

**UNITED STATES DEPARTMENT OF COMMERCE National Oceanic and Atmospheric Administration** NATIONAL MARINE FISHERIES SERVICE West Coast Region 501 West Ocean Boulevard, Suite 4200 LONG BEACH, CA 90802

**Refer to NMFS No: WCRO-2023-02924**

October 7, 2024

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Mr. Matthew Davis Forest Supervisor Payette National Forest 500 North Mission Street, Building 2 McCall, Idaho 83638

Lt. Col. Kathryn Werback U.S. Army Corps of Engineers Walla Walla District 201 N. 3rd Avenue Walla Walla, Washington 99362-1876

Re: Endangered Species Act Section 7(a)(2) Biological Opinion and Magnuson–Stevens Fishery Conservation and Management Act Essential Fish Habitat Response for the Stibnite Gold Project, South Fork Salmon River HUC 17060208, Valley County, Idaho; Lemhi River HUC 17060204, Lemhi County, Idaho.

Dear Mr. Davis and Lt. Col. Werback:

Thank you for your letter of March 26, 2024, requesting initiation of consultation with NOAA's National Marine Fisheries Service (NMFS) pursuant to section 7 of the Endangered Species Act of 1973 (ESA) (16 U.S.C. 1531 et seq.) for the Stibnite Gold Project. Thank you, also, for your request for consultation pursuant to the essential fish habitat (EFH) provisions in Section 305(b) of the Magnuson–Stevens Fishery Conservation and Management Act (MSA) [16 U.S.C. 1855(b)] for this action. Upon review, we determined your submittal was sufficient and initiated consultation on March 26, 2024. On July 25, 2024, the U.S. Forest Service (USFS) submitted additional information, modifying the proposed action.

In this biological opinion (opinion), NMFS concludes that the action, as proposed, is not likely to jeopardize the continued existence of Snake River (SR) spring/summer Chinook salmon (*Oncorhynchus tshawytscha*) and SR Basin steelhead (*O. mykiss*). NMFS also determined the action will not destroy or adversely modify designated critical habitat for these species. Rationale for our conclusions is provided in the attached opinion.

The USFS and the U.S. Army Corps of Engineers (USACE) also determined that the proposed action may affect, but is not likely to adversely affect Southern Resident killer whale (*Orcinus orca*) and their designated critical habitat. NMFS concurs with this determination and provides the rationale for this determination in this opinion.



As required by section 7 of the ESA, NMFS provides an incidental take statement (ITS) with the opinion. The ITS describes reasonable and prudent measures (RPM) NMFS considers necessary or appropriate to minimize the impact of incidental take associated with this action. The take statement sets forth terms and conditions, including reporting requirements, that the USFS, USACE, Perpetua Resources, Inc., and any permittee who performs any portion of the action, must comply with in order to be exempt from the ESA take prohibition.

This document also includes the results of our analysis of the action's effects on EFH pursuant to section 305(b) of the MSA, and includes thirteen Conservation Recommendations to avoid, minimize, or otherwise offset potential adverse effects on EFH. These Conservation Recommendations are similar, but not identical to the ESA terms and conditions. Section 305(b)(4)(B) of the MSA requires federal agencies to provide a detailed written response to NMFS within 30 days after receiving these recommendations. If the response is inconsistent with the EFH Conservation Recommendations, the USFS and USACE must explain why the recommendations will not be followed, including the justification for any disagreements over the effects of the action and the recommendations. In response to increased oversight of overall EFH program effectiveness by the Office of Management and Budget, NMFS established a quarterly reporting requirement to determine how many Conservation Recommendations are provided as part of each EFH consultation and how many are adopted by the action agency. Therefore, in your statutory reply to the EFH portion of this consultation, NMFS asks that you clearly identify the number of Conservation Recommendations accepted.

Please contact Bill Lind, Southern Snake Branch Office, at (208) 391-1282, [Bill.lind@noaa.gov](mailto:Bill.lind@noaa.gov) or Johnna Sandow, Northern Snake Branch Office, at (208) 378-5737, [Johnna.sandow@noaa.gov](mailto:Johnna.sandow@noaa.gov) if you have any questions concerning this consultation, or if you require additional information.

Sincerely.

Jennifer Quan Regional Administrator West Coast Region

Enclosure

cc: K. Knesek – PNF R. Rymerson - PNF C. Nalder – PNF K. Urbanek - USACE B. Wilson - USACE K. Hendricks - USFWS C. Wise - USFWS M. Lopez - NPT C. Colter – SBT B. Gibson - SPT

#### **Endangered Species Act (ESA) Section 7(a)(2) Biological Opinion and Magnuson–Stevens Fishery Conservation and Management Act Essential Fish Habitat Response**

Stibnite Gold Project

NMFS Consultation Number: WCRO-2023-02924



Affected Species and NMFS' Determinations:





**Consultation Conducted By:** National Marine Fisheries Service, West Coast Region

**Issued By:**  $\leftarrow$   $\mu$ Jennifer Quan

Regional Administrator West Coast Region National Marine Fisheries Service

**Date:** *October 7, 2024*

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### **1. INTRODUCTION**

<span id="page-20-0"></span>This Introduction section provides information relevant to the other sections of this document and is incorporated by reference into Sections 2 and 3, below.

# <span id="page-20-1"></span>**1.1. Background**

National Marine Fisheries Service (NMFS) prepared the biological opinion (opinion) and incidental take statement (ITS) portions of this document in accordance with section 7(b) of the Endangered Species Act (ESA) of 1973 (16 U.S.C. 1531 et seq.), as amended, and implementing regulations at 50 CFR 402.

We also completed an essential fish habitat (EFH) consultation on the proposed action, in accordance with section 305(b)(2) of the Magnuson–Stevens Fishery Conservation and Management Act (MSA) (16 U.S.C. 1801 et seq.) and implementing regulations at 50 CFR 600.

We completed pre-dissemination review of this document using standards for utility, integrity, and objectivity in compliance with applicable guidelines issued under the Data Quality Act (DQA) (section 515 of the Treasury and General Government Appropriations Act for Fiscal Year 2001, Public Law 106-554). The document will be available within 2 weeks at the NOAA Library Institutional Repository [https://repository.library.noaa.gov/welcome]. A complete record of this consultation is on file at the Snake Basin Office in Boise, Idaho.

# <span id="page-20-2"></span>**1.2. Consultation History**

Perpetua Resources Idaho, Inc. (Perpetua) proposed the Stibnite Gold Project (SGP or Project) in central Idaho. The mine at Stibnite occurs on private, state, and public lands administered by the Boise National Forest (BNF), Payette National Forest (PNF), and the Bureau of Reclamation (BOR) in Valley County, Idaho (Figure 1), while the Lemhi River restoration portion of the project occurs in Lemhi County, Idaho (Figure 2). The SGP is located in the Stibnite-Yellow Pine Mining District (District) in central Idaho, near the Frank Church River of No Return Wilderness (FC-RNRW) and along the Lemhi River downstream from Leadore, Idaho. A summary of the consultation history follows. For a more detailed description of early coordination efforts and the consultation history, please see Appendix A of the final biological assessment (BA) (Stantec 2024).

NMFS and the U.S. Fish and Wildlife Service (FWS) have been meeting with the action agencies monthly since 2018 in early coordination efforts. Regularly scheduled (monthly) consultation meetings were initiated on June 21, 2018, with participation of the U.S. Forest Service (USFS), AECOM (USFS contractor), Perpetua, FWS, NMFS, U.S. Army Corps of Engineers (USACE), and the U.S. Environmental Protection Agency (EPA). Meetings continued through March 2020, and covered a wide variety of topics related to analysis methodology, data, BA preparation, etc.

The PNF and USACE submitted their first draft BA on October 26, 2023. Upon review of the draft BA, NMFS issued an insufficiency letter on November 22, 2023, outlining information needed in order to provide a complete initiation package. The action agencies and their

contractor Stantec later met with NMFS and the FWS via conference call to discuss their approach to BA revisions on January 22 and February 7, 2024. A revised BA was submitted to NMFS on March 26, 2024, and consultation was initiated at this time.

The BNF initiated informal consultation for the permitting of Perpetua's Burntlog Route Geophysical Investigation (BRGI) Project on February 24, 2022. NMFS issued a concurrence letter for the project on March 14, 2022 (NMFS No: WCRO-2022-00428). On July 11, 2024, the BNF withdrew approval for the BRGI. Approval for the BRGI will now be considered in the NEPA decision for the larger SGP. On July 25, 2024, NMFS received a letter from the PNF transmitting a revised BA, fully incorporating the BRGI project into the SGP proposed action.

The USFS, as the lead Federal agency, evaluated the mining project proposal under regulations found at 36 Code of Federal Regulations (CFR) 228 subpart A, submitting the BA in accordance with Section 7 of the ESA (16 United States Code [USC] 1536(c)) and Section 305(b) of the MSA. In addition, the USACE evaluated the SGP per its requirements under the Clean Water Act (CWA). The USFS' approval of the Plan of Restoration and Operations, and the USACE's issuance of the CWA 404 permit, provide the federal nexus to ESA consultation. The PNF designated Perpetua as the non-federal representative for the SGP on March 14, 2018. The PNF also granted Perpetua applicant status for the SGP on August 30, 2017.

In the BA, the action agencies determined that the proposed action "may affect," and is "likely to adversely affect Snake River (SR) spring/summer Chinook salmon (*Oncorhynchus tshawytscha*), SR Basin steelhead (*O. mykiss*), and their designated critical habitats (DCH). The USFS also determined that the proposed action "may affect," but is "not likely to adversely affect" Southern Resident Killer Whale (SRKW) (*Orcinus orca*) and their DCH. Our concurrence is documented in the NLAA Determinations section (Section 2.21).

Updates to the regulations governing interagency consultation (50 CFR part 402) were effective on May 6, 2024 (89 Fed. Reg. 24268). We are applying the updated regulations to this consultation. The 2024 regulatory changes, like those from 2019, were intended to improve and clarify the consultation process, and, with one exception from 2024 (offsetting reasonable and prudent measures [RPMs]), were not intended to result in changes to the Services' existing practice in implementing section 7(a)(2) of the Act. 89 Fed. Reg. at 24268; 84 Fed. Reg. at 45015. We have considered the prior rules and affirm that the substantive analysis and conclusions articulated in this opinion and ITS would not have been any different under the 2019 regulations or pre-2019 regulations.

On August 8, 2024, NMFS provided a copy of the draft proposed action to the PNF. NMFS received an August 19, 2024 email from the PNF explaining they completed their review and the did not have any substantive comments or recommended changes for the propose action. We also shared a copy of the draft terms and conditions with the PNF, USACE, Nez Perce Tribe, Shoshone-Bannock Tribes, and Shoshone-Paiute Tribes on September 12, 2024.

In preparing this opinion, NMFS relied upon information from the BA (Stantec 2024) and its supporting documentation, published scientific literature, and various government documents (e.g., recovery plans, 5-year reviews, listing decisions, etc.). This information provided the basis for our determinations as to whether the PNF and USACE can ensure that their proposed action is not likely to jeopardize the continued existence of ESA-listed species, and is not likely to result in the destruction or adverse modification of DCH.



<span id="page-23-0"></span>**Figure 1. Stibnite Gold Project Location, Stibnite, Idaho.**



<span id="page-24-0"></span>**Figure 2. Lemhi River Restoration Project, Lemhi County, Idaho.**

## <span id="page-25-0"></span>**1.3. Proposed Federal Action**

Under the ESA, "action" means all activities or programs of any kind authorized, funded, or carried out, in whole or in part, by Federal agencies (see 50 CFR 402.02). Under the MSA, "Federal action" means any action authorized, funded, or undertaken, or proposed to be authorized, funded, or undertaken by a Federal agency (see 50 CFR 600.910).

This section of the opinion includes relevant excerpts of the proposed action as described in the BA (Stantec 2024), supplemented by additional information provided in the following support documents: Compensatory Stream and Wetland Mitigation Plan (CMP) (Tetra Tech 2023); Fisheries and Aquatic Resources Mitigation Plan (FMP) (Brown and Caldwell, Rio ASE, and BioAnalysts 2021); Fishway Operations and Management Plan (FOMP) (Brown and Caldwell, McMillen Jacobs, and BioAnalysts 2021); and, the Stream Design Report (Rio ASE 2021).

#### <span id="page-25-1"></span>**1.3.1. Overview**

The Proposed Action is based upon Perpetua's Modified Plan of Operations submitted in October 2021 for the SGP (Perpetua 2021a), referred to as the 2021 Modified Mine Plan (2021 MMP). Mine operations will occur on patented mining claims owned or controlled by Perpetua and on unpatented mining claims and other areas of federal public lands comprised of national forest system (NFS) lands that are administered by the PNF. Supporting infrastructure corridors (access and transmission line) are located on the BNF, Idaho Department of Lands (IDL), BOR, and non-federal lands. This document includes only sections of the proposed action relevant to effects on ESA-listed anadromous fish or their DCH. For a complete description of the proposed action, please see the BA (Stantec 2024).

Perpetua proposes to develop a mine operation that produces gold and silver doré, and antimony concentrates from ore deposits associated with their mining claims in the project area. The estimated recoverable mineral resource consists of: 4.2 million ounces of gold, 1.7 million ounces of silver, and 115 million pounds of antimony.

Development of the mineral resource will include construction of access and haul roads; construction of supporting infrastructure; open pit mining; ore processing; placement of tailings in a tailings storage facility (TSF); and placement of development rock. New access to the SGP will be provided by the proposed Burntlog Route, which will be a combination of widening the existing Burnt Log Road (Forest Road [FR] 447), Thunder Mountain Road (FR 50375), Meadow Creek Lookout Road (FR 51290), and constructing new connecting road segments of approximately 15 miles (Figure 3). Development of the Burntlog Route will entail 340.9 acres of new cut and fill activity (including borrow sources) along existing and newly constructed roadways.

To provide electric power for the SGP, an existing powerline will be upgraded and a new transmission line from a new Johnson Creek substation to the mine will be constructed. Additional off-site support facilities to be constructed along access corridors include the Stibnite Gold Logistics Facility (SGLF) and the Burntlog Access Route Maintenance Facility. The SGLF will house administrative offices, the assay laboratory, and a warehouse while the maintenance facility will be the headquarters for road maintenance and snow removal. The proposed facilities and access roads are shown on Figure 3 and Figure 4. The Operations Area Boundary shown on Figures 3 and 4 is the boundary within which Perpetua will control public access.

The proposed action includes activities that will result in permanent impacts to waters of the United States (WOTUS) including wetlands. Therefore, Perpetua has proposed a CMP as part of their CWA 404 permit. The CMP addresses compensatory mitigation for permanent impacts that will be accomplished through a combination of mitigation bank credits and the creation of new wetlands, streams, and enhancing and reclaiming existing wetlands and streams in the general vicinity of the impact areas. The CMP also addresses compensatory mitigation to reduce the temporal loss of aquatic functions and potential risks associated with actions described in the CMP.

As part of the CMP, Perpetua has proposed the Lemhi Regional Land Trust's (LRLT) Little Springs Conservation Easement and Restoration Project in an effort to use offsite mitigation to offset temporal losses to fish habitat in the SFSR drainage. The LRLT is negotiating a perpetual conservation easement for the property to ensure the restoration benefits and associated mitigation credits persist for at least the length of the predicted temporal loss for the SGP. This project reach is located on private ranch lands on the upper Lemhi River, approximately 12 miles northwest of Leadore, Idaho.

The components of the proposed action are described in the following sections in terms of overall land management and affected areas, and project phases: construction; operations; exploration; and closure and reclamation, including post-closure monitoring.



<span id="page-27-0"></span>**Figure 3. Project Overview and Location for the Stibnite Gold Project.**



<span id="page-28-0"></span>**Figure 4. Operations Area Boundary and Mine Site Layout, Stibnite, Idaho.**

#### <span id="page-29-0"></span>**1.4. Land Management and Affected Areas**

Table 1 provides a summary of land management or ownership by estimated SGP component for the maximum affected area proposed and also includes acreages of new disturbance and redisturbance by SGP component and ownership.

Component		<b>Perpetua</b> <b>Private</b>	. <b>Other</b> <b>Private</b>	<b>PNF</b>	<b>BNF</b>	SCNF <sup>4</sup>	<b>BOR</b>	<b>IDL</b>	<b>Totals</b>
Mine Site	New Disturbance	48.2	$\theta$	$767.9 + 65^2$	$\Omega$	$\Omega$	0	$\Omega$	881.1
	Re-disturbance	456.7	$\theta$	402.3	$\Omega$	$\Omega$	0	$\Omega$	859.0
Off-site Facilities	New Disturbance	24.3	$\Omega$	$\Omega$	4.5	$\Omega$	$\theta$	$\Omega$	28.8
	Re-disturbance	$\Omega$	$\theta$	$\theta$	$\Omega$	$\Omega$	0	$\Omega$	$\Omega$
Access Roads	New Disturbance	$\Omega$	$\Omega$	81.6	253.8	5.5	0	$\Omega$	340.9
	Re-disturbance	1.9	4.5	26.9	102.5	8.7	$\Omega$	$\Omega$	144.5
Utilities $1$	New Disturbance	2.9	105.9	61.4	221.8	$\Omega$	3.5	26.0	421.5
	Re-disturbance		174	19.4	350.6	$\theta$	9	36.1	590.1
Disturbance Totals	<b>Total New Disturbance</b>	75.4	105.9	$910.9 + 65^2$	480.1	5.5	3.5	26.0	1,672.3
	<b>Total Re-disturbance</b>	459.6	178.5	448.6	453.1	8.7	9	36.1	1,593.6
<b>Total New and Re-Disturbance</b>		535.0	284.4	1,424.5	933.2	14.2	12.5	62.1	$3,265.9^3$

<span id="page-29-2"></span>**Table 1. Land Management and Acreage by Component for the Proposed Action.**

**Key:** PNF – Payette National Forest; BNF – Boise National Forest; Salmon-Challis National Forest; BOR – Bureau of Reclamation; IDL – Idaho Department of Lands.

**Notes:** <sup>1</sup> Utilities affected areas include both existing utility corridors and access routes, and new utility corridors and access routes. Some existing utility access routes will be upgraded.

 $2$  Approximately 65 affected acres associated with temporary surface exploration pads and roads (SGP component) have an unknown land ownership because the exact locations of these exploration areas are not yet known; however, these are included in the PNF SGP subtotal.

 $3$  Items, subtotals, and totals may not add up to the grand total due to rounding.

<sup>4</sup> Approximately 14 acres of land is administered by the PNF but is within the boundary of the SCNF.

#### <span id="page-29-1"></span>**1.5. Phases and Timeline**

The proposed action will take place over a period of approximately 20 to 25 years, not including the long-term, post-closure environmental monitoring or potential long-term water treatment. The phases of the SGP are described in subsequent sections and include: (1) construction (approximately 3 years; Mine Years (MY) -3 through -1); (2) mining and ore processing operations (approx. 15 years; MYs 1 through 15); (3) surface and underground exploration (approx. 17 years, beginning during construction and continuing concurrent with operations; MYs -2 through 15); and (4) closure and reclamation (MY 16+). Most activities in the Closure and Reclamation period will be completed within five years. However, closure water management and water treatment are expected to continue for as long as 25 years based on predictions for draindown and consolidation of the TSF (MYs 16 through 40) (Perpetua 2021a). The environmental monitoring phase will continue for as long as needed to demonstrate that the site has been fully reclaimed. Figure 5 provides an illustration of the timing of construction and operations activities and the initiation of the closure phase.



<span id="page-30-2"></span>**Figure 5. Stibnite Gold Project Phases and Timeline.**

## <span id="page-30-0"></span>**1.6. Site Preparation, Access, and Infrastructure**

## <span id="page-30-1"></span>**1.6.1. Overview**

Preparing the proposed action will require construction of surface facilities, haul roads, and water management features. Environmental design features (EDF) for facilities associated with regulatory requirements are summarized in Table 2. Supporting infrastructure will include transmission lines, substations, communication sites, access roads, and a fish tunnel.

Additionally, removal of some features from past mining activities (legacy mining features) will be initiated during the construction phase. Perpetua will install 15 to 20 temporary trailers on private lands adjacent to the existing exploration camp (located in the proposed ore processing area) to accommodate construction crews; these temporary trailers will be used during site preparation and early construction until the worker housing facility is constructed.



<span id="page-31-0"></span>









Prior to site preparation and construction of surface facilities, vegetation will be removed from operating areas. Trees, deadwood, shrubs, and slash not needed to construct windrows at the edge of Burntlog Route disturbance (to function as sediment barriers), will be chipped, and suitable soil will be separately salvaged and stockpiled (except for a small portion that will be 'live handled') for use as part of site reclamation and restoration. Portions of the salvaged soil will be blended with the chipped wood to create growth media (GM).

The existing potable water supply system at the exploration camp will be used and expanded for the construction camp. The existing system will be supplemented with deliveries of potable water, if needed. Supplemental water sources (i.e., water deliveries) will be used by personnel in remote construction areas. Sanitation during construction will be provided through the existing sewage treatment system adjacent to the exploration camp. In addition, portable sanitary facilities will be located throughout the SGP and at remote construction areas.
Construction of the Burntlog Route will occur from both ends of the route at the same time on a seasonal basis (May to November), but construction could occur outside of this time period if conditions allow (i.e., snow free). The southern portion workforce will be housed in three temporary trailer camps located within construction borrow sources or staging areas. The northern portion workforce will be housed at the temporary trailer construction camp at the SGP. Some construction workers could be housed in Cascade, Idaho.

### **1.6.2. Growth Media Stockpiles**

Suitable GM within the area proposed for operations will be salvaged following vegetation clearing and moved to growth media stockpiles (GMS) either within the Fiddle Valley or at the Worker Housing Facility. Another short-term GMS will be located within the footprint of the TSF. Growth media from the new construction of the Burntlog Route will be stockpiled in the borrow source areas used for construction and widening of the route and in windrows along the edges of fill slopes. GMS will be stabilized, seeded, and mulched to protect the stockpile from wind and water erosion. A total of approximately 1,657,246 bank cubic yards (BCY) of suitable soils (GM and seed bank material [SBM]) will need to be salvaged from the SGP for reclamation. A total of approximately 860,373 BCY of GM, chipped wood blend, and SBM are available for salvage at the SGP.

To achieve the reclamation success criteria and offset the GM deficits, 1.5 million BCY of unconsolidated overburden (chiefly alluvial and glacial materials from the Yellow Pine Pit [YPP]) will be stored in the Fiddle GMS to allow use as cover material for reclamation of the TSF, TSF Buttress, and Hangar Flats Pit (HFP) backfill.

#### **1.6.3. Access Roads**

During the construction phase, site access to the SGP will cross 43 streams along existing roads, and crossing 28 streams along the Burntlog Route, including the existing Burntlog Route (Figure 6) (USFS 2023a). In addition to these stream crossings, approximately 6.5 miles (18 percent of its 36-mile length) of the Johnson Creek Route is located in close proximity to streams (i.e., within 100 feet). A total of 65 vehicle trips per day will occur during the construction phase (USFS 2023b). These trips will consist of 20 light vehicles and 45 heavy vehicles. The 65 trips will be along the Johnson Creek route.

During the mining and ore processing operations phase (approximately 15 years), a total of 50 vehicle trips per day are anticipated on average (year-round) utilizing the Burntlog Route. The 50 trips will consist of 17 light vehicles and 33 heavy vehicles. During the closure and reclamation phase, traffic along the Burntlog Route will be reduced to a total of 27 vehicle trips per day (year-round).

#### **Warm Lake Road**

Warm Lake Road (County Road [CR] 10-579) is a two-lane (one lane each direction), asphaltpaved roadway with lane markings and is open year-round to all vehicles from Idaho State Highway (SH) 55 to Warm Lake. Warm Lake Road starts in Cascade at an intersection with SH 55 and continues eastward for approximately 35 miles, ending at Johnson Creek Road (CR 10-

413) at Landmark. Warm Lake Road is under the jurisdiction of Valley County, but Valley County currently does not maintain Warm Lake Road in winter beyond Warm Lake Lodge. During years with adequate snowpack, an 8-mile segment of the Warm Lake Road route east of Warm Lake Lodge is used as an over snow vehicle (OSV) route, allowing access into Landmark and other areas.

The SGP will require year-round passenger and delivery truck access from the onset of construction through the life of the mine. Perpetua will conduct wintertime maintenance east of Warm Lake Lodge to ensure safe, year-round access to the sole route of ingress/egress to the SGP for all mine support traffic. This will include snow removal and road sanding, as appropriate, to maintain a safe driving surface. Commitments for wintertime maintenance of Warm Lake Road will be documented in a Road Maintenance Agreement with Valley County.



**Figure 6. Burntlog Route, Stibnite Gold Project.**

Perpetua wintertime maintenance and use of Warm Lake Road will result in two changes to current traffic conditions:

- Warm Lake Road east of Warm Lake Lodge will not be available as a recreational OSV route from the start of construction through reclamation of the SGP. To replace this recreational use, a dedicated alternative OSV route will be established from the Warm Lake area to Landmark via the Cabin Creek/Trout Creek drainages and adjacent to the Johnson Creek Road. Establishing this replacement OSV route will minimize the interactions between SGP traffic and recreational traffic in the winter. The proposed OSV route is illustrated in Figure 7.
- Expanded wintertime public vehicle access on Warm Lake Road east of Warm Lake Lodge will commingle SGP and public travel.

Warm Lake Road and its supporting infrastructure (i.e., for stormwater management) are not being expanded or modified.

The USFS is not a party to Perpetua's Road Maintenance Agreement with Valley County, the owner of the Warm Lake Road, Johnson Creek Road, and Stibnite Road. Therefore, the USFS will not be involved in the review, implementation, or enforcement of the agreement from a road maintenance perspective. However, as part of Project approval, the USFS will require the Project implementation of environmental requirements pertaining to road use and maintenance indicated in this document (Appendix A, Table A-5).

In the event that road maintenance requires more substantial efforts than typical maintenance (e.g., landslide or avalanche recovery), the USFS will engage with Valley County and Perpetua on efforts that will affect USFS lands outside the current road footprint and roadside support structures (e.g., ditches, culverts). Maintenance activities within the existing road footprint and support structures will not require additional USFS engagement. Activities involving USFS land not currently utilized by the road and support structures will require additional USFS engagement and potential permitting.



**Figure 7. OSV Routes – Stibnite Gold Project.**

#### 1.6.3.2. **Johnson Creek Route**

During the initial construction period of the Burntlog Route (approximately 2 to 3 years), minerelated traffic will access the SGP from SH 55, north of the city of Cascade, via Warm Lake Road for approximately 34 miles. Then, north on Johnson Creek Road (CR 10-413) for approximately 25 miles to the village of Yellow Pine, and from Yellow Pine east approximately 14 miles to the SGP via the Stibnite Road (CR 50-412). The portion of the route that includes Johnson Creek Road and Stibnite Road is known as the Johnson Creek Route. This route is primarily situated topographically adjacent to the valley bottom, paralleling Johnson Creek and the East Fork South Fork Salmon River (EFSFSR).

Johnson Creek Road is a county-maintained, native-surface road that is open to vehicles with seasonal restrictions due to snow. During the winter, Valley County plows approximately 10 miles of Johnson Creek Road from Yellow Pine south to Wapiti Meadow Ranch and grooms the remaining 17 miles of Johnson Creek Road from Wapiti Meadow Ranch to Warm Lake Road at Landmark for OSV use.

The Stibnite Road portion of the route is also a county-maintained native surface road, open to all vehicles with seasonal restrictions due to snow. This road is plowed in the winter by Perpetua through an agreement with Valley County to allow site access for exploration activities. Seasonal restrictions and measures implemented under the Golden Meadows Exploration Project to restrict access will remain in place during the three-year construction period. Upon construction of the Burntlog Route, winter plowing of the Stibnite Road for SGP access will be discontinued. Stibnite Road connects to Thunder Mountain Road on the southeastern portion of the Stibnite site and currently provides seasonal (non-winter) public access through the site.

Minor surface improvements (such as ditch and culvert repair, winter snow removal, resurfacing [i.e., gravel addition] if required, and summer dust suppression) will occur on the Johnson Creek Route under the proposed action to reduce sediment runoff and dust generation. However, there will be no road alignment modification or widening of the road prism of these existing roads along the Johnson Creek Route, as the current road is able to accommodate the equipment and materials needed to be transported during the construction period. In practice, resurfacing, dust suppression, and repairs will be conducted on an as needed basis. Resurfacing and repair work will generally occur annually following the winter season. Dust suppression using water application will be frequent (i.e., every few weeks) during the summer season.

Use of chemical dust suppressants such as magnesium chloride  $(MgCl<sub>2</sub>)$  will occur near the start of the summer season. Water application will utilize over-the-road water trucks (e.g., 2,000 gallon) that will fill from diversion points authorized by Idaho Department of Water Resources (IDWR) water rights approval (e.g., from the groundwater well at the Landmark Maintenance Facility).

Prior to construction of the Burntlog Route, the Johnson Creek Route will be used for fuel transport with precautionary measures being: (1) staged spill response kits; (2) pilot cars equipped with spill response kits; (3) radio contact with hauling trucks; (4) only day-time fuel deliveries; and (5) driver training on route. Once the Burntlog Route is completed, fuel transport using the Johnson Creek Route will be discontinued.

Portions of Johnson Creek Road (i.e., Landmark to Wapiti Meadows) are currently used as a groomed OSV trail during winter and use of the Johnson Creek Route by mine-related construction traffic will conflict with this existing groomed OSV trail. Thus, while the Burntlog Route (described below) is under construction, a temporary 16-foot-wide groomed OSV trail adjacent to Johnson Creek Road between the proposed Cabin Creek Groomed OSV Route and Landmark will be constructed (Figure 7). However, the OSV trail from Trout Creek Campground to Wapiti Meadows will be closed until construction of the Burntlog Route is complete; once mine traffic moves to that route, then the OSV route will return to Johnson Creek Road and will reconnect Landmark with Wapiti Meadows.

Perpetua has an existing agreement with Valley County for maintenance of Johnson Creek and Stibnite Roads, including performing maintenance measures to repair segments that have deteriorated. Appropriate revisions to the road maintenance agreement will be established for use of the Johnson Creek Route as a construction route and to ensure year‐round access in accordance with Valley County's public road easement stipulations. Once construction of the Burntlog Route has been completed (2-3 years), the Johnson Creek Route will no longer be used by mine-related traffic.

# 1.6.3.3. **Burntlog Route**

The Burntlog Route will connect the eastern end of Warm Lake Road (at Landmark) to the SGP (to the northeast) by widening and improving approximately 23 miles of existing roads, including the full length of the existing Burntlog Road (FR 447) and segments of Meadow Creek Lookout Road (FR 51290) and Thunder Mountain Road (FR 50375). The three road segments will be connected with two new road segments totaling approximately 15 miles. Burntlog Road is currently a native surface road that is open year-round to all vehicles with seasonal restrictions due to snow. The last 0.25 to 0.5 mile of the existing road is closed and motorized traffic prohibited. Meadow Creek Lookout Road is a native surface road, open year-round to all vehicles. The Burntlog Route is primarily situated topographically on mid-slopes and ridgeline.

Improvements on the existing roads that comprise the Burntlog Route will include:

- Straightening tight corners to allow for improved safety and traffic visibility;
- Maintaining grades of less than 10 percent in all practicable locations;
- Placing sub-base material and surfacing with gravel;
- Application of a road binding agent in localized segments to suppress dust, increase stability, and reduce sediment runoff;
- Widening the existing road surface (currently approximately12 ft. wide) to a 21 ft. wide travel way (approximately 26 ft. including shoulders); and,
- Installing side-ditching, culverts, guardrails, and bridges, where necessary, with EDF to provide fish passage and limit potential sediment delivery to streams.

Figure 6 shows the proposed Burntlog Route, which includes the proposed new road construction. A segment of new road construction for the Burntlog Route will be located on the south side of the Riordan Creek drainage and cross Riordan Creek north of Black Lake. The approximately 5.3-mile road segment will have 12 stream crossings, three of which cross perennial streams. Along the Burntlog Route, culverts designed and installed to allow fish

passage will be used for fish-bearing stream segments crossed by the route. Upon construction of the Burntlog Route, the route will be used for fuel transport with precautionary measures being: (1) staged spill response kits; (2) pilot cars equipped with spill response kits; (3) radio contact with hauling trucks; (4) only day-time fuel deliveries; and (5) driver training on route. After construction is complete, public use will be allowed on Burntlog Route when other public access roads are blocked by mine operations.

The connection segment between the end of Burntlog Road and Meadow Creek Lookout Road is approximately 11 miles and will cross Trapper Creek. The second connector between the Meadow Creek Lookout Road and Thunder Mountain Road will be approximately 4 miles and links up with Thunder Mountain Road approximately 2 miles south of the SGP. Minor surface improvements (e.g., blading) will occur on the portions of the existing Thunder Mountain Road and Meadow Creek Lookout Road that will not become part of the Burntlog Route to provide a safe road surface for transportation of construction equipment required to build the Burntlog Route. There will be no road alignment modification or widening of the portions of the existing roads that are not part of the Burntlog Route.

Construction of new segments of the Burntlog Route will utilize cut and fill techniques to create a level surface for installation of the roadway. Most cut and fill will be conducted by mechanical construction equipment (i.e., dozers, rollers, graders) to relocate and compact unconsolidated materials then place a gravel road surface. In instances where consolidated bedrock material is encountered in cut areas, blasting may be used to break up the bedrock to complete the cut. Areas requiring blasting will typically occur on steeper side slopes in upland areas where unconsolidated soil and cover materials overlying bedrock may have limited thickness.

Primary SGP access will shift from the Johnson Creek Route to the Burntlog Route near the end of the construction phase. The Burntlog Route will avoid environmental and human health and safety risks associated with the Johnson Creek Route which passes through identified areas for avalanches, landslides, and floods. This route will decrease the potential for spill risk adjacent to fish-bearing streams.

The Burntlog Route road design prioritizes safe transportation conditions (e.g., appropriate road grade, width, turning radii, etc.) that minimize the risk associated with traffic incidents as well as minimize impacts to environmental resources. The EDF, reclamation, and mitigation measures to minimize or offset impacts are described in the BA and compiled in Appendix A, Tables A-1 to A-9.

#### **Burntlog Route Borrow Sources, Staging Areas, and Construction Camps**

Up to eight borrow sites will be established along the Burntlog Route (Figure 6) to meet construction and ongoing maintenance throughout the life of the mine and to support decommissioning following mine closure while avoiding Riparian Conservation Areas (RCA). Additionally, those same eight borrow areas will be utilized for staging of equipment and supplies. Three construction camps will be located within the disturbance created by borrow sources or staging areas. The construction camps will be for trailer parking. Each trailer will be equipped with fresh water and sanitary waste storage.

#### **Culverts and Bridges**

Construction of the Burntlog Route (i.e., improvement of the existing FS 447 plus new road development) will require installation of bridges and culverts to cross stream segments and to manage stormwater diverted from the roadway. Design documents refer to culverts that crossstream segments containing perennial flows as "Stream Crossing Culverts" while culverts that do not cross flowing streams but instead manage stormwater are referred to as "Relief Culverts".

EDF for bridges and culverts consider the criteria described in the Forest Service Structures Handbook (USFS 2014), the Valley County Roadway Design Guide (Valley County 2008), and NMFS guidelines (NMFS 2022a). These criteria are summarized in Table 3.

The number of bridges, stream crossing culverts, and relief culverts by drainage are summarized in Table 4. Bridges and culverts which have been installed within the last 20 years that are in good condition will be retained if possible.



#### **Table 3. Bridge and Culvert Design Criteria.**

#### **Table 4. Burntlog Route Bridges and Culverts.**



\*Stream culverts sized based on 100-year flows (see below for fish passage information).

\*\*Relief culverts sized for 25-year events based on drainage area runoff analyses.

All of the bridges cross stream segments with fish passage while nine of the 25 stream crossing culverts cross segments with fish passage based on drain area analyses and environmental DNA  $(eDNA)^1$  data. In addition to the Burntlog Route access road crossings, there will be one haul road crossing of the EFSFSR in the mining area along with an existing box culvert. Design information for the crossings with fish passage are summarized in Table 5. Final designs for bridges, culverts, and plate arches are pending geotechnical assessment of the crossing locations. Conceptual design drawings are shown in Figures 3.4-3 through 3.4-10 of the BA (Stantec 2024).

Project incorporation of site preparation, staging, and sequencing stream crossing work has been developed for instream work such as the Burntlog Route stream crossings as described in the FMP, Section 5.4.7 (Brown and Caldwell, Rio ASE, and BioAnalysts 2021). A planning team with representation from project management, engineering, and fish biology will be assembled to coordinate with construction personnel and equipment operators to plan the staging and sequence for work area isolation, fish capture and removal, and dewatering, i.e.:

- Scheduling within an appropriate instream work window (see Section 1.11.2.2);
- Establishing the length of channel to be isolated for each crossing;
- Conducting work area isolation and fish salvage in consideration of habitat requirements, flow and temperature conditions, and exposure to turbidity or other unfavorable conditions; and,
- Dewatering via a bypass flume or culvert with diversion by sandbags, sheet piling, or cofferdam.

When stream segments require dewatering for bridge or culvert installation, they will be isolated using a method appropriate for the location, including block nets, sandbags, diversion, pumps, sheetpiling, flashboards, cofferdams, and other structures. The specific method will depend on the stream segment location, diversion sequencing, operational requirements, segment length, segment slope, flow conditions, depth, and fish salvage (see below). All isolation barriers will be monitored during installation and operation. Partial dewatering will generally be conducted during low-flow periods to facilitate stream segment isolation and fish salvage. Whenever possible, dewatering will not begin until fish have been captured and removed for relocation. However, depending on the location and water depth, it may be necessary to partially draw down the water first to perform fish removal. Partial dewatering before fish salvage operations begin may also improve fish capture efficiency by reducing the total volume of stream habitat that needs to be salvaged. In those cases, dewatering pumps will be screened to meet NOAA Fisheries and IDFG standards to avoid entrainment of juvenile fish. Fish capture from work area isolation will consist of:

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<sup>&</sup>lt;sup>1</sup> Environmental DNA sampling is a technique that identifies DNA found in the environment (e.g., water, soil, air) from cellular material shed by organisms that has accumulated in the surrounding water, soil, air, etc. It allows for testing without having to sample the organism itself; and is a tool that allows for monitoring and detection of species which may be present in low numbers and therefore difficult to find.

- Slowly reducing flow in the work area to allow some fish to leave volitionally;
- Installation of block nets upstream and downstream of the isolation area with the nets secured to stream channel bed and banks until fish capture is complete and exclusion of fish from the work area is necessary;
- Hourly monitoring of block nets during instream disturbance in the work area;
- If block nets are in place for more than one day, they will be monitored daily to ensure they are secured to banks and are free of organic accumulation plus monitored every four hours for fish impingement if located in bull trout (*Salvelinus confluentus*) spawning and rearing habitat (unless a variance is granted by the USFS);
- Seining the isolated area to capture and relocate fish;
- If areas are isolated overnight, minnow traps will be placed overnight in conjunction with seining;
- Collecting any remaining fish by hand or dip nets as dewatering continues; and,
- If all other techniques have been exhausted, electrofishing may be used to capture remaining fish under electrofishing conservation measures.

Captured fish will be relocated as quickly as possible to pre-planned release areas using aerated and shaded transport buckets holding limited numbers of fish of comparable size to minimize predation. Dead fish will not be stored in transport buckets but will be left on the streambank to avoid mortality counting errors.

Upon completion of the instream work, flow diversions will be removed slowly to allow gradual rewatering of the isolated stream segment to minimize turbidity. Once the stream segment is rewatered, the upstream and downstream block nets will be removed.

Erosion and sediment control for in-water work for the Burntlog Route will be consistent with controls used for other aspects of the project (see below, Section 1.11, and Appendix A, Tables A-1, A-2, A-5). Turbidity monitoring and protocols will include:

- Turbidity monitoring will be required and shall be completed in accordance with designated protocols (for the type of planned work);
- Work will be performed in a manner that does not cause turbidity exceedances within the waterway;
- If turbidity exceedances do occur, the work will stop to address the turbidity issues; and,
- Construction discharge water will be collected to remove debris and sediment and will meet turbidity requirements for discharging back to receiving streams.



# **Table 5. Design Information for Fish Crossing Bridges and Culverts.**

Sediment controls will include the implementation and use of the following as needed in appropriate locations:

- Instream work will conform with the work, turbidity, and dewatering procedures as specified in design conservation measures (RioASE 2023) and adhere to Bonneville Power Administration Habitat Improvement Program conservation measures;
- Placement of fine mesh silt fences and straw waddles;
- Minimization of equipment wet crossings with vehicles and machinery crossing at right angles to the main channel whenever possible;
- No construction equipment stream crossings will occur within 300 feet (ft.) upstream or 100 ft. downstream of an existing redd or spawning fish;
- After construction, temporary stream crossings will be removed and banks restored while adhering to turbidity requirements;
- Cofferdams and diversion structures will have 1-ft. of freeboard;
- Dewatering pump discharge will be released onto floodplain areas away from wetlands and construction activities where discharge will fully infiltrate prior to reaching wetlands and surface waters unless otherwise approved;
- Any return flows from dewatering discharge will meet turbidity requirements;
- Bag fill materials will be clean, washed, and rounded material meeting standard specifications for drain rock, streambed aggregate, streambed sediments, or streambed cobbles; and,
- Work activities within the ordinary high-water channel will conform with the water quality standards established for the project.

### **Road Maintenance Measures**

Road maintenance will be conducted per the project design, USFS requirements, and requirements of the Road Maintenance Agreement with Valley County. Routine maintenance will involve resurfacing, fixing holes, best management practices (BMPs), grading, ditch cleaning and use of traffic signs. For non-routine road maintenance (e.g., activities outside the current road prism, new construction, or reconstruction), the USFS will engage with Valley County when activities involve repairs or new infrastructure that departs from the current roadway and infrastructure footprint. The USFS will focus on enforcement of its existing requirements when activities occur within the current footprint.

To assess road conditions and fulfillment of requirements, the USFS will meet annually with Perpetua to discuss road maintenance needs, road maintenance activities, and best practices that must be employed to minimize impacts to Federally-protected resources. The USFS will present an annual summary of the implemented and planned road maintenance activities to the Interested Agency Review Board (IARB) (see Section 1.11.1 for additional detail). These activities and reports will include the status of road maintenance requirements for the Project including:

• Use of gravel for road surfacing that meets American Association of State Highway and Transportation standards, design specifications for particle size (90 to 100 passing 1 inch, 85 to 95 percent passing 3/4 inch, 70 to 83 percent passing 3/8 inch, 47 to 62 percent passing No. 4 sieve, 27 to 40 percent passing No. 16 sieve, 18 to 27 percent passing No.

40 sieve, and 10 to 16 percent passing No. 200 sieve), percent fracture (75, one face), and plasticity index (4 to 10), and does not rapidly degrade into fine material.

- Avoids side-casting of snow where it has the potential to dam adjacent streams.
- Use of dust suppressants  $MgCl<sub>2</sub>$ , calcium chloride (CaCl<sub>2</sub>), or lignin-based chemicals such as lignin sulfonate with IARB approval required for use of any other dust suppressant.
- Application of dust abatement will be centered in the road so that all the chemical is absorbed before leaving the road surface when the road is within 25 feet of stream channels.
- Avoid installation of berms along the outside edge of roads unless an outside berm was specifically designed to be part of the road and low-energy drainage is provided for.
- Grade and shape roads in a manner that conserves existing surface material and designed drainage.
- Remove fines that cannot be bladed into the road surface by end-hauling to areas outside RCAs for disposal (i.e., no side-casting of materials); slides and rock failures of more than one half cubic yard will be hauled to disposal sites outside RCAs; scattered clean rocks (i.e., 1" plus) may be raked or bladed off the road except within 100 ft. of streams.
- Maintain blocked motorized access on all roads and road segments that are not open to the public, particularly service roads for the power transmission line, and gravel roads through RCAs when the roads will be used daily or used by heavy equipment.

#### **1.6.4. Public Access**

During construction of the SGP and completion of the Burntlog Route, to the degree practicable, the public will continue to have access on forest roads currently available to the public (Figure 3) and following construction road use with seasonal restrictions per current conditions will return. A new 4-mile long, 12-ft.-wide gravel road will be constructed to provide public access from Stibnite Road (FR 50412) to Thunder Mountain Road through the SGP (Figure 4). The road will be constructed on a widened bench on the west side within the YPP, then head south of the YPP, where this road will utilize an underpass to cross under an SGP haul road and continue southward, parallel to and on the east side of the mine haul road on a partially revegetated portion of a former haul road (Figure 4). Southwest of the ore processing area, the public access road will connect with Thunder Mountain Road and continue toward the worker housing facility, exiting the SGP to the southeast.

During operations, the public access road through the SGP will provide seasonal use, open to all vehicles; access will not be provided in winter when impassable (current county maintenance standards) and signs will inform the public of seasonal and temporary closures.

Public access will continue along Johnson Creek Road and Burntlog Road. Total closures of half-day to multiple days could occur during construction work on Stibnite Road between the village of Yellow Pine and the SGP, part of Thunder Mountain Road, and Burntlog Road. The long duration road closures will primarily occur in the mine site area associated with modification of the YPP wall to start construction of the fishway tunnel and construction of a light vehicle underpass below the mine equipment haul road. Public use of the Burntlog Route will provide motorized access to Meadow Creek Lookout Road (FR 51290) and Monumental Summit. Other routes available for public use are shown on Figure 3.

### **Warm Lake to Landmark Groomed OSV Trail.**

Due to year-round access to the SGP along the Burntlog Route, an existing, approximately 8.5 mile-long groomed OSV trail from Warm Lake to Landmark will be closed for the life of the SGP. To replace this recreational use, a dedicated alternative OSV route will be established from Warm Lake area to Landmark via the Cabin Creek/Trout Creek drainages and Johnson Creek Road (Figure 7). The trail will be established largely along existing roads using a snowplow wing attachment requiring some vegetation and tree removal for safe snowplowing.

Near Warm Lake, an approximately 2-acre parking area will be established west of South Fork Road on FR 474B. A new 3.2-mile groomer access trail will be established from the parking area to the USFS Warm Lake Project Camp south of Paradise Valley Road (FR 488) where the groomer will be stored. An approximate 0.1-mile segment will be groomed from the intersection of Paradise Valley Road and FR 488A to Warm Lake Road. The Cabin Creek Road (FR 467) portion of the groomed OSV trail will extend approximately 13 miles to the Trout Creek Campground on Johnson Creek Road. Portions of Cabin Creek Road will require stream crossing improvements, localized road widening, and surface grading to support the OSV route grooming equipment.

### 1.6.4.2. **Johnson Creek Groomed OSV Trail.**

From Trout Creek Campground to Landmark, an approximately 8-mile temporary groomed OSV trail will be created and maintained on NFS lands adjacent to the west side of Johnson Creek Road (CR 10-413). Portions of the temporary groomed OSV trail (approximately 16 ft. wide) will be established using a snowplow wing attachment requiring some vegetation and tree removal to allow for safe snowplowing. In areas where topography and vegetation prevent using the wing attachment to establish the groomed OSV trail, sections will merge with Johnson Creek Road. During construction, the replacement OSV route will include an additional 0.34 of a mile segment east along the Warm Lake Road connecting Johnson Creek Road to Deadwood-Stanley Road (FR 579) (Figure 7).

#### **Warm Lake Area OSV Connection.**

A 16-foot-wide groomed OSV trail will be created and maintained north of Warm Lake Road to connect the southern end of the Cabin Creek Road OSV trail to the Warm Lake Road (FR 579). It will also provide access to North Shoreline Drive (FR 489) from the Cabin Creek Road OSV trail. This 0.3-mile route will be used throughout construction and operations and will require the removal of some vegetation and trees.

#### **Temporary OSV Closure Trout Creek Campground to Wapiti Meadows.**

OSV access will be temporarily halted between Trout Creek Campground and Wapiti Meadows (about 9 miles north of Trout Creek Campground on Johnson Creek Road; Figure 7) for approximately 2 to 3 years during construction of the Burntlog Route. Once construction of the Burntlog Route has been completed, the Johnson Creek Route will no longer be used by minerelated traffic and the OSV route will be returned to the unplowed Johnson Creek Road and extended northward to provide approximately 17 miles of groomed OSV access between

Landmark and Wapiti Meadows. Resumption of OSV access between Trout Creek Campground and Wapiti Meadows will occur following construction of the Burntlog Route.

## **1.6.5. Traffic**

Traffic associated with SGP construction will occur year-round, depending upon road and weather conditions. Construction-related traffic and material hauling will be most concentrated from May through November, and personnel will be transported primarily using buses and vans. The total estimated annual average daily traffic (AADT) for construction activities driving from the SGLF and the SGP is listed in Table 6.

<b>Phase</b>	Route	<b>Transport Type</b>	<b>AADT</b>	
Construction	<b>SGLF</b> to SGP	<b>HV</b>	45	
		LV	20	
<b>Subtotal</b>			65	
Operations	<b>SGLF</b> to SGP	<b>HV</b>	33	
		LV	17	
<b>Subtotal</b>			50	
<b>Reclamation and Closure</b>	<b>SGLF</b> to SGP	<b>HV</b>	15	
		LV	12	
<b>Subtotal</b>			27	

**Table 6. Project Construction and Operations Stibnite Gold Project Traffic.**

AADT – annual average daily traffic; HV – heavy vehicle; LV – light vehicle.

**Note:** Table has been modified from Table 3.4-5 in the BA to include only traffic within watersheds occupied by ESA-listed Chinook salmon or steelhead.

SGLF to SGP - Stibnite Gold Logistics Facility to Stibnite Gold Project via the Johnson Creek Route during construction and via the Burntlog Route during operations.

### **1.6.6. Water Treatment During Construction**

During construction, mine-impacted water will be generated and will require treatment before being discharged to receiving streams. Water treatment plants will be modular, vendor-supplied equipment package skids placed on improved pads with covers and freeze protection for sensitive piping and equipment located in the process plant and YPP work areas to treat for analytes including cadmium, copper, lead, mercury, silver, thallium, zinc, arsenic, and antimony. Peak capacity on-site for construction water treatment requirements is expected to be 300 gallons per minute (gpm) (or 0.67 cubic feet per second [cfs]) with average flows of 18 gpm (0.04 cfs) and 128 gpm (0.29 cfs) during the first and second years of mine site construction, respectively. Water treatment plant residuals will be sent to the TSF for disposal. See Section 1.7.10.1 for additional details.

### **1.6.7. Transmission Line Upgrades**

In order to serve Perpetua's 60-megawatt (MW) load requirement for the SGP, Idaho Power Company (IPC) will rebuild or construct 72.8-miles of transmission line and associated facilities (Figure 8). The existing Cascade to Warm Lake 69-kilovolt (kV) transmission line, and much of the Lake Fork to Cascade and the Warm Lake to Yellow Pine 69-kV transmission lines, will be rebuilt to 138-kV clearances and capacity (Perpetua 2021b)<sup>2</sup>. A new Johnson Creek Substation will be constructed and a new 9.1-mile, 138-kV transmission line will be built between the new Johnson Creek Substation and the new Stibnite Substation at the SGP. The existing single-phase distribution line between the proposed Johnson Creek Substation and the village of Yellow Pine will remain intact. A new single-phase underground distribution line, within the existing road right of way (ROW), will be built along Johnson Creek Road between the Johnson Creek Substation and Wapiti Meadows to the south.

Changes to the existing IPC system for SGP operations will include:

- Upgrade approximately 59.1 miles of the existing 12.5-kV and 69-kV transmission lines between the Lake Fork and Johnson Creek Substations to 138-kV service. The ROW will be 50 to 100 ft. (depending on slope aspect) and existing transmission line support structures will be replaced with taller structures.
- A new approximate 9.1-mile, 138-kV line will be constructed from the Johnson Creek substation to a new substation at the SGP, partially within a former transmission line ROW. The ROW for the new transmission line will be approximately 100 ft. wide. At the SGP, transformers will reduce the voltage from 138-kV to 34.9-kV for distribution to facilities through overhead distribution lines or underground conduits.
- A new substation (Johnson Creek Substation), approximately 0.7 mile south of the Johnson Creek Airstrip on NFS lands, will be built to provide low voltage distribution to Yellow Pine and electricity to the SGP (Figure 8). The substation is outside RCAs (see the Wetlands Specialist Report, Figure 5-5b for location details; USFS 2023c).
- Install approximately 3 miles of new underground power distribution along Johnson Creek Road from the Johnson Creek substation south to Wapiti Meadows. This underground power distribution line is within the existing Johnson Creek Road in a segment that does not cross Johnson Creek (Perpetua 2021b; Maps 59 through 62).

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<sup>&</sup>lt;sup>2</sup> The transmission line from Cascade Dam, including the Cascade switching station, and the Scott Valley substation, are located in the Payette River subbasin, and outside the range of anadromous fish. Because activities in the Payette River drainage do not have the potential to affect Chinook salmon, steelhead, or their designated critical habitat, they will not be further described. Please see the BA (Stantec 2024) for more details about these components of the SGP.



**Figure 8. Transmission Line and Associated Infrastructure.**

Utilities associated with the SGP (existing transmission line upgrades and structure work, ROW clearing, new transmission line, and transmission line access roads) will cross 37 different streams. Of the 37 streams that will be crossed, 26 will be related to the upgrade of existing IPC transmission lines. The existing transmission line currently crosses multiple streams, including Cabin Creek, Trout Creek, and Riordan Creek. The ROW overlaps with 132.4 acres of RCAs. However, the utility poles are not directly along the creeks or within the RCA, and the line is currently kept cleared for access when necessary. Upgrades of these lines, while requiring a wider clearing zone, the effects will be limited to trimming of trees that pose a fire risk to the power line.

The transmission line extends across lands managed by the BNF, PNF, BOR, IDL, and private lands (Figure 8). Table 7 summarizes the transmission line segments by land ownership crossed.

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<b>Line Segment</b>	<b>Total</b>	<b>USFS</b>		<b>Private</b>		<b>State or Local</b>	
	<b>Miles</b>	<b>Miles</b>	$\frac{0}{0}$	<b>Miles</b>	$\frac{0}{0}$	<b>Miles</b>	$\frac{0}{0}$
Cascade Switching Station to Johnson Creek (Ck.)	43.6	31.5	72.2%	6.6	15.1%	5.5	12.6%
Johnson Ck. to Stibnite	9.1	8.7	95.6%	0.4	4.4%		
Johnson Ck. to Wapiti Meadows Distribution (underground)	3.1	2.6	83.9%	0.5	16.1%		

**Table 7. Transmission Line Segment Summary by Land Ownership.**

*Note: Crossings on BOR land are located from Lake Fork to the Cascade Switching Station in the Payette River drainage, crossing streams outside the range of anadromous fish, and are therefore not included in the table. Source: Land ownership derived from parcel data (Valley County 2019).*

Both temporary and permanent disturbances will be required for the construction of the transmission line and substations. While existing structure locations will be used when possible, the removal and installation of new structures will require temporary disturbance. Where possible, single-pole structures will be installed rather than H-frame structures to minimize the structure disturbance footprint. Table 8 lists areas permanently disturbed for each transmission line structure type.

<b>Structure Type</b>	<b>Area Required Permanently</b>			
<b>Single Pole Tangent Structure</b>	16 square feet (sq. ft.), 4-ft. by 4-ft. base			
Single Pole Guyed Structure	28 sq. ft., 4-ft. by 4-ft. base, 3 x multi-helix screw anchors			
H-Frame Tangent Structure	64 sq. ft., 16- ft. by 4- ft. base			
	156 sq. ft., 37- ft. by 4- ft. base			
H-Frame Guyed Structure	Up to 500 sq. ft., for up to 10, 5-ft. by 10-ft. down guy wire plate anchors			

**Table 8. Land Permanently Disturbed for Transmission Line Structures.**

Each transmission line structure site needs a construction space large enough to remove the existing structure, excavate structure foundation holes, and install new structure poles and any guys and anchors. Temporary disturbance is based on a 100-ft. by 60-ft. pad for each structure location. Some temporary disturbance areas will be 100-ft. by 100-ft. pads. Lands affected during construction by line segment and substations and the land status are listed in Table 9.

<b>Line Segment/Project Component</b>	<b>USFS</b>	<b>Private</b>	<b>State or Local</b>	Total <sup>1</sup>	
Cascade Switching Station to Johnson Creek <sup>1</sup>					
Access, Existing (Minor Improvements, 0-50%)	55.0	2.0	0.3	57.3	
Access, Existing (Major Improvements, 50-100%)	65.7	4.4	7.5	77.5	
Access, New (Bladed)	2.8	0.7	1.2	4.7	
Access, New (Overland Travel)	0.9	7.7	1.4	10.0	
Access, Temporary (Minor Improvements, 0-50%)	$-$	1.6	0.2	1.7	
Access, Temporary (Overland Travel)	--	2.0	--	2.0	
<b>Pulling-Tensioning Sites</b>	17.3	4.4	3.1	24.7	
<b>Staging Areas</b>	17.3	9.9	$-$	27.1	
<b>Structures</b>	31.7	12.3	6.4	50.4	
Structures, (Remove Existing)		4.7	< 0.1	4.8	
Substation, Cascade Switching Station	$-$	2.6	--	2.6	
Substation, Johnson Creek	1.1	$-$	$-$	1.1	
Substation (Scott Valley), SGLF	$ -$	0.9	$-$	0.9	
Substation, Thunderbolt Drop Substation	0.1	$-$	$-$	0.1	
Substation, Warm Lake	0.3		$-$	0.3	
Cascade Switching Station to Johnson Creek - Total <sup>1</sup>	192.2	53.2	20.1	265.2	
<b>Johnson Creek to Stibnite</b>					
Access, Existing (Minor Improvements, 0-50%)	10.9	1.1	$-$	12.0	
Access, Existing (Major Improvements, 50-100%)	36.5	1.2	$\overline{a}$	37.6	
Access, New (Bladed)	15.3	0.6	$-1$	15.9	
Access, New (Overland Travel)	--	$- -$	--	$\overline{a}$	
<b>Pulling-Tensioning Sites</b>	6.5	0.5	--	7.0	
<b>Staging Areas</b>	9.7	9.2		18.9	
<b>Structures</b>	8.7	0.9	$-$	9.7	
Johnson Creek to Stibnite - Total <sup>1</sup>	87.6	13.5	$-$	101.1	

**Table 9. Land Affected During Construction by Line Segment/Project Component and Land Status (acres).**

*Source: Land ownership derived from parcel data (Valley County 2019).*

**<sup>1</sup>**Only portions of the transmission line in the South Fork Salmon River and Johnson Creek are occupied by Chinook salmon and steelhead in this transmission line segment.

**Key:** SGLF = Stibnite Gold Logistics Facility.

Lands required permanently for Project operations by route segment and land status are listed in Table 10.



#### **Table 10. Land Permanently Disturbed During Operations by Line Segment/Project Component and Land Status (acres).**

Source: Land ownership derived from parcel data (Valley County 2019).

<sup>1</sup> Only portions of the transmission line in the South Fork Salmon River and Johnson Creek are occupied by Chinook salmon and steelhead in this transmission line segment.

Key: SGLF = Stibnite Gold Logistic Facility

#### **Access Roads**

In addition to the transmission line work detailed above, the existing road network used to access these structures may require maintenance/improvements to allow construction equipment safe access into the power line corridor. While the existing road network proximate to the transmission line ROW will be used to the maximum extent possible, some new service roads (roads used solely by Perpetua or IPC to access Project facilities) could be needed to reach structure locations without existing access (Table 11).

Additionally, overland service routes will be required from the existing access road to reach structure locations without current access. These overland service routes will not require blade work (i.e., recontouring). A 14-foot-wide ROW is being requested for the existing/proposed roads outside of the power line corridor ROW to accommodate construction and maintenance equipment. For FR 467, a 16-foot-wide ROW is being requested to accommodate OSV.

During construction, the new section of transmission line between the Johnson Creek substation and the SGP will require major improvements to Horse Heaven Road (FR 416W) and NFS Trail 233 (no name), and approximately 4 miles of new spur roads will be constructed. Minor upgrades to Cabin Creek Road (FR 50467) will also be required.

Road maintenance requirements prior to construction will vary depending on the type of road, level of use, and condition of the road. However, maintenance generally will consist of clearing vegetation and rocks, as well as repairing cut and fill slope failures, as necessary, to allow for a 14-ft.-wide road surface. In most cases, the roads will be left as close to an undeveloped nature (i.e., two-track road) as possible without creating environmental degradation (e.g., erosion or rutting from poor water drainage). Equipment to perform the required road maintenance will include hand tools (e.g., chainsaws), track driven machines (bulldozers and graders) and crewhaul vehicles (such as 4-wheel-drive pickups and/or off-highway vehicles [OHV; includes all terrain vehicles, utility task vehicles, and side-by-sides]). Roads will be opened/cleared for use by trucks transporting materials, excavators, drill rigs, bucket trucks, pickup trucks, and crewhaul vehicles. Specific actions, such as installing water bars and dips to control erosion and stormwater, will be implemented to reduce construction impacts and will follow standard designs.

Access road construction and disturbance can typically be summarized into five types of access roads:

*Existing (No Improvement)* – These existing roads provide access to structures and will not require improvement. Minor maintenance activities such as pruning of vegetation for construction vehicle access and applying water to the road to reduce dust may be required.

*Existing (Minor Improvement) –* These existing roads provide access to structures and should not require significant improvement to utilize for construction. Existing road widths typically vary from 14-ft.-wide access roads to 24-ft.-wide gravel roads with 14 ft. being the minimum needed to accommodate construction traffic. Minor maintenance activities such as applying water to the road to reduce dust and improve workability of the soil for blading and compaction, and blading may be required during and after construction to support construction traffic and return the road to a preconstruction condition.

*Existing (Major Improvement) –* These existing roads provide access to the structures and may require major reconstruction work. These roads appear to be in questionable condition and will likely require major reconstruction to support construction traffic. Existing road widths may be as narrow as 8 ft. for primitive two-track roads that need reconstruction to widen the driving surface to 14 ft., with curve widening and turnouts added to accommodate construction traffic. Overall disturbance width is estimated to be an average of 20 ft., which includes cut/fill slopes and other impacts associated with reconstruction. Maintenance activities such as applying water to the road, to reduce dust and improve workability of the soil, and blading may be required during and after

construction to support construction traffic. Aggregate/crushed rock placement may be required to maintain the existing road.

*New (Overland Travel) –*These roads traverse existing agricultural fields or open areas and are not expected to require grading work to support construction traffic. No permanent road construction is anticipated on these routes, and any earthwork or aggregate imported will be reclaimed after construction. Temporary driving surface is estimated to be 14 ft. to accommodate construction traffic. Sections of road that cross wet fields or wetlands may have temporary matting installed to provide a stable surface to support construction equipment without disturbing the ground. Minor work such as grade smoothing at ditches or large rock removal may be required to provide a drivable surface.

*New (Bladed) –* New bladed roads are typically required where the existing ground has a significant cross slope or traverse's terrain that needs to be bladed smooth. Construction of the road prism will require excavation and placement of fill material to provide a stable driving surface. The driving surface is constructed to a minimum width of 14 ft. and includes curve widening and turnouts to accommodate construction traffic. Overall disturbance width is estimated to be an average of 35 ft., which includes cut/fill slopes and other impacts associated with construction. Earthwork quantities are typically balanced for each road by adjusting the grade to balance material being cut versus filled. Surfacing rock is not typically placed on these roads unless required by stakeholders or needed to support construction traffic.

Table 11 provides a summary of miles of access roads by route segment and land status.

Line Segment/Access Type		<b>USFS</b>	<b>Private</b>	<b>State or</b> Local	Total <sup>1</sup>		
<b>Cascade Switching Station to Johnson Creek</b>							
Access, Existing (No Improvements)	--	5.1	4.2	4.6	13.9		
Access, Existing (Minor Improvements, 0-50%)	$- -$	18.9	0.7	0.1	19.7		
Access, Existing (Major Improvements, 50-100%)	--	17.8	1.1	2.0	20.9		
Access, New (Bladed)	$- -$	0.6	0.1	0.3	1.0		
Access, New (Overland Travel)	$- -$	0.4	4.0	0.7	5.1		
Access, Temporary (Minor Improvements, 0-50%)	$-$		0.5	0.1	0.6		
Access, Temporary (Overland Travel)	$-$		1.0		1.0		
Cascade Switching Station to Johnson Creek - Total <sup>1</sup>	$-$	42.8	11.6	7.7	62.1		
<b>Johnson Creek to Stibnite</b>							
Access, Existing (No Improvements)	$-1$	< 0.1	0.7	$-$	0.7		
Access, Existing (Minor Improvements, 0-50%)	--	3.7	0.4		4.1		
Access, Existing (Major Improvements, 50-100%)	$- -$	10.1	0.3	--	10.3		
Access, New (Bladed)		3.5	0.1		3.7		

**Table 11. Miles of Access Roads by Line Segment and Land Ownership.**



Source: Land ownership derived from parcel data (Valley County 2019).

<sup>1</sup> Totals may not sum correctly due to rounding.

#### **Substations**

Several substations along the transmission line from Cascade to Yellow Pine will need to be upgraded from 69-kV to 138-kV. A 138-kV metering substation will be placed in the Johnson Creek area to feed the village of Yellow Pine and serve as a metering point for the Stibnite 138 kV line. The substations will be operated and maintained by IPC. Table 10 provides the area that is needed, by land ownership, for each of the substations. Additional details regarding the upgrades needed to existing substations and the construction of new substations are available in the Electrical Transmission POD (Perpetua 2021b).

Routine operation and maintenance activities that may be conducted by IPS as necessary and without prior notification to the USFS include: routine inspections (air and ground), clearing of vegetation (including hazard tree removal) to prevent encroachment into the minimum vegetation clearance distance, access road inspection and routine maintenance, wood treatment to retard rotting, and application of fire retardant to base of poles. A more detailed description of these activities is provided in Section 3.4.7.9 of the BA (Stantec 2024).

Substation maintenance activities will include equipment testing, preventative repair, and procedures for providing continual service and maintaining electrical service. Typical substation maintenance does not require ground-disturbing activity, although ground disturbance could be required to replace damaged equipment, oil containment facilities, or other miscellaneous items.

#### **1.6.8. Off-site Facilities**

Perpetua will require off-site facilities to support mine-related activities, of which only the Burntlog Maintenance Facility is located within the SFSR subbasin. This facility will support. road maintenance and snow removal activities. This facility will be located on NFS land within a previously disturbed borrow source site 4.4 miles east of the junction of Johnson Creek Road and Warm Lake Road (Figure 3) and will be accessed via the Burntlog Route with two points of ingress/egress. The facility footprint will be approximately 3.5 acres and will include three main buildings: a 7,000-square-foot maintenance building; a 7,000-square-foot aggregates storage building; and a 4,050-square-foot equipment shelter (Figure 9).



**Figure 9. Burntlog Maintenance Facility.**

The Burntlog Maintenance Facility will also contain a fuel station, electric generator, propane tank, outdoor storage area, and worker sleeping quarters. It will house sanding/snowplowing trucks, snow blowers, road graders, and support equipment in the equipment shelter or maintenance buildings. The Burntlog Maintenance Facility will require a domestic groundwater well to service the facility. This well and associated water right will require permitting through the IDWR.

This facility will include a double-contained fuel storage area housing three above-ground 2,500gallon fuel tanks for on-road diesel, off-road diesel, and unleaded gasoline. Additionally, a 1,000-gallon used oil tank will be located inside the maintenance facility and a 1,000-gallon propane tank will be located at the facility for heating.

Additional features of this facility could include covered stockpiles of coarse sand and gravel for winter sanding activities; temporary or emergency on-site housing for road maintenance crews during periods of heavy snow removal needs and other winter maintenance activities; and communications equipment including a tower. This facility could also serve to support snowmobile trail grooming and grooming equipment storage as needed.

### **1.7. Mine Operations**

The SGP will consist of mining three primary mineral deposits and the re-mining of historical tailings using conventional open pit shovel and truck mining methods. Ore from three open pits (Yellow Pine, Hangar Flats, and West End Pits) will be sent to either the crusher, located near the processing plant, or one of several ore stockpiles in various locations within the Operations Area Boundary (Figure 4) (M3 2021). Pre-stripping, or removing the overlying soil and rock (i.e., development rock) to access the mineral deposit, will commence during the construction phase in MY -2. Ore removal and processing will begin in MY 1 (operations phase) and continue year-round for approximately 15 years. Mine operations will occur in the area of two historical open pit mined areas (Yellow Pine and West End) and one new open pit (Hangar Flats) that includes former underground mining and mineral processing facilities.

In general, ore mined from the three open pits will be hauled directly to the primary crusher area; however, during extended periods when the ore tonnage or ore type from the pits exceed the availability of the ore processing plant, the ore will be stockpiled and processed at a future time. Development rock (also commonly referred to as waste rock) will be hauled to the TSF embankment or placed in one of four destinations: the TSF Buttress or the Yellow Pine, Hangar Flats, and West End open pits once they are mined out.

### **1.7.1. Open Pits**

Figure 4 shows the location and extent of the three pits to be mined. A general sequence for mining, assuming 15 years of mine operations as shown on Figure 5, will be as follows:

- $\bullet$  YPP MYs 1 through 7
- $\bullet$  HFP MYs 4 through 7
- West End Pit (WEP)– MYs 7 through 12
- Stockpile mining  $-MYs$  12 through 15

The YPP will be in the northern portion of the SGP, in the same general location as a historical open pit mining area. The pit will be expanded to include a shallower mining area to the northeast previously mined as the Homestake Pit. The EFSFSR currently flows through the legacy YPP, forming a small pit lake (YPP Lake) when the EFSFSR flowed into the pit after it was abandoned in the 1950s.

The WEP will be in the northeast portion of the SGP, east of and at a higher elevation than the YPP, generally situated between Sugar Creek to the north and Midnight Creek to the south. The WEP will be in the same general location as historical open pit mining where multiple open pits, mine benches, waste rock dumps, and areas of deep backfill exist. The existing Stibnite Pit is within the southern portion of the WEP, and once expanded will be known as the Midnight Pit.

The HFP will be in the central portion of the SGP, generally encompassing steep south and southeast facing slopes and the adjacent Meadow Creek valley floor at the toe of these slopes. Historical mining activity in this area was primarily underground but the proposed pit also will encompass the site of the former Bradley mill and smelter, the Hecla heap leach, and Stibnite Mine Inc. leach pads. Table 3.5-1 in the BA (Stantec 2024) provides a summary of characteristics for each pit.

Partial dewatering of the open pits will occur prior to and concurrent to renewed SGP mining. Shallow alluvial and deeper bedrock wells will be drilled adjacent to the pits to intercept and pump groundwater before it flows into the pits. During mine operations, groundwater seepage and in-pit surface water runoff will be collected for reuse in the ore processing plant or treated and discharged, according to whether there was a water deficit or surplus at a given time. Additional details on pit water management can be found in Section 1.7.10.

#### **1.7.2. Drilling and Blasting**

Drilling and blasting will be used to break ore and development rock in the mine pits (see M3 2021 for additional details as well as Section 3.9 in the BA for information on controls). Following drilling, explosives will be used to break rock into fragments that are suitable for loading into equipment. An Explosives and Blasting Management Plan will be prepared as part of the final mine plan. This plan will include blasting measures techniques, charge sizes, and setbacks to minimize effects on fish and wildlife as summarized below and described in the FMP. Explosives storage, transport, handling, and use will comply with applicable Department of Homeland Security, Bureau of Alcohol, Tobacco, Firearms and Explosives, Department of Transportation, and Mine Safety and Health Administration (MSHA) regulations.

● Perpetua will employ blasting setback distances and other controlled blasting techniques following industry BMPs (modifying blasting variables including charge size, and vibration and overpressure monitoring) to minimize impacts to fish from blasting. Perpetua will follow up with monitoring in early stages of operation to evaluate effectiveness and refine blasting protocols in coordination with federal, state, and tribal agencies, if needed. Blasting setbacks are described in the FMP. Blasting for the YPP will be in proximity to the EFSFSR while blasting for the HFP will be in proximity to the Meadow Creek diversion, above its confluence with the EFSFSR.

- Blasting peak particle velocity will be  $\langle 7.3 \text{ pounds per square inch (psi) (50 kilopascals)}$ [kPa]) where fish are present.
- Blasting airblast overpressure will be  $\lt$  2.0 inches per second (in/s) (51 mm/s) during sensitive stage (embryo incubation before epiboly is complete).
- Required setbacks for blasting are set to meet maximum overpressure and maximum peak particle velocity, including a 239-ft. blasting setback on 20-ft. benches, and 419 ft. setback on 40-ft. benches (as measured from the closest point the blast field to stream and lake habitats).
- Perpetua will develop an Explosives and Blasting Plan that will ensure compliance with the blasting requirements of the MSHA, 30 CFR Part 56, Subpart E – Explosives and Part 57, Subpart E – Explosives. The blasting plan will include the setback distances and options for other mitigative measures and BMPs.
- Observe setback distances for each blasting activity, wherever possible. In the event a blasting activity will not meet the required setback distance, Perpetua will attempt to adjust the bench height or blast intensity to minimize potential adverse effects to fish communities in the nearby stream. Where the setback distance cannot be met and alterations to the blasting protocol will not adequately mitigate potential harm to fish communities, Perpetua will implement measures to isolate, capture, and relocate ESAlisted fish species from the stream segment where potential for impact exists.

Perpetua developed a spreadsheet tool to compute the required setback distances from fishbearing streams and lakes. The spreadsheet tool was developed using the following steps:

- 1. The equations used in the spreadsheet were taken from Guidelines for the Use of Explosives in or Near Canadian Fisheries Waters (Wright and Hopky 1998), including equating the peak particle velocity to charge weight and distance.
- 2. The standards used came from the Alaska Blasting Standard for the Proper Protection of Fish (TR 13-03) (ADFG 1991) described above—the blast overpressure threshold limit was set at 7.3 psi and peak particle velocity threshold was set at 2.0 in/s.

The spreadsheet tool was then populated with the anticipated drill and blast assumptions (bench height, drill hole diameter, stemming length, powder column height, powder volume and charge per hole [weight]).

Perpetua used these required setback distances to do a high-level review of streams and lakes in closest proximity to areas where blasting may be required. This review identified areas where the blasting may be within the 239-ft. and 419-ft. setbacks. These include some stream segments adjacent to the EFSFSR diversion tunnel, YPP, WEP, the TSF and HFP where Meadow Creek is closest to the pit. This analysis does not definitively identify areas where impacts will occur but points to areas where adjustments to blasting methods may be needed to reduce the required blasting buffer (bench heights, charge size, and detonation pattern, etc.). According to the proposed implementation, such areas will be the locations at which initial calculation and testing of blast effects (instantaneous pressure and peak particle velocity) will be conducted. It is also noteworthy that blasting in some of the areas identified may not occur until non-consolidated materials are removed without blasting, such that the distance between blast sites and streams will have increased. The predicted setback distances will be verified via blast monitoring

instrumentation (blast seismographs and pressure transducers) utilized for test blasts at three sites with different physical conditions.

### **1.7.3. Rock Hauling and Storage**

Rock loading and haulage will use a development fleet and a production mining fleet. Mine development excavation required to establish haul truck access roads, access limestone, and prestrip pits prior to production mining will use a fleet of medium sized excavators, wheel loaders, and 45-ton articulated trucks. This development fleet will also be used to salvage GM and support reclamation activities. Production mining will use a conventional diesel truck and shovel fleet consisting of two 28-cubic yard hydraulic shovels, approximately sixteen 150-ton haul trucks, and one 28-cubic yard wheel loader. The wheel loader will be used primarily to load haul trucks during shovel maintenance and to load stockpiled ore as needed. The ore will be hauled directly to the primary crusher or the run-of-mine ore stockpile at the ore processing facilities.

### **1.7.4. Ore Management**

Ore from the open pits will be hauled to and placed directly into the ore processing plant, except during periods when the amount or type exceeds the availability of the ore processing plant, the excess ore will be stockpiled in unlined facilities on top or within other mine disturbance areas. Seven long-term ore stockpiles and one short-term stockpile will be used to manage the excess ore (Figure 4). The long-term ore stockpiles will be located on and near the TSF Buttress and HFP and the short-term stockpile will be located near the crusher. The short-term stockpile will hold ore that will be processed within weeks while the long-term stockpiles will hold ore for a period of months to years until the process has the capacity to receive the stored ore.

Highest-grade ore will be sent directly to the crusher, or to the short-term stockpile area near the crusher where it will likely be processed within a few days. Lower-grade ore will be sent to the long-term ore stockpiles where it will remain for months or longer. Some of the ore sent to the low-grade ore stockpiles will be re-handled during active mine operations, and some will be rehandled and processed once open pit mining has ceased. If metal prices do not support processing of some of the long-term stockpiles, the stockpiled material will be covered as part of TSF Buttress closure activities (Section 1.9).

Three long-term ore stockpiles will be on the TSF Buttress on the north side of the valley. Two stockpiles will be adjacent to the HFP and extended onto the pit footprint after it is backfilled. A stockpile within the WEP footprint will temporarily store ore mined during West End Road development and pre-stripping. Ore storage in long-term stockpiles peaks in Year 11 with approximately 19 million tons.

### **1.7.5. Development Rock Production and Storage**

Development rock from the three open pits will be sent to five different permanent destinations over the mine life including the TSF embankment and rind fills; the TSF Buttress; the mined-out YPP; the mined-out HFP; and the Midnight area within the mined-out WEP. In addition to these five areas, other destinations will receive development rock from the three open pits including a temporary ore stockpile base within the WEP, a foundation for stockpiling growth medium and

recovered seed bank material, a reclamation materials stockpile located on the TSF Buttress, and miscellaneous projects such as road fills and ore stockpile foundations. Approximately 280 million tons of development rock from active mining areas will be used to construct the TSF embankment and buttress, and placed in the mined-out pits, as described in BA Table 3.5-2 (Stantec 2024).

After the main portion of the YPP Pit has been mined and mining commences in the northern portion of the pit, development rock will be end-dumped into the YPP as backfill. The dumped development rock will not be mechanically compacted, except as it nears the final reclaimed surface elevation of the backfilled area.

The upper lifts of the backfill will be placed by direct dumping and compaction. The final backfill will be covered with a geosynthetic liner and soil/rock cover, and the EFSFSR and Stibnite Lake will be established across the backfill in a geosynthetic-lined stream/floodplain corridor. The inclusion of the lined Stibnite Lake on the YPP backfill will help buffer temperature extremes in the EFSFSR and replace the fish habitat of the existing YPP Lake. The 16-million-gallon lake feature was designed based on results of lake temperature modeling to reduce diurnal temperature fluctuations, in particular, to lower the maximum temperature. Development rock to backfill the YPP will be sourced predominantly from the WEP, with minor quantities originating from the YPP and HFP.

Upon construction of the Stibnite Lake, the lake feature will be filled with 16 million gallons of water diverted from the EFSFSR upstream from the tunnel location. This diversion will flow through the restored portion of the EFSFSR located on top of the YPP backfill until entering Stibnite Lake. Once Stibnite Lake is filled, it will outflow to another segment of restored stream channel on top of the YPP backfill which subsequently enters the EFSFSR channel north of the YPP. The diverted flow rate will be a portion of the total EFSFSR flow for a period of several weeks to minimize sediment generation from the restored stream channel and to maintain flows in the EFSFSR and tunnel to support fish habitat and passage. The diverted portion will be based on the available flow in the EFSFSR while maintaining habitat and passage and will fill Stibnite Lake feature slowly (i.e., a 1 cfs diversion will require approximately 24 days to fill the feature).

Once mining ceases at the HFP, development rock to backfill the HFP will be sourced predominantly from the WEP. The Midnight Pit, a portion of the WEP in the southeast corner of the pit near Midnight Creek, will be backfilled concurrent to mining the WEP, with development rock from the WEP once mining in the area to be backfilled is completed.

In addition to the permanent development rock storage described above, a temporary development rock storage facility (DRSF) will be constructed within the WEP during road construction and pre-stripping activities. This temporary DRSF will contain approximately 2.5 million tons and serve as the base for the West End In-Pit stockpile. Since this is a temporary DRSF entirely within the footprint of the WEP, material will be rehandled during regular mine operations at the WEP and relocated to other facilities for permanent development rock storage.

Perpetua has conducted geotechnical investigations supporting the design of the development rock backfills. Because backfills will be below grade, they will not be susceptible to mass failure events in the post-closure period. Development rock in the above grade TSF embankment will be placed per a design with a 5.85 Factor of Safety that will not be susceptible to mass failure events in the post-closure period (Tierra Group 2021).

Surface water and groundwater management for facilities that permanently store development rock are discussed in Section 1.7.10, Surface Water and Groundwater Management. The Development Rock Management Plan describes procedures and methods for mining, haulage, and placement of development rock that is produced and stored across the SGP during operations (Brown and Caldwell 2022).

### **1.7.6. Spent Ore and Legacy Tailings Removal in Meadow Creek Valley**

The Meadow Creek Valley contains legacy materials created from historical mining activities. Legacy materials include development rock, spent ore in the unlined spent ore disposal area (SODA), the Bradley Mill Tailings, and run-of-mine and crushed ore in the historical lined heap leach pads. An Environmental Legacy Management Plan (Perpetua 2021c) describes procedures and methods for active management of legacy materials encountered during construction and mining operations. While the TSF is being built and expanded, Perpetua will remove and reuse as construction material the 7.5 million tons of spent ore within the unlined SODA and other areas (Hecla and Stibnite Mine Inc. leach pads). Physical and chemical testing of the legacy material will determine if the material is suitable for construction uses (e.g., TSF starter dam material) and determine the final placement of the material. Legacy tailings removal will be a component of early-Project ore processing using water to mobilize legacy tailings and collecting excess water in the SODA contact water pond. The water will be initially sourced at approximately 800 gpm (1.7 cfs) from dewatering wells, industrial supply wells, and the EFSFSR freshwater intake, then recirculated with an expected reclaim efficiency of 80 percent. Water not reclaimed will be entrained in the tailings within the TSF or lost to evaporation. The temporary water addition and pumping facility to make up for entrainment and evaporation will be an enclosed, heated structure located within the limits of the SODA.

The legacy tailings will be pumped to the ore processing facility. During the first four years or so of ore processing operations, Perpetua will remove and reprocess the three million tons of Bradley tailings underlying the SODA using approximately 1.5 million gallons of water recirculated daily per the water usage forecasts for the overall Project. If other legacy materials are encountered during construction they will be removed and hauled off site to an appropriate disposal facility, placed in the TSF, used as pit backfill or construction material, or left in place, depending on testing to determine physical and chemical suitability. Physical suitability will be based on the material's geotechnical characteristics (e.g., grain size, shear strength) compared to the geotechnical specifications of the facility at their location. Chemical suitability will be based on the potential for leaching of the materials to affect water quality (i.e., acid-base accounting and kinetic geochemical testing) as described in the Environmental Legacy Management Plan (Perpetua 2021c). Legacy development rock not used for TSF construction purposes or reprocessed will be placed in pit backfills or used for the TSF Buttress.

### **1.7.7. Ore Processing**

During operations, approximately 115 million tons of ore will be mined from the three proposed pits and processed at the mill facilities during the approximately 15-year process facility operation. At full operation, targeted ore production will range from 20,000 to 25,000 tons per day, which will be transported to the processing facility to separate the gold, silver, and antimony from the ore. The ore processing is summarized in the following sections. Additional details on ore processing can be found in section 17 of SGP's updated feasibility study (M3 2021).

Ore feed for processing will be sourced from either the open pits, Bradley tailings, the SODA, the short-term stockpiles, or long-term stockpiles. The ore processing flow sheet is shown on BA Figure 3.5-1 (Stantec 2024). The ore processing facility and associated support infrastructure are shown on Figure 4.

The ore processing area will be designed to provide for containment of ore processing materials, chemicals, wastes, and surface runoff. Potentially hazardous chemicals and wastes will be stored within buildings or areas with both primary and secondary containment. Surface runoff within the ore processing area will be directed to a contact water pond for collection. Any leaks or spills escaping both primary and secondary containment will flow to the contact water pond for collection and will not discharge off site.

The processing will result in production of an antimony mineral concentrate, gold- and silverrich doré, tailings, and other waste products (e.g., small quantities of other non-saleable metals recycled back into the process). Tailings disposal is discussed in Section 1.7.8, Tailings Storage Facility.

### **Crushing and Grinding**

Ore will be hauled to the crusher where it will be crushed and ground to reduce the size of the rock to separate the gold, silver, and antimony-bearing minerals from the host rock. Mined ore will typically be direct-dumped into the jaw crusher or stockpiled at the uncovered run-of-mine stockpile area near the crusher. Stockpiled ore will be loaded into the crusher dump pocket, based on crusher availability, using a loader. The residence time for material in the stockpile will be short (i.e., days); therefore, there will not be sufficient time for infiltration through the stockpile material to the subsurface. Surface water runoff from the run-of-mine ore stockpile area will be captured and directed to a pond and be used in the ore processing facility (Section 1.7.10).

Following crushing, the crushed ore will be transported via conveyor to a dome-shaped, covered stockpile. Dust emission controls, such as water sprays and/or bag house dust collectors, will reduce dust from crushing, conveying, and stockpiling. Apron feeders below the crushed ore stockpile will convey the ore to a semi-autogenous grinding mill followed by a ball mill for additional size reduction of the ore. Grinding will occur within an enclosed building to reduce noise levels. Grinding with process water will reduce the ore to the size of fine sand in a water slurry for further processing.

### **On-site Lime Generation**

Ground limestone and lime are needed for pH adjustment in the SGP ore processing plant. Rather than trucking these materials to site from an off-site source, a limestone bed in the WEP is of suitable quality and quantity to satisfy the life-of-mine SGP requirements for lime. Over the life of the mine, approximately 130,000 to 318,000 tons of limestone will be mined annually, averaging approximately 240,000 tons per year. Approximately 25 to 30 percent of the limestone mined annually will be crushed and run through an on-site lime kiln to produce metallurgical lime powder, with the remainder (70 to 75 percent) will be crushed and stockpiled for direct use as limestone. Both ore and limestone will be temporarily stored at the run-of-mine stockpile area.

The on-site lime generation will require additional equipment, which will be placed within the ore processing area. This equipment will include: limestone crusher and conveyor, propane-fired kiln (200 tons per day output capacity), kiln combustion air system including preheat heat exchanger, propane storage tank plus vaporizer, air compressor, receivers, and dryers for plant air and instrument air at kiln area, roll crusher for kiln product discharge, conveyors for moving feed and product materials, off-gas fume filter for kiln discharge, dust collector kiln feed bin, storage bin for kiln feed material; and storage bin for lime products. The limestone crusher, screens, conveyors, and feed bins will not be enclosed. Dust will be controlled in a similar manner to the ore crushing and conveying process through the use of water sprays and/or bag house dust collectors.

### **Antimony Flotation**

Two flotation circuits will be utilized; one circuit produces an antimony concentrate, and the other produces a gold-rich sulfide concentrate. Ore high in antimony will be processed by the antimony circuit to produce an antimony concentrate (M3 2021). Following grinding, the ground ore slurry will be mixed with lime and small amounts of sodium cyanide or equivalent to inhibit flotation of the gold-bearing minerals (pyrite and arsenopyrite). Lead nitrate or equivalent will be added and then a sulfur- and phosphate‐bearing organic chemical. These chemicals make the stibnite mineral particles hydrophobic where the particles then attach to air bubbles and float to the surface in the stibnite flotation tanks. The gold-bearing mineral particles which do not adhere to the bubbles in the stibnite flotation tanks will drop to the bottom of the flotation tanks and be routed to the subsequent gold flotation circuit for further processing. The antimony flotation facility will have interior curbing high enough to contain 110 percent of the volume of the largest tank.

The stibnite-laden bubbles form a froth and will be collected from the top of the stibnite flotation tanks. The stibnite concentrate froth will be subjected to one or two additional flotation steps to further clean the concentrate. The resultant antimony-rich concentrate will be finally thickened and filtered. The final antimony concentrate will be placed in 2-ton supersack containers ready for shipment off site for further refining.

#### **Antimony Concentrate Transport**

The antimony concentrate will contain approximately 55 to 60 percent antimony by weight. The remaining balance, 40 to 45 percent by weight, of the concentrate includes sulfur and common

minerals with trace amounts of gold, silver, and mercury. As described in the Transportation Management Plan (TMP) (Perpetua 2022) for transportation of antimony concentrate, Perpetua will load the sealed 2-ton super sacks containing the concentrate into a shipping container at the processing facility. Perpetua will load the concentrate by forklift and hooked lifting racks to safely move the super sacks, which are equipped with lifting straps, into fully enclosed shipping containers for the full course of their transport from the SGP site to their final destination. The supersacks and shipping container will provide primary and secondary containment for the antimony concentrate (Perpetua 2022). The concentrate will be trucked via SH 55 to a commercial truck, train, barge, ship loading facility depending on the refinery location. An estimated one to two truckloads of antimony concentrate will be hauled off site each day. It is assumed that the concentrate, when sold, will be shipped to facilities outside of the U.S. for smelting and refining because there are currently no smelters in the U.S. with capacity for refining the antimony concentrate.

### **Gold and Silver Flotation**

Low-antimony mill feed will be processed in the gold flotation circuit only, bypassing the antimony circuit (M3 2021). Gold and silver flotation are a process similar to that described for stibnite flotation, and will be housed in the same building, but using different chemicals to float pyrite and arsenopyrite, the minerals that contain the gold and silver. The flotation building will have interior curbing high enough to contain 110 percent of the volume of the largest tank. The flotation froth, with particles containing gold and silver, will be collected and pumped to the gold concentrate thickener to further separate the gold/silver mineral particles from the process water which will be recycled. The particles from gold flotation that do not float will become the tailings slurry. The gold and silver concentrations of the tailings will be regularly monitored and, if the concentrations are high enough to warrant further processing, they will be sent to the leaching circuit; otherwise, the tailings will be thickened to recycle additional process water and then routed to the TSF as described below.

#### **Oxidation and Neutralization**

An autoclave pressure-oxidation system will be used to oxidize the gold- and silver-bearing sulfide minerals comprising the gold and silver concentrate to liberate the gold and silver for subsequent leaching. Before the gold concentrate is pumped into the autoclave, it will be mixed with appropriate amounts of ground limestone to maintain a constant free acid level of approximately 10 grams per liter in the autoclave. This value was established through bench and pilot-scale metallurgical testing to promote the formation of stable, crystalline arsenic compounds in the autoclave. Oxygen will be injected into the autoclave to promote the oxidation reaction, and the temperature in the autoclave will be maintained at approximately 220°C. Water will be injected into the autoclave as needed to control the temperature. After pressure oxidation, the acidic slurry containing gold and silver will be neutralized using slurried lime and other chemicals and cooled in two forced draft cooling towers. The neutralized slurry will then be sent to the leach circuit for recovery of gold and silver from the slurry.

When increasing arsenic levels are observed, the oxidized slurry will be treated with hot arsenic cure (HAC) prior to neutralization. Metallurgical tests showed that this process promotes

formation of the stable crystalline form of the arsenic precipitate enhancing environmental stability of arsenic.

The autoclave system will be housed in a steel frame building set on concrete foundations, with interior curbing to provide secondary containment. Air emissions from the pressure oxidation facility will be captured in a series of air pollution controls, and the material collected will be disposed of as a solid waste or a hazardous waste depending on the waste characterization.

### **Gold and Silver Leaching and Carbon Adsorption**

The gold and silver leaching component of the recovery process will be regulated by the Idaho Department of Environmental Quality (IDEQ) under the Cyanidation Rule (Idaho Administrative Procedures Act [IDAPA] 58.01.13) and will be designed and operated consistent with the International Cyanide Management Code for the Manufacture, Transport, and Use of Cyanide in the Production of Gold (Perpetua 2021a). Gold and silver leaching and carbon adsorption will occur in a steel frame building set on concrete foundations, with secondary containment of 110 percent of the volume of the largest tank and could include audible alarms, interlock systems, and/or sumps, as spill control measures (Initiative for Responsible Mining Assurance [IRMA] 2018).

The leaching to recover gold and silver from the oxidized gold and silver concentrate slurry will occur in large carbon-in-pulp tanks which will be fully contained to capture, retain, and recycle process solutions. Sodium cyanide will be added to the tanks containing the neutralized solution to form a gold-silver-cyanide complex and activated carbon will then be added to the tanks to promote the adsorption of the gold-silver-cyanide complex onto the carbon (BA Figure 3.5-1) (Stantec 2024). The pH of the slurry in the leach circuit will be closely managed at an elevated level to maintain the cyanide in a stable soluble form.

The loaded carbon with gold-silver-cyanide complex attached will then be collected on screens and sent to the carbon stripping circuit. Inside sealed tanks, the carbon with the gold-silvercyanide complex will be washed with an acid solution to remove impurities, rinsed with fresh water, and stripped of the gold using a hot alkaline elution solution. The resulting gold and silver-bearing elution solution will be piped to the electrowinning and refinery area.

The acid solution used during carbon stripping will be reused until it loses its effectiveness. The solution will be neutralized and sent to the tailings thickener for pumping to the TSF. Air emissions from the leaching facility will be captured in a series of air pollution controls, and the material collected will be disposed of as a solid waste or a hazardous waste depending on characterization of the waste.

#### **Gold and Silver Electrowinning and Refining**

The gold and silver electrowinning and refinery facility will be a closed-circuit system with 110 percent containment of the largest vessel. The elution solution pumped into electrowinning cells which will electrolytically precipitate the precious metals into a solid sludge that will be removed from the elution solution with a filter. The solid precipitate will then be heated in a retort system to drive off and collect any contained mercury. The gold and silver precipitate from the retort

will then be mixed with flux and then placed into an induction furnace and heated. The molten material from the induction furnace, consisting of gold and silver metal and slag, will be poured into molds to cool. The slag will be recycled within the mill circuit and the doré gold/silver bars will be shipped off site to refineries for further processing and refining.

Air emissions from the induction furnace and retort will be treated in a series of emission controls. Mercury metal will be securely stored prior to shipment to a certified hazardous waste disposal facility.

# **Tailings Neutralization Circuit**

Cyanide-bearing process slurry from the carbon-in-leach circuit will be neutralized within the ore processing plant to less than approximately 10 milligrams per liter weak acid dissociable cyanide before being pumped to the TSF. Residual cyanide in the slurry will be treated using a sodium metabisulfite and air system to oxidize cyanide to form cyanate. After neutralization, tailings will be routed to one or more tailings thickeners, to partially dewater the tailings before they are pumped to the TSF. The process water separated from the thickened tailings slurry will be recycled within the ore processing facility. The neutralized and thickened tailings slurry will be pumped to the TSF.

### **Tailings Pipeline Maintenance Pond**

Lined tailings pipeline maintenance ponds will be located at the truck shop and at the ore processing facility, to which tailings slurry from the tailings pipeline between the mill and the TSF and process water from the tailings reclaim pipeline could drain by gravity during maintenance shutdowns or if there were a leak in either pipeline. The ponds will typically be empty except during maintenance or unforeseen problems with the tailings or reclaim water pipelines, pumping system, or TSF. The ponds are designed to contain the contents of the pipelines and the runoff from the pond and open-trench portions of the lined pipeline corridor from a 100-year, 24-hour storm event plus snowmelt.

### **1.7.8. Tailings Storage Facility**

The TSF will be located on NFS lands within the Meadow Creek valley (Figure 4). The TSF, its embankment, and associated water diversions will occupy approximately 423 acres at final buildout with approximately 405 acres of new disturbance. Perpetua has conducted geotechnical and geophysical investigations to support the design of the TSF and associated buttresses. The TSF at the end of operations will be capable of holding approximately 120 million tons of tailings, the operational water pool, and precipitation falling within the TSF and contributing watershed up to the 24-hour Probable Maximum Precipitation event of 11.74 inches of rainfall. Additional details on ore processing can be found in section 18 of SGP's updated feasibility study (M3 2021).

The TSF will consist of a rockfill embankment, a fully lined impoundment, and appurtenant water management features. The TSF Buttress located immediately downstream of, and abutting against, the TSF embankment will substantially enhance embankment stability.
EDFs were established based on the facility size and risk using applicable dam safety and water quality regulations and industry best practice for the TSF embankment on a stand-alone basis; the addition of the buttress substantially increases the safety factor for the design to about double the minimum requirements. The upstream face of the TSF embankment and the Meadow Creek valley, where the TSF impoundment will be located, will be fully lined to minimize leakage. The TSF will be surrounded by an 8-ft. high, chain-link fence designed to keep wildlife, such as deer and elk, from entering the impoundment area.

The TSF includes an engineered, rockfill starter embankment. Historical development rock (i.e., waste rock), spent ore from the historical SODA and heap leach areas, and development rock from mine pits will be used for the TSF embankment construction. The TSF Buttress will be built by first constructing a ramp along the north side of the valley to access the crest of the TSF embankment and upper portions of the buttress (BA Figure 3.5-2) (Stantec 2024). Historical spent ores from the SODA and Hecla heap leach will be placed as bedding on the upstream face of the embankment or impoundment fill prior to placement of the liner to minimize interaction with infiltrating surface water. The starter embankment will be constructed to an elevation of 6,850 feet (or 245 feet above the existing ground surface). The TSF Buttress will then be constructed upwards to further access TSF embankment lifts while the base expands down the valley (eastward) as historical spent ore and legacy tailings are removed from the valley bottom. Engineered fill will be placed against steep slopes within the impoundment to flatten and smooth slopes to facilitate liner placement. This method of construction will allow for controlled material placement across the valley from the ramp north of the valley to the south side. The TSF Buttress will provide additional short- and long-term geotechnical stability. The final embankment height will be 475 ft. at a crest elevation of 7,080 ft. (BA Figure 3.5-3) (Stantec 2024).

# **TSF Underdrain System**

The TSF will have an underdrain groundwater collection and conveyance system located beneath the liner. Prior to construction, the area will be evaluated for springs and seeps. Evaluations will consist of visually identifying intermittent wet areas (seeps), areas with flowing water (springs), or areas supporting increased plant growth when compared to surrounding areas (see section 18 of M3 2021 for additional detail).

Groundwater underdrains will be a series of parallel drains with branching laterals, instead of a single valley bottom drain, due to the broad u-shaped nature of the Meadow Creek valley. Pipes will transition from perforated (able to collect groundwater) to solid-wall (for conveyance only) as they exit their respective collection areas (impoundment and embankment) and flow underneath the buttress to the outlet. Underdrain flows will be collected in a sump downstream of the toe of the buttress, monitored for water quality, then either discharged to Meadow Creek surface water through a permitted Idaho Pollution Discharge Elimination System (IPDES) discharge, or pumped to the ore processing facility or a contact water pond for either treatment and discharge or use as makeup water for the mill process. The TSF liner system will then be installed in the TSF impoundment area over the underdrain system.

Underdrains will be installed beneath the TSF Buttress to ensure that groundwater does not saturate the base of that fill and potentially lead to water quality impacts or geotechnical

instability; however, little if any flow is expected in the buttress underdrains owing to lower observed groundwater levels beneath the buttress. Underdrain collection sumps and downgradient monitoring wells will be used for TSF leak detection.

# *1.7.8.1.1. TSF Liner System*

Due to water quality regulations and the presence of dissolved metals (chiefly arsenic and antimony, with trace mercury) and residual cyanide in the tailings pore water and supernatant pool, the TSF impoundment (including the upstream embankment face) will be composite lined with geosynthetic materials to prevent seepage of process water or transport of tailings out of the facility. A network of geosynthetic drains will be placed above portions of the geomembrane liner to reduce hydraulic head on the liner and excess pore pressure in the overlying tailings. The drains will report to a sump near the upstream embankment toe, and the water will be pumped out to the pool or reclaim system for reuse (M3 2021).

A composite liner consisting of a 60‐mil, single‐sided, textured, linear low-density polyethylene liner over a geosynthetic clay liner (GCL) will be employed to contain the tailings. Before placement of the liner within the TSF, the subgrade will be re-worked and compacted, or a minimum of 12 inches of buffer/liner bedding fill will be placed if re-working and compaction of native materials is not expected to meet subgrade design specifications as defined under IDAPA 50.01.13 (Rules for Ore Processing by Cyanidation). Geosynthetic overliner drains will be placed above portions of the liner to reduce hydraulic head on the liner and pore pressure in the overlying tailings solids during operations. The drains will direct water that migrates through the tailings to a sump near the upstream toe of the embankment, and the water will then be pumped out to the tailings pool within the impoundment or the reclaim system for reuse in the mill.

Facilities that use cyanide in their mineral extraction process are required to obtain a permit from the IDEQ and follow the Rules for Ore Processing by Cyanidation (IDAPA 50.01.13). The liner system proposed for the SGP meets the requirements of the rule under which the Project's Cyanidation permit is expected to be issued.

# *1.7.8.1.2. TSF Management Support Facilities*

Light vehicle roads and haul roads will provide access between the ore processing facility and the TSF, and the tailings delivery and reclaim water return pipelines will parallel the haul road. Secondary containment in the event of a pipeline break will consist of a geosynthetic wrap or an open geosynthetic lined trench. Further, the pipeline corridor will drain to one of two pipeline maintenance ponds – one at the truck shop and one at the ore processing facility. Electrically powered pumps will be located at the ore processing facility to pump tailings to the TSF and reclaim pumps will be located at the TSF to return water to the ore processing facility for reuse.

# *1.7.8.1.3. TSF Water Management*

Thickened tailings slurry will be pumped to the TSF (see section 18 of M3 [2021] for additional details). The TSF will be designed and operated as a closed-circuit, zero-discharge facility meaning no tailings water will be discharged during the mining operations phase to surface water or groundwater except in compliance with applicable permits and regulations. As the tailings consolidate, water collected in or falling on the surface of the TSF will form the supernatant pool on top of the tailings and be reclaimed for use in ore processing. Cyanide levels in the TSF reclaim water will be monitored throughout operations to ensure they remain in compliance with issued approvals and permits.

# **1.7.9. Mine Support Infrastructure**

SGP infrastructure to support surface mining and ore processing operations will include the following:

- A one-story mine administration building that will be sided or painted and roofed in neutral colors.
- A maintenance workshop which will store materials and supplies as discussed in Section 1.7.14, Materials, Supplies, Chemical Reagents, and Wastes.
- A truck wash facility which will include an oil/water separation system and water treatment facilities to enable recycling of the wash water.
- A worker housing facility (Figure 10), which will be constructed on NFS lands adjacent to Thunder Mountain Road (FR 50375) and will accommodate up to 500 people. This facility will include dormitories, food service, and recreation facilities, along with the supporting infrastructure of power, water supply, and wastewater treatment plant. The SGP main gate and security building will be collocated with the worker housing facility.
- Haul roads to transport ore, development rock, and reclamation materials from mining or storage areas, and to transport vehicles to the maintenance workshop. A typical haul road travelway will be approximately 87 ft. wide (81.1 ft. of running surface and 5 ft. of safety berm width). The haul roads will be built and maintained for year-round access and will be surfaced with gravel materials. Road maintenance activities will be conducted to manage fugitive dust emissions and maintain stormwater management features. The total disturbance associated with haul roads and other Project access roads based on the Reclamation Closure Plan is estimated to be 127.5 acres.
- Culverts will be installed where haul roads cross drainages or to direct stormwater to collection and retention structures. Culvert inlets and outlets will be lined with rock riprap, or equivalent, as needed to prevent erosion and protect water quality. Crossings of known fish-bearing streams will be constructed to support fish passage, with appropriately designed and constructed culverts or bridges.
- Service roads and paths that will provide an internal access system for employees and visitors to the site. The service roads will typically be 12 to 15 ft. wide; some will be graveled or covered with rock aggregate, while others will be two-track roads. There will be no planned public use of the SGP service roads or trails. The path system will enable SGP pedestrian traffic to move safely throughout the SGP operating area. Service roads and paths will be located within the overall disturbance area defined for the SGP and existing roads will be used to the extent possible.

Employee and visitor parking that will be maintained during construction and operations. During construction, the gravel parking areas will be located at the new worker housing facility, near the contractor/construction laydown areas, and at the Scout Portal. As operations are initiated, gravel parking areas will be maintained for buses, vans, and other miscellaneous vehicles for employees, contractors, vendors, and visitors at the new worker housing facility, at the shop area, and near the mine administration office.



**Figure 10. Worker Housing Facility, Stibnite Gold Project.**

# **1.7.10. Surface and Groundwater Management**

# **Water Use and Water Balance.**

The water balance is an accounting of inflows, outflows, and storage for various components of the mining and ore processing system. Actual volumes for water balance inputs and outputs could vary seasonally and annually from the volumes estimated. In particular, the seasonal basis for dust control is related to the time of year where the ground is not snow-covered or does not have enough ambient moisture present to control dust. This period is generally from June to October but can start earlier or extend later depending on precipitation conditions experienced each year. Precipitation events between June and October will also result in temporary periods during and immediately following the precipitation where dust control is not required due to the presence of ambient moisture. A water balance flow diagram for the mining and ore processing

operations phase is provided in BA Figure 3.5-5 (Stantec 2024) with components of the water balance described below.

# *1.7.10.1.1. Water Use and Supply*

Sources of water are required for ore processing, surface and underground exploration, dust control, and potable use. Water for industrial and mining uses will be supplied from water pumped from the dewatering wells located around the Hangar Flats, Yellow Pine, and West End Pits; industrial water supply wells; contact water storage ponds; a surface water supply intake on the EFSFSR; and process water recycled within the ore processing and tailings circuit. Dewatering production varies over the mine life from 100 gpm (0.22 cfs) to 2,200 gpm (4.9 cfs); industrial supply varies between zero and 1,300 gpm (2.9 cfs); contact water varies between zero and 1,600 gpm (3.6 cfs), and EFSFSR surface diversion varies between zero to 2,020 gpm (4.5 cfs). The surface water supply intake will be located immediately downstream of the debris screen before diverted flow enters the south portal of the EFSFSR tunnel. The intake will be equipped with a fish screen designed in accordance with the NMFS 2011 Anadromous Salmonid Facility Design guidance. Dedicated wells will provide potable water for worker consumption and sanitary use. Projected water use for the SGP is described in Table 12.

Component	Construction and Start-Up (gpm)	Operations (gpm)	Closure and Reclamation (gpm)
Underground and surface exploration	50	50	0
Surface dust control (seasonal basis)	33	66	16.5
Ore processing including tailings storage	$\Omega$	3.900	$\Omega$
Potable or domestic use	26	12	4
Sub-Total Use	109	4,028	20.5
Contingency (10%)	11	403	↑
<b>Total Estimated Use</b>	120	4,431	22.5

**Table 12. Estimated Gross Fresh and Recycled Water Usage.**

**Source:** Perpetua 2021a

gpm = gallons per minute

 $\overline{a}$ 

As shown in Table 12, ore processing facility operations will represent approximately 97 percent of water use associated with the SGP. A separate wellfield of up to four wells will be developed in the EFSFSR drainage adjacent to the worker housing facility to provide potable water for the housing facility. The use of water from pit dewatering, contact water from precipitation runoff, surface water, and development of separate wellfields for supplemental industrial water and potable water at the worker housing facility will require permitting through the IDWR as new water rights or transfer of the place of use for one of Perpetua's existing water rights. Perpetua has submitted an application to IDWR for a total diversion of up to 9.84 cfs (4,416 gpm) for use by the SGP<sup>3</sup>. A description of water rights authorizing diversion of water for use at the SGP main facility, and at the Landmark Maintenance facility, is in Section 1.7.10.1.9.

<sup>&</sup>lt;sup>3</sup> As of July 5, 2024, IDWR had approved water rights with a total maximum diversion rate of 9.8 cfs for use at the SGP mine site, and an additional 0.04 cfs for use at the Landmark maintenance facility.

### *1.7.10.1.2. Water for Ore Processing.*

Ore processing is the primary driver for water use. Process water will require a continuous supply with approximately 80 percent of process use reclaimed from the TSF (i.e., approximately 3,000 gpm [6.7 cfs]). Water sources for ore processing include water from pit dewatering and water supply wells, contact water, EFSFSR surface water intake, and water recycled from the TSF. Outflows from ore processing include tailings slurry conveyed to the TSF and evaporative losses from various process components.

The majority of the water needed for ore processing will be recycled (reclaimed) from the TSF. Reclaim water will be pumped from the supernatant water pool at the TSF to the reclaim water tank at the ore processing facility. Makeup water will be supplied from pit dewatering in wells located around the Hangar Flats, Yellow Pine, and West End Pits; water supply wells; contact water; and surface water intake in the EFSFSR. Water will be pumped from the pit dewatering wells to freshwater tanks near the ore processing facility site. These tank facilities also could supply water for exploration drilling, development drilling, in-pit road dust control, and emergency fire suppression. The freshwater tanks could store approximately 360,000 gallons of water; 240,000 will be available for process uses, and the remaining 120,000 gallons will be maintained for fire suppression. Additional water needed for ore processing, domestic uses, etc. will be diverted from the EFSFSR and from groundwater wells, in accordance with State of Idaho water rights (see Section 1.7.10.1.9).

### *1.7.10.1.3. Water at the TSF*

Inflows to the TSF include tailings slurry and precipitation. The TSF will store tailings solids, water entrained with the tailings, and free water atop the tailings (supernatant pool). Stormwater and snow falling directly on the TSF and water from the supernatant pool, that forms as the tailings consolidate, will be stored in the TSF and will be reclaimed for ore processing. Water infiltrating to the base of the TSF will be captured by the liner overdrains, enter a sump, and be pumped back to the supernatant pond. The volume of available reclaim water will be influenced by the ore processing volumes, precipitation, and evaporation. The reclaim water will be pumped from the TSF to the reclaim water tank located at the ore processing facility. During periods of site-wide water excess, reclaim can be curtailed and contact water could be used directly in ore processing to facilitate emptying the contact water ponds, while retaining water in the TSF for use in an upcoming dry season. Periods of site-wide excess water coincide with the periods of greatest mine dewatering in MYs 4, 5, and 6. There is a potential for excess water in other years (MYs -1 through 12) if there is greater than average precipitation events contributing to more contact water collection. Local stormwater and snowmelt runoff from outside the TSF footprint and the existing Meadow Creek will be routed around the TSF.

# *1.7.10.1.4. Water for Potable Use*

Potable water will be needed for worker consumption and sanitary use. Groundwater will be the primary source of water for potable use at the SGP. An existing well located near the exploration camp in the EFSFSR drainage will be used to supply an independent water circuit, along with a separate wellfield in the EFSFSR drainage adjacent to the worker housing facility. Wells also

will be drilled for potable and industrial or commercial water uses at the Burntlog Maintenance Facility and the SGLF. See Section 1.7.10.1.9 for information on water rights associated with the proposed action.

# *1.7.10.1.5. Water Treatment*

The Project's water treatment requirements, objectives, and methods are described in detail in the Stibnite Gold Project Water Management Plan (Brown and Caldwell 2021b) and summarized here. Three water types will require treatment over the life of the SGP: (1) contact water from mine facilities, which includes dewatering water (construction through closure); (2) process water from the TSF (closure); and (3) sanitary wastewater (construction through early closure). During operations, treating and releasing contact water will generally be limited to periods when a significant amount of dewatering water is being produced, or seasonally in wet years (i.e., Spring runoffs). Outside of that time, much of the collected contact water could be put to beneficial use in the mill. During construction and at closure, absent a water demand for ore processing, less contact water will be consumed and proportionally more will be disposed of through evaporation or treatment and discharge. From construction through early closure, the camp and offices will produce sanitary wastewater needing treatment. Additional water treatment that could be required during post-closure is discussed in Section 1.9.13, Post-Closure Water Treatment. Permit discharge limits will be developed according to IDEQ and CWA requirements and the limits will be established by the IPDES permit issued by the IDEQ.

The sources proposed for operational water treatment by Perpetua include:

- Contact water from dewatering of the Yellow Pine, Hangar Flats, and West End Pits;
- Stormwater runoff (including snowmelt) from the pits, TSF Buttress, Bradley tailings, SODA, Hecla heap leach, run-of-mine ore stockpile area, truck shop, and ore processing facility;
- Toe seepage from the TSF Buttress and long-term ore stockpiles;
- Groundwater produced by the dewatering system; and,
- Sanitary wastewater from the worker housing facility, truck shop, ore processing facility, and administrative buildings.

The conceptual water treatment system during operations will need to adhere to surface water quality standards for regulated constituents, most notably arsenic and antimony. The discharge quality will meet IDEQ standards for all regulated constituents. The discharge rate will be between zero and 2,000 gpm (4.5 cfs) at a primary location at the process plant (near the confluence of Meadow Creek and EFSFSR) and a secondary location on Meadow Creek east of the HFP area. This water will be used to supplement flows in Meadow Creek if needed. The outfalls will be installed to minimize sedimentation and maintain total suspended solid (TSS) within IDEQ accepted limits. Water treatment discharge is predicted to be 2.5<sup>o</sup>C higher than ambient flow in Meadow Creek. This effect will be offset through the use of diversion pipes around the TSF which maintain cooler water from upstream, resulting in stream flow 2.3°C cooler in the EFSFSR below the treatment plant outfalls. Thus, coupled with the timing of water treatment needs with respect to the mining sequence and dewatering excess, treatment methods and capacity will be phased. During construction and early in operations, a modular, mobile,

two-stage iron coprecipitation system will be utilized. Early in operations, this system will be replaced by a two-stage iron coprecipitation system located near the ore processing facility. Residuals (sludge) from the water treatment during construction will be stored in a small impoundment in the TSF footprint. During operations and closure, the residuals will be stored in the TSF. Due to contact water runoff seasonality, reuse, and equalization storage (i.e., ponds), average treatment rates are often significantly less than nominal treatment capacity, except during the HFP dewatering when a substantial proportion of treated water will be from relatively constant dewatering flows.

Dewatering flows will be met with a staged water treatment strategy. The construction time period is paired with 300 gpm (0.7 cfs) of peak capacity from package iron coprecipitation systems. The first three years of operations will require 1,000 gpm (2.2 cfs) of total treatment capacity, using an iron coprecipitation system that will remain until closure. During peak simultaneous dewatering of the YPP and the HFP, an additional 1,000 gpm (2.2 cfs) of modular water treatment capacity will be brought online for approximately three years, then treatment capacity will be scaled back to 1,000 gpm (2.2 cfs) for the remainder of operations and early closure.

Prior to closure, a new closure water treatment plant will be constructed to accommodate treatment of water from the TSF, which will include iron coprecipitation and the application of reverse osmosis membrane treatment. After mine closure and final reclamation of the TSF Buttress and pit backfill surfaces, contact water treatment will no longer be required because installation of a geosynthetic liner, growth material cover, and revegetation will preclude contact of surface runoff with mined materials; but process water treatment for the TSF (Section 1.9.13) will continue longer, through approximately year 40. The closure treatment plant will be located on private land at the TSF Buttress as the TSF will ultimately be the only remaining water source requiring treatment.

Enhanced evaporation, using snowmaker style misters located over the lined TSF, collection ponds, and/or pits, will supplement the treatment system, in particular to prevent surplus process water accumulation in the TSF and eliminate contact water inventory, if necessary, when environmental conditions are conducive to evaporation. Predicted dewatering rates were combined with estimated volumes of mine-impacted waters from the Site-Wide Water Balance (Perpetua 2021d) to forecast the volume requirements for water treatment during operations and closure. Water treatment is required whenever the volume of produced groundwater plus mineimpacted waters exceed the consumptive use demands for the Project. Hence, the water treatment volume estimate represents the sum of predicted mine-impacted water values (e.g., dewatering production, contact water) less the consumptive use by the Project (i.e., process water). These volumes ranged from 2,000 gpm (4.5 cfs) during the years of highest dewatering production down to zero flow from the collection of mine-impacted waters post-closure. Estimates also included potential variability associated with meteoric conditions on the generation of contact water to develop potential contact water volumes associated with the range between the  $5<sup>th</sup>$  and  $95<sup>th</sup>$  percentiles of predicted volumes. Contact water storage ponds will be used to provide temporary storage of contact water flows. The location of these storage ponds is constrained by topography, other proposed mine facilities, legacy materials, and near-surface groundwater levels. The ponds are also located to manage runoff in proximity to the watergenerating areas (Table 13). The Project water management system is designed with storage capacity for meteoric water events so that water destined for treatment can be contained until it can be transferred to the water treatment plant (WTP) for constituent removal at the plant's 2,000 gpm (4.5 cfs) design rate (Table 14). Contact water ponds will be geomembrane-lined earthen facilities, equipped with emergency spillways and designed to contain runoff volumes associated with design storm runoff events (Table 15).

<b>Pond Name</b>	<b>Location</b>	<b>Duration</b> in Mine <b>Years</b>	<b>Facilities Served</b>	<b>Pumped Inflow</b> <b>Source</b>
<b>Hangar Flats</b> Pond	In footprint of Hangar Flats Pit (HFP)	$-2$ to 4	TSF Embankment and Buttress, HFP, SODA	Gravity inflow
Soda Pond	East of TSF Buttress in footprint of SODA/Bradley tailings	3 to 17	<b>TSF Buttress, SODA</b>	Gravity inflow
West End Pond	Downstream and north of West End Pit (WEP) in the West End Creek drainage	$-1$ to 9	<b>WEP</b>	WEP sumps
Midnight Pond	Upstream and south of the Yellow Pine Pit (YPP) near the confluence of Midnight Creek and the EFSFSR	$-2$ to 15	WEP, YPP	WEP sumps, YPP sumps
North Truck Shop Pond	In Meadow Creek valley in the footprint of the truck shop area	$-2$ to 17	<b>Truck Shop</b>	Gravity inflow
South Truck Shop Pond	In Meadow Creek valley in the footprint of the truck shop area	$-2$ to 17	Truck Shop, HFP	HFP sumps
North Plant Pond	North of Garnet Creek on the northern side of the process plant site	$-2$ to 17	Process Plant site, ore stockpile	Gravity inflow
<b>Central Plant</b> Pond	North of Garnet Creek in the $-2$ to 17 central portion of the process plant site		Process Plant site	Midnight Pond, South Truck Shop Pond, North Truck Shop Pond, Hangar <b>Flats Pond</b>
<b>Scout Pond</b>	North of Garnet Creek on the eastern side of the process plant site		Scout stockpile	Gravity inflow

**Table 13. SGP Contact Water Pond Locations.**

Pond	<b>Pond Capacity</b> (excluding) freeboard; acre-feet)	<b>Design Storm</b> <b>Runoff</b> (acre-feet)	Freeboard (feet)	<b>Embankment</b> <b>Height</b> (feet)
Hangar Flats Pond	201.8	33.9	3	35.0
Soda Pond	147.7	24.6	3	29.4
West End Pond	28.7	39.3 <sup>1</sup>	3	60.5
Midnight Pond	83.9	16.8	3	72.7
North Truck Shop Pond	3.2	3.2	2	N/A
South Truck Shop Pond	18.3	17.9	2	N/A
North Plant Pond	7.5	7.3	$\mathfrak{D}$	N/A
<b>Central Plant Pond</b>	4.3	4.3	2	N/A
<b>Scout Pond</b>	9.0	9.0	$\overline{c}$	N/A

**Table 14. SGP Contact Water Ponds Design Summaries.**

Source: Brown and Caldwell 2021b, Table 6-2.

<sup>1</sup>West End Pond can contain the 100-year, 24-hour storm volume (25.8 acre-feet). Additional potential volume from snowmelt will be managed using in-pit sumps or pumping stored water from West End Pond to Midnight Pond or YPP.

 $n/a$  = not applicable – ponds excavated into the sub-grade

The installation of geosynthetic liner systems on the top surface of the TSF, TSF Buttress, YPP backfill, and Hangar Flats backfill inhibits the generation of contact water in the post-closure period plus drainage of the water entrained in the TSF results in the abatement of contact water flows after approximately 40 years.

# *1.7.10.1.6. Contact Water Pond Chemistry*

During operations, contact water from SGP facilities, and occasionally pit dewatering water, will be directed to site contact water collection ponds and subsequently directed to the WTP. Inflow sources to each collection pond, and predicted analytes of concern for each contact water pond, are provided in Table 3.5-6 of the BA (Stantec 2024). Open pit dewatering water that is not directed to site contact water collection ponds will be pumped directly to the WTP. The predicted quality of the treatment plant inflow is summarized in Table 3.5-7 of the BA.

# *1.7.10.1.7. TSF Embankment and Buttress*

During the construction and early operations phases, Hangar Flats Pond will be located near the northeast toe of the TSF Buttress to provide contact water storage. Runoff and toe seepage from the TSF Buttress and remaining legacy materials in SODA will be conveyed to the Hangar Flats Pond using a series of runoff collection channels or berms, internal collections sumps, pumps, and pipelines as needed. The SODA Pond will be constructed south of the TSF Buttress to provide contact water storage for the remaining years of operations and closure, as the Hangar Flats Pond will be deconstructed as the HFP is mined below the valley bottom. Details regarding the TSF Buttress design can be found in Perpetua 2021a. A summary of the information follows.

At final buildout, the TSF Buttress and adjacent TSF Embankment will contain 142 million tons of material, comprising 85.5 million tons (60 percent) of non-potentially acid generating (PAG) development rock from the YPP, 22 million tons (16 percent) of non-PAG development rock

from the WEP, 14.3 million tons (10 percent) of non-PAG development rock from the HFP, 6.4 million tons (4 percent) of PAG development rock, 11.7 million tons (8 percent) of borrow material, 1.25 million tons (0.9 percent) of spent ore from the Hecla Heap, 0.85 million tons (0.6 percent) of spent ore from the SODA, and 0.2 million tons (0.1 percent) mine waste placed on the former Stibnite Mine, Inc. on/off leach pads during the Stibnite Administrative Settlement and Order on Consent (ASAOC) action. Active 'blending' of the development rock during operations is not proposed. During operations, ore stockpiles 1, 2, 3 and 4 will be located on top of the TSF Buttress and are assumed to contribute to solute loading from the facility during the operational period only. These stockpiles are assumed to have been completely removed and processed prior to closure.

At closure, the TSF Embankment and Buttress will be regraded to promote positive drainage and a low permeability geosynthetic cover will be placed over the entire facility, which will be designed to limit infiltration through the underlying development rock (Perpetua 2021a). The geosynthetic cover will be overlain by an inert soil/rock layer and GM and revegetated. Both the runoff and the toe seepage from the TSF Embankment and Buttress report to a contact water pond and then to the WTP.

The mine-affected waters that report to the ground surface will be subject to consumptive use in ore processing with any water production above consumptive use subject to water treatment and discharge. To summarize, these mine-affected waters that will be subject to water treatment include: dewatering production, waters collected in contact water ponds, stockpile runoff and toe seepage, TSF Buttress runoff and toe seepage, and post-closure TSF facility solutions.

Waters infiltrating into the subsurface under the mine facilities will mix with alluvial groundwater and are not subject to water treatment except in instances where alluvial groundwater is subsequently pumped for mine dewatering.

The Site-wide Water Balance model (Perpetua 2021d) provides a forecast for the volumes of water that will require water treatment for the operating and post closure time-periods. A principal driver for predicting water treatment rates will be uncertainty in future precipitation rates and their effect on contact water. A 120-year precipitation record was utilized to develop percentile estimates for meteoric inputs to the water balance. Initially, the volumes of water destined for water treatment will be less than 500 gpm (1.1 cfs) because dewatering and seepage rates from newly constructed facilities will be ramping up at the same time that consumptive use demand for processing needs will be at its largest and consuming contact water as a supply. Over time, water treatment volumes will increase through about MY 6 to approximately 2,000 gpm (4.5 cfs) as dewatering production and seepage rates will constitute a higher percentage of diversion for process water in those years, displacing contact water as a source. Differences in actual versus predicted dewatering rates will have limited effect on water treatment needs because diversion from industrial supply wells or surface waters will be reduced to offset any increase dewatering production (USFS 2023a). Following MY 6, predicted dewatering rates will decline removing most of the need for water treatment as water recycling will be needed to meet consumptive use demands, except during seasonal runoff periods when contact water volumes will increase. Any short-term volumes in excess of the water treatment capacity (i.e., following a large storm event) will result in water storage within the TSF and/or contact water ponds.

In the closure and post-closure periods, beginning in MY 15, volume of mine-affected waters requiring water treatment will range seasonally up to approximately 1,000 gpm (2.2 cfs) until geosynthetic cover installations (planned to commence in MY 19) could be completed in MY 23 to prevent mixing of surface water runoff and contact waters with consolidation water. Once the cover installations are in effect, volumes consisting of residual seepage and TSF consolidation water will continue to be treated but will decrease from approximately 200 gpm (0.4 cfs) down to very minor, unmeasurable flow as the tailings solids consolidate and stop emitting water. To meet applicable discharge standards, the target post-treatment concentrations for analytes were identified for the water treatment plant design (Table 15).

<b>Parameter</b>	<b>Units</b>	<b>Treatment Objective</b> <sup>1</sup>
pH (range)	s.u.	$6.9 - 9.0$
Silver	mg/L	0.0007
Arsenic	mg/L	0.01
Cadmium	mg/L	0.00033
Chromium (III)	mg/L	0.035
Chromium (IV)	$mg/L$	0.0106
Copper	mg/L	0.0025
Mercury	mg/L	0.000012
Nickel	mg/L	0.024
Lead	mg/L	0.0009
Antimony	mg/L	0.0052
Sulfate	mg/L	250
Thallium	mg/L	0.005
Zinc	mg/L	0.054
Nitrate/Nitrite	mg/L as N	10
Ammonia	mg/L as N	2.1
Cyanide, Total	mg/L	0.0052
Cyanide, WAD	mg/L	0.0039
<b>Total Dissolved Solids (TDS)</b>	mg/L	500

**Table 15. Target Post-Water Treatment Plant Effluent Analyte Concentrations.**

**Source:** Brown and Caldwell 2021b

<sup>1</sup>Treatment objectives are equivalent to the strictest potentially applied water quality standard.

Brown and Caldwell (2021b) performed an assessment of the viability of potentially applicable water treatment technologies to the predicted maximum influent water chemistry and identified the following technologies, described below, to incorporate into the Project design for the construction, operational, and post-closure periods.

Temporary treatment systems will be employed during the construction period until the Project's WTP is constructed and commissioned. These temporary systems will utilize trailer-mounted or skid-mounted equipment packages containing membrane treatment and/or iron coprecipitation systems that can be set up with limited lead time. Figure 3.5-9 in the BA (Stantec 2024) illustrates the construction period water treatment flowsheet.

Figure 3.5-10 in the BA (Stantec 2024) illustrates the operational period water treatment plan flowsheet with a design capacity of 2,000 gpm (4.5 cfs). For the operational period water chemistry, a treatment process consisting of sodium hypochlorite oxidation, two-stage iron coprecipitation with ferric sulfate, and solids separation with contingent mercury precipitation via organic sulfide precipitant addition between iron precipitation stages was selected. Influent waters will be stored in lined storage ponds for flow equalization and pumped into the WTP. This operational water treatment generally targets dissolved nitrate, metals, and oxyanions in influent solution, primarily arsenic and antimony. Addition of the mercury-sequestering precipitant is included as a contingency for the design to account for uncertainties regarding the effectiveness of iron coprecipitation in reducing dissolved mercury and methylmercury concentrations to levels below applicable receiving stream standards. Residual solids from the treatment plant will be placed in the TSF.

Under an IPDES permit, the water treatment plant effluent will be directed to Meadow Creek at a location upstream of the HFP when flow augmentation is required (i.e., when Hangar Flats groundwater pumping results in decreased Meadow Creek baseflow) and otherwise to the EFSFSR for the remainder of operations.

For the post-closure period, the water treatment process will need to be augmented to treat cyanide, sulfate, and total dissolved solids (TDS) concentrations that will be derived from the remaining inventory of TSF process water and tailings consolidation seepage (BA Figure 3.5-11, Stantec 2024). The treatment process begins with chemical oxidation followed by iron coprecipitation to remove a significant fraction of dissolved metals. Organic sulfide precipitation of mercury will be provided. Softening will be performed via lime and soda ash to remove calcium and magnesium. Adjustment of pH will be provided in advance of ultrafiltration to remove carryover solids from the solids contact clarifier and prevent particulate fouling of the reverse osmosis (RO) membranes. The RO membrane treatment will separate the dissolved solids into a concentrated brine while the permeate water will be pH adjusted and re-mineralized prior to discharge to Meadow Creek via an IPDES-permitted outfall. Treatment plant residual solids will be placed in the TSF until its cover was completed, and thereafter dewatered and disposed of in a location constructed in the TSF above the cover.

The operations phase water treatment plant will treat mine-impacted water and discharge to the EFSFSR (or Meadow Creek if flow augmentation is necessary) through reclamation of operational components through MY 18. Prior to MY 15, the reclamation and closure phase WTP will be constructed on top of the TSF Buttress where it will treat mine-impacted water to Meadow Creek through the completion of water treatment requirements estimated to be in MY 40.

# *1.7.10.1.8. Sanitary Wastewater Treatment*

The worker housing, administration building, warehouse, maintenance shops, and underground exploration surface facilities will produce sanitary wastewater. Wastewater from the administration building, warehouse, maintenance shops, and underground facility will be collected in tanks for transport to a sanitary wastewater treatment plant equipped with a septage receiving system located near the worker housing facility. The sanitary wastewater treatment

plant will consist of a package plant containing a membrane bioreactor or equivalent system to treat wastewater to applicable discharge permit requirements. The volume of wastewater influent will depend on the number of personnel working on site and is expected to be approximately 50,000 gallons per day (gpd) (0.15 acre-feet/day [afd]) during the construction period and 25,000 gpd (0.08 afd) during operations (Brown and Caldwell 2021b).

Sanitary wastewater treatment plant effluent will be discharged to the EFSFSR at an IPDES permitted location near the worker housing facility. Treatment residuals will be dewatered and transported to a permitted, off-site landfill for disposal.

# *1.7.10.1.9. State of Idaho Permits and Cyanidation Permit*

The State of Idaho has regulatory authority over its IPDES process. The SGP will need permits issued by the IDEQ to discharge treated water from the WTP and the sanitary wastewater treatment plant. Under the IPDES program, IDEQ will establish specific discharge limits for constituents of interest plus monitoring and reporting requirements for the system based on its regulatory criteria.

The SGP will also need a Cyanidation Permit issued by IDEQ to allow the use of cyanide in its ore processing. Under this permit, IDEQ will institute permit obligations regarding the handling and containment of process solutions as well as responses to upset conditions. In addition, the permit will also contain requirements for the ultimate treatment and disposal of process water. The descriptions of handling TSF water in this report are consistent with the requirements of the Cyanidation Permit regulations.

The IDWR regulates mine tailings impoundments with dams higher than 30 ft. and administers regulations that may have to be considered when a tailings impoundment affects surface water hydrology. The IDWR also is responsible for administration of water rights, well construction standards, dam safety, and stream channel alteration. All water rights to implement the SGP will need to be granted to the applicant by the State of Idaho through IDWR.

In addition to the reclaimed water, described in Section 1.7.10.1.2, water will also be diverted from the EFSFSR and from groundwater wells. The single surface water diversion will be used to divert up to 4.5 cfs to serve three water rights (Water Right Nos. 77-7122, 77-7293, 77-14378) and will meet NMFS guidelines for screening and for upstream and downstream fish passage. All diversion of water within the EFSFSR drainage will be done in accordance with the water rights in Table 16. Although some of these water rights are in the permitting process, that process has been ongoing for several years and NMFS does not expect additional changes prior to project implementation.

#### **Table 16. Water Rights Authorizing Water Diversion Needed to Support Mining Activities, Ore Processing, Domestic Uses Associated with the Proposed Action, etc.**



1. Surface water diversion from the EFSFSR is limited to 4.5 cfs; diversions are conditional on flows in the EFSFSR, at the POD, of at least 5.0 cfs from October 1 – June 29 and 7.25 cfs from June 20 – September 30; diversions are conditional on flows of at least 3.0 cfs in lower Meadow Creek; and diversions cannot reduce flows in the EFSFSR by more than 20% when flows are less than 25 cfs downstream from Sugar Creek.

2. This water right is for the Landmark maintenance facility in the Johnson Creek drainage, which is outside of the flow analysis area.

Source: Information was obtained from the IDWR water rights search engine

(https://research.idwr.idaho.gov/apps/waterrights/wrajsearch/wradjsearch.aspx)

Key:  $cfs = cubic$  feet per second;  $AF = Acc$  Feet

In addition to the conditions described in the footnotes for Table 16, water diverted for use at the SGP must comply with additional conditions that are designed to protect instream flows for the Wild and Scenic River (WSR) reaches of the Salmon River, described in water rights 75-13316 and 77-11941. These conditions include mitigations that were designed by Perpetua Resources and proposed during development of the proposed action. Details of the water rights described in Table 16 are available at:

[https://research.idwr.idaho.gov/apps/waterrights/wrajsearch/wradjsearch.aspx.](https://research.idwr.idaho.gov/apps/waterrights/wrajsearch/wradjsearch.aspx)

The two USFS WSR water rights on the Salmon River are measured at the Shoup gage, which is upstream of the MFSR confluence. These are instream, non-consumptive water rights that maintain flows for the WSR designated segment of the Salmon River. When flows measured at the Shoup gage are less than 13,600 cfs, the minimum instream flow rates provided by the water rights range from 1,200 cfs for the period of September 1 to September 15 to 9,450 cfs for the period of June 1 to June 15. The SFSR joins this WSR segment of the mainstem Salmon River approximately 64.6 miles downstream from the SGP area. The BA explains that these water rights are subordinated to all water rights claims filed in the Snake River Basin Adjudication as

of the effective date (September 1, 2003) of the Stipulation among the U.S., the State of Idaho, and other objectors. They also are subordinated to specified quantities of future beneficial use rights. Additional detailed information regarding these two water rights can be found in water right reports (referenced by water right number) available on the IDWR website [\(https://idwr.idaho.gov/water-rights/\)](https://idwr.idaho.gov/water-rights/).

The surface water diversion, screen, and bypass facility will be constructed prior to operation of the SGP. Groundwater pumping will be via existing wells and via new wells that may be drilled before or during operation of the SGP. This includes a new well at the Landmark Maintenance Facility in the Burntlog Creek drainage.

The BA explains that no water right with a junior priority date can deplete the water needed to maintain the Idaho Water Resource Board (IWRB) maintained minimum streamflow water right on the EFSFSR (Water Right 77-14190), unless allowed as a condition of approval of the proposed junior water right. All the existing water rights at the SGP predate the priority date of April 1, 2005, associated with water right 77-14190. Any new water rights permits will have a junior priority date, but the minimum stream right (77-14190) on the EFSFSR is subordinate to all future domestic, commercial, municipal, and industrial uses, and up to 8.2 cfs of new nondomestic, commercial, municipal, and industrial uses. This will allow authorization of up to 8.2 cfs of new non-domestic, commercial, municipal and industrial water rights to which water right 77-14190 will be subordinate.

# *1.7.10.1.10. Drainage Area Alteration*

The WEP Lake will be excavated in the Sugar Creek drainage, the catchment area of the Lake will be 185 acres, and the fill volume will be 7,027 acre-feet. In a median flow year, approximately 0.39 cfs runs off of the 185 acres that will become the WEP Lake catchment area via Sugar Creek. After MY 1, that water instead flow into the WEP, until the lake fills.

# **Surface Water Management**

To manage surface water at the SGP, existing streams that run through areas proposed for mining related disturbance will be diverted. Temporary diversions will be used within the SGP to keep non-contact water separated from contact water. Contact water is water that flows into or through disturbed areas and mining facilities and could have the potential to pick up increased levels of sediment, metals, and other possible contaminants which cannot be discharged into surface water and groundwater without proper treatment. Non-contact water is meteoric water that does not contact disturbed areas or mining facilities.

# *1.7.10.2.1. Stream Diversions Around Mining Features*

Existing streams will be temporarily diverted around SGP facilities, within constructed surface water channels. Diversion channel segments constructed in erodible materials will be lined with riprap to prevent erosion. Rock-cut channels will be constructed on steep slopes and in areas with shallow or at-surface bedrock, will have low erosion potential, and not require riprap lining. Certain channel segments constructed over fill or excavated in permeable materials will be lined with a geosynthetic liner to prevent seepage. A geotextile and/or transition layer of sand/gravel

followed by riprap will be placed over the liner for erosion protection. Certain diversion sections will be piped as dictated by terrain or the need to limit warming of water. Diversions will be sized for high flows in diverted creeks, i.e., approximately 7.4 cfs for Hennesey Creek, 1.4 cfs for West End Creek, and 700 cfs for the EFSFSR tunnel. The underground diversion for Hennesey Creek will be 18- to 24-in. in diameter while the underground diversion for West End Creek will be 8 to 12-in. in diameter. The dimensions of the EFSFSR tunnel will be 15 ft. high by 15 ft. wide. These diversions will be in place through the mine operations period until replaced by the restored stream channels during the reclamation and closure period.

During mine operations, summer low flows in perennial diversion channels around the TSF impoundment and buttress (Meadow Creek), YPP (Hennessy Creek), and WEP (West End Creek) will be piped underground as an EDF to maintain cold stream temperatures. Eight- to 12 in. diameter pipes, sized to convey August baseflow, will be installed under the diversion channels in the riprap channel lining or under the adjacent access road to carry low flows. Stream flow will enter pipes through inlets at the same locations stream and tributary inflows will be diverted into the constructed channel. These combination diversions (combination of underground piping below diversion ditches) will be operated such that the pipe will contain flow year-round and surface flow will be intermittent. That is, surface flow will occur when the diverted flow is greater than the capacity of the pipeline (Richard Rymerson, personal communication, May 14, 2024). Some diversions, such as portions of Hennessy and West End Creeks, and the EFSFSR tunnel, will be entirely underground, in which case conduits will be larger and sized for high flows.

Streams will be routed into the diversion channels by constructing a temporary flow barrier, such as a diversion berm or cofferdam, to redirect flows from the existing streams into the diversion channels. Additional protection, such as riprap or energy dissipation structures, may be needed at the channel entrances and exits to ensure velocities do not scour the existing streambed or bank. Where needed, trash racks or similar debris removal structures will be installed at the channel entrances to prevent large wood and other debris from entering the diversions.

To help ensure the stream diversions are completed in a manner protective of fish inhabiting the streams, Perpetua has developed a plan for isolating channel segments, dewatering, and salvaging and relocating fish during dewatering or maintenance of natural stream channels and diversion channels, as described in the FMP (Brown and Caldwell, Rio ASE, and BioAnalysts 2021). Stream segments to be dewatered may be isolated using a variety of methods as appropriate for the site conditions. Fish salvage methods are described in Appendix A, Table A-1. Sediment controls will be installed, where needed, to reduce stream turbidity and prevent sedimentation of the downstream receiving streams.

Fish handling and salvage operations will be required for the following SGP activities:

- Construction of the TSF/DRSF;
- HFP mining;
- Stream restoration and enhancement activities:
- EFSRSR Tunnel construction and dewatering of the YPP;
- EFSFSR Tunnel maintenance: and,
- Culvert or bridge construction or replacement.

All temporary dewatering and diversion efforts for activities, such as stream repair, culvert maintenance, or temporary stream impacts from other mining activities, will have the proper fish exclusion screening, or other method, to minimize the risk of fish becoming entrained in the pump and/or diversion. Stream diversions around the TSF, TSF Buttress, YPP, Process Plant site, and Fiddle GMS will be assessed on a case-by-case basis for whether fish exclusion is necessary based on the diversion structure, channel dimensions, and likely fish presence. Further details on these potential exclusions are provided in the FMP and are incorporated here by reference (Brown and Caldwell, Rio ASE, and BioAnalysts 2021).

### *1.7.10.2.2. EFSFSR Temporary Diversion Tunnel*

Currently, the EFSFSR flows into and through the YPP Lake. The cascade at the inflow to the pit lake currently blocks upstream fish passage. A tunnel will be built in MY -1 to direct the EFSFSR around the west side of YPP to allow mining in the pit and fish passage during construction and operations (Figure 11). The tunnel will be approximately 0.9-mile-long and 15 ft. high by 15 ft. wide. The tunnel will include a fishway stream channel designed to provide for upstream and downstream passage of migratory and anadromous salmonid fish.

The tunnel has been designed so that fish could swim through its entire length in both directions (Brown and Caldwell, McMillen Jacobs and BioAnalysts 2021). To encourage fish passage, lowenergy lighting will be installed in the tunnel and set on timers to simulate daylight. A trash rack will be constructed near the upstream entrance to the tunnel to prevent large wood, boulders, and other debris from entering the tunnel, and will be periodically cleaned. The spaces between the trash rack bars will be sized to allow adult Chinook salmon passage. A surface water supply intake with fish screens will be installed upstream of the trash rack at a control weir to divert water from the EFSFSR for ore processing makeup when necessary.

A parallel roadway will be constructed in the tunnel to allow equipment and personnel access for monitoring, inspection, and maintenance. The accessway will function as a floodway for high flows, greater than the normal flow range within the fishway.

The tunnel fishway will incorporate concrete weirs, designed to produce hydraulic conditions that could be successfully navigated by fish (McMillen Jacobs 2018). The south portal (upstream end) of the tunnel will include a sediment collection and drop out area, a resting pool, trash rack, flow control weir, and picket panels. The north portal, located at the downstream end of the tunnel, will include an orientation pool for downstream migrating juvenile fish with an adult exclusion barrier to reduce potential predation, a separate adult fish holding/resting pool, rock weirs and a transition zone. A barrier to prevent upstream movement by fish would be established in the EFSFSR upstream of the north portal, using a picket type weir with a fish trap.



**Figure 11. Cutaway View of Fish Passage Tunnel.**

Activation of the EFSFSR tunnel is described in detail in the FOMP (Brown and Caldwell, McMillen Jacobs, and BioAnalysts 2021), and summarized here. Activation would occur during an approved in-water work window and would follow the following sequence:

- 1. Pre-wash the fishway and accessway of the EFSFSR tunnel before watering up the tunnel. Water would be diverted from the stream into the fishway and accessway using a screened pump or seined opening in the cofferdam. Typical 6-inch trash pumps operate at a maximum pumping capacity of approximately 2 cfs, but pre-wash flow should be regulated to less than ¼ of the total flow upstream of the water switch location. This would fill the fishway until water reaches the downstream end. Turbid wash water would be detained and pumped to a predetermined upland location rather than discharged directly to fish bearing waters. Water would be detained and pumped from the adult holding pool.
- 2. Prepare EFSFSR tunnel for water by installing seine (or other appropriate fish barrier) at downstream and upstream ends to prevent fish from moving into the EFSFSR tunnel until 2/3 of the total streamflow is available in the EFSFSR tunnel (after step #4). Starting early in the morning, introduce 1/3 of the flow into the EFSFSR tunnel fishway over a period of 2–4 hours. There is approximately 3,134 feet of existing EFSFSR channel (not including the YPP) lying between the EFSFSR tunnel entrance and exit, with water flowing at an estimated velocity of 2 feet per second (fps); it would take a minimum of approximately 30 minutes for the reduction in flow (as flow introduced into the EFSFSR tunnel) to be noticed at the downstream end of the existing EFSFSR channel. Two to four hours would provide sufficient time for the existing channel to slowly reduce flows and allow some natural outmigration of fish from the reach into the YPP.
- 3. Monitor turbidity in north portal adult holding pool in accordance with state and federal regulations. For example, Idaho water quality standards state that the activity needs to be modified to reduce turbidity if instantaneous turbidity exceeds 50 nephelometric turbidity units [NTU]) above background or exceeds more than 25 NTU (above background) for more than 10 consecutive days (IDEQ 2017a). Continue monitoring until turbidity is within acceptable levels.
- 4. Prepare to introduce the second 1/3 of the flow (up to a total of 2/3) to the fishway by installing a seine (or other appropriate fish barrier) at the upstream end of the existing channel to prevent fish from moving downstream into a low-flow segment of existing channel. Introduce this additional 1/3 (2/3 total flow) flow into the fishway over 2–4 hours.
- 5. Repeat step #3 above.
- 6. Salvage fish from the existing natural stream channel before dropping below 1/3 of the incoming flow. Fish salvage within the YPP should begin shortly after incoming flow is cut off from entering the YPP. Depending on salvage techniques, a seine (or other appropriate fish exclusion device) may be installed at the downstream end of the existing channel being salvaged.
- 7. Remove seine nets blocking the EFSFSR tunnel entrance and exit and allow fish to move both downstream and upstream through the fishway.
- 8. Introduce the final 1/3 of the flow over a period of 2–4 hours. Once 100% of the flow is in the fishway, complete construction of diversion berm in the existing channel and remove seines from existing channel.

Specific details on the north and south portals, plus the overall design, function, operation, and maintenance of the diversion tunnel are thoroughly described in the FOMP (Brown and Caldwell, McMillen Jacobs, and BioAnalysts 2021).

# *1.7.10.2.3. Midnight Creek*

Midnight Creek is a first order, perennial, non-fish-bearing stream. The Midnight Creek stream diversion will reroute approximately 0.3 mile of the lower portion of Midnight Creek to the south, away from where it currently enters the YPP lake. The rerouted creek will be piped under haul roads so that it will enter the EFSFSR upstream of the proposed tunnel portal (Figure 12). The Midnight Creek diversion will manage flows in Midnight Creek during YPP operations and backfill activities. The creek will be restored when the newly developed EFSFSR alignment over the backfilled pit is complete and stabilized as described in Section 1.9.4.

# *1.7.10.2.4. Hennessy Creek*

Hennessy Creek is a first order, perennial, non-fish-bearing stream. Hennessy Creek will be diverted south of YPP in a pipe along the public access road at the western edge of the pit (Figure 12). The diversion will include an impounding structure, overflow weir, and diversion cleanout basin. Diverted flows will be routed to Fiddle Creek downstream of the existing Stibnite Road culvert crossing, ultimately placing Hennessy Creek flows into the EFSFSR upstream of the south tunnel portal and disconnecting flow from the current unlined ditch passing alongside the Northwest Bradley dumps. Overflow, if any, will follow the existing stream channel into the YPP.

# *1.7.10.2.5. Fiddle Creek*

Fiddle Creek is a second order, perennial, fish-bearing stream. Fiddle Creek will not be diverted. Rather, small stormwater diversions will route hillslope runoff around the Fiddle GMS and a culvert will route Fiddle Creek under the GMS, GMS access road, and public access road.

# *1.7.10.2.6. West End Creek*

West End Creek is a first order, non-perennial, non-fish-bearing stream. The approximately 1.5 mile-long West End Creek stream diversion will reroute West End Creek around the north side of the legacy West End DRSF and cross the upper benches of the WEP (Figure 12). The diversion will consist of a lined channel along the upper legacy DRSF, and a pipe in the segments along a steep hillside above the WEP, within the pit, and along the steep hillside alongside the lower legacy DRSF down to the outlet at the existing stream channel. The lined

channel portion will be designed to convey flows from a minimum 25-year storm event plus 2 ft. of freeboard.

# *1.7.10.2.7. Garnet Creek*

Garnet Creek is a perennial, first order, non-fish-bearing stream. During construction, Garnet Creek will be re-routed downstream of the ore processing facility to a relocated confluence with the EFSFSR (Figure 12). The diverted length of Garnet Creek will be approximately 600 linear ft.

Details of the Garnet Creek re-route are depicted on Drawing GSK-002 in Appendix B of the Water Management Plan (Brown and Caldwell 2021b). Above the early restoration reach which passes through the processing plant site area, Garnet Creek will be routed along the upper processing plant site access road in a riprap channel, then cross under the ore processing facility roads in culverts, with EDF to reduce sediment loading to the stream, and to protect water quality. At closure, this segment of Garnet Creek will be restored, along with created wetlands at the plant site.

# *1.7.10.2.8. Meadow Creek*

Meadow Creek is a perennial, third order, fish-bearing stream. Approximately 2 miles of Meadow Creek will be diverted around the south side of the TSF and TSF Buttress. The diversion will direct flows back into the existing SODA diversion upstream of the HFP (Figure 13). The new diversion will consist of a rock-cut channel in segments along the steep hillsides above the TSF and buttress, and an excavated channel in alluvium across tributary valley segments. Channel segments excavated in erodible or permeable materials will be lined with rock riprap and/or geosynthetic liner to prevent erosion and to minimize seepage. The Meadow Creek diversion channel around the TSF and TSF Buttress will be designed to convey flows from a minimum 100-year storm event with 1-ft. of freeboard.

The stream will also be diverted around the HFP. The Meadow Creek channel will be moved away from the pit to the south/southeast and reconstructed as a permanent, sinuous channel and floodplain to allow potential for spawning habitat and establishment of riparian habitat within the floodplain. A liner will be installed under the stream/floodplain corridor to minimize water seepage into the HFP or the pit dewatering well system, and to avoid potential pit wall instability or loss of stream habitat as a result of stream dewatering. The Meadow Creek diversion channel/floodplain corridor around the HFP will be designed to convey flows from a minimum 100-year storm event with 3 ft. of freeboard; as a natural channel design, the stream channel itself will be designed for bankfull flows (i.e., 1.5-year recurrence). This diversion will be permanent and incorporates design aspects to resemble natural channels not applied to temporary diversions of the other creeks. This permanent design accounts for channel migration, flooding, riparian development, and biological habitat. Details of the Meadow Creek diversion appear in the Water Management Plan, Appendix B (Brown and Caldwell 2021b), while details of its restoration appear in the CMP, Attachment D and are incorporated here by reference (Tetra Tech 2023).



**Figure 12. Proposed Action Water Management Plan – EFSFSR below Meadow Creek.**



**Figure 13. Proposed Action Water Management Plan – Meadow Creek.**

### *1.7.10.2.9. East Fork Meadow Creek*

East Fork Meadow Creek (EFMC, aka Blowout Creek [BC]) is a first order, perennial, fishbearing stream outside the Project operational footprint. EFMC was impacted by the failure of a water storage dam in 1965 creating a steep actively eroding channel that conveys EFMC. In its current condition, EFMC is a downcutting creek with conditions not suitable for fish occupancy. There are fish in the wetlands area upstream of EFMC that will be excluded from the EFMC restoration area by proper fish exclusion screening.

The design for the EFMC repairs appears in the Water Management Plan; Section 6.1.4 which references the 2021 Rio ASE Stream Design Report. The design entails three reaches: (1) the segment downstream from the wetlands and upstream from the downcutting creek area (design reach BC1); (2) the downcutting creek area (BC2); and (3) the approximately 400-ft. segment between the downcutting area and the confluence with Meadow Creek (BC3). Perpetua proposes to stabilize and repair the failed area of EFMC in the actively eroding chute and raise groundwater levels in the meadow upstream of the former dam site to restore wetland hydrology. A structure to control the grade of the creek will raise groundwater levels in the meadow and a coarse rock drain will address ongoing erosion of the channel side slopes that currently deliver sediment directly to the creek, while facilitating construction of a permanent surface channel. This will be an SGP EDF and restoration effort, as the EFMC chute and upper meadow are unrelated to and unaffected by the proposed mine features.

The restored stream channels in EFMC (Tetra Tech 2023, CMP, Attachment D, Sheets 66 through 71) utilize placed streambed materials to maintain flows as surface water. These reaches have gradients greater than 14 percent which will promote runoff as surface water. The EFMC restoration will affect 2,668 linear ft. of stream below a 2,000 cubic yard structure placed as a rock grade control at the outlet of the wetlands area to the eroded channel. The rock grade control will be placed during the construction phase of the Project. For details on this construction, see the CMP Attachment D, Sheets 64 and 65 (Tetra Tech 2023). Upon exiting the rock drain (i.e., design reach BC2), flow will be conveyed by a restored stream channel to the confluence with Meadow Creek (i.e., design reach BC3, Tetra Tech 2023, Attachment D, Sheets 69 through 71). The lower portion of the EFMC alluvial fan will be an important borrow area for this and other restoration projects and is included in Project disturbance. Therefore, the restored stream channel will be constructed approximately 100 feet north of the existing channel which overlies sediments from the downcutting segment which will act as borrow materials for the restoration projects.

During construction and early mining, Perpetua will construct grade control and water retention features near the old reservoir water retention dam location to elevate the groundwater level and stream water surface sufficiently to restore wetland hydrology in the surrounding meadow. The retention structure will impound portions of the meadow channel, which will fill with sediment over time.

A coarse rock drain will be constructed within the chute downstream of the failed dam site to isolate the flow of EFMC from the actively eroding chute side slopes and to prevent further erosion of the gully bottom, facilitating subsequent restoration of a surface channel on top of the drain. The rock drain will also provide area for the collection and retention of side-slope erosion material rather than allowing that material to potentially contribute sediment to EFMC. As the rock drain fills with sediment, it will become closed off from the stream channel and flow will revert to the designed surface channel.

The existing alluvial fan in lower EFMC, located adjacent to Meadow Creek, will be partially removed, mostly during mine operations for borrow materials, and the area reclaimed. A surface diversion will be constructed at the margin of the lower alluvial fan to facilitate borrow excavation, and this stream reach subsequently restored.

# *1.7.10.2.10. East Fork South Fork Salmon River*

Enhancement in the EFSFSR will begin in MY -1, and efforts will be focused on increasing hydraulic and geomorphic diversity while removing potential fish passage barriers. Enhancement in these stream reaches is intended to provide habitat for spawning and rearing Chinook salmon and steelhead. LWD and rock clusters will be used to enhance instream conditions, increase instream friction, sort sediment, and create localized velocity gradients. Grade control structures such as engineered riffles and/or channel-spanning rock or wood may be used to facilitate the development of relatively large pools intended to accommodate adult salmonids migrating upstream through these relatively steep reaches. The newly constructed EFSFSR, constructed across the surface of the YPP backfill, will be designed to interact with its floodplain, and will include side-channels, LWD, boulders, wetlands, and the lowest reaches of Hennessy, Garnet, and Midnight Creeks. See Appendix D of the Stream Design Report (Rio ASE 2021) for more detail and reach-specific designs.

#### *1.7.10.2.11. Non-Contact Stormwater Diversions*

Non-contact stormwater is meteoric water (i.e., precipitation) that does not contact tailings, open pits, the TSF, TSF Buttress, spent heap leached ore, and tailings from past mine operations, or any other mining related surfaces. IDAPA 20.03.02 and 37,03.05 set stormwater design criteria (e.g., storm events for design sizing and freeboard requirements). Stormwater runoff from undisturbed areas upslope of mine features in the major drainages will be captured in stream diversion channels described above or in other channels that will direct runoff away from mine disturbed areas (Figures 12 and 13). Smaller-scale diversion channels or earthen berms will be used, where necessary, to divert stormwater around other mine infrastructure. Non-contact water will be managed with features to reduce erosion and sediment delivery to streams. Where sedimentation is a concern, non-contact water stormwater diversions will be routed to sediment catch basins where the water can evaporate, infiltrate, or discharge into the stream system after settling. Energy dissipation structures will be installed at the non-contact surface outfalls as needed.

# *1.7.10.2.12. Contact Water*

Water that contacts mining disturbances and has the potential to impact water quality is termed contact water. Contact water includes, but is not limited to, runoff from mine facilities such as the TSF, TSF Buttress, stockpiles, mine pits, haul roads constructed of development rock, toe

seepage of precipitation infiltrating through the stockpiles, and underground exploration water. The TSF Buttress and stockpiles are unlined facilities. Therefore, water incident on the TSF Buttress and stockpiles that does not runoff will infiltrate into the buttress and then emerge as toe seepage or infiltrate into the subsurface and groundwater. The volume of runoff, toe seepage, and infiltration are accounted for in the site wide water balance (Perpetua 2021d) with runoff and toe seepage collected as contact water, and the effects of the infiltration are accounted for in the assessment of groundwater chemistry. Collection of contact water will begin during the first year of on-site construction and will continue throughout operations and the closure and reclamation phases. Contact water will be captured in channels and sumps and routed to the ore processing facility, contact water storage ponds, water treatment plant, or enhanced evaporation systems. In unusually high runoff periods collected water may be allowed to remain in the pits or the TSF temporarily, excess contact water from outside of the pits may be routed to mine pits for temporary storage. Contact water storage ponds will be lined to minimize leakage. Water in the contact water storage ponds could be pumped to the mill for use, treated and discharged in accordance with applicable requirements, or evaporated. Contact water in the mine pits will be directed to in-pit sumps in the lowest part of the pit and piped to the mill for use, to other contact water storage ponds, to water treatment or evaporation, or into trucks for spraying for dust control within open pits and on stockpiles or TSF Buttress. Any contact water beneficially used in the ore processing or for dust control or stored for more than 24 hours then treated and discharged will require water rights permitting, including mitigation as outlined in the water right permit, through the IDWR prior to use.

Contact water which exceeds regulatory discharge standards set by IDEQ and that cannot be used during operations will be disposed of through a variety of methods including forced evaporation using sprayers located within the TSF or other managed areas or treated and discharged. Water will be treated to meet IPDES permit limits and treated water will then be discharged through IPDES permitted outfalls to the EFSFSR or Meadow Creek.

Runoff from haul roads and access roads outside of pits, ore stockpiles, or development rock storage areas may be of sufficiently good quality to be eligible for coverage under the Multi-Sector General Permit (MSGP) for Stormwater Associated with Industrial Activities. Eligibility will depend upon the materials used for road construction and will be determined through coordination with IDEQ with oversight by EPA. Construction materials will be required to meet the 500 mg/kg arsenic concentration criteria associated with the protection of surface water from effects of metal leaching from the construction materials. The establishment of this criteria is detailed in the Development Rock Management Plan (Brown and Caldwell 2022). Runoff covered under the MSGP will be managed with a variety of EDF and conventional stormwater control measures to ensure the protection of surface water quality.

# *1.7.10.2.13. Surface Water Outfalls*

The specific number and exact locations of outfalls will be determined via IPDES permitting through IDEQ. Approximate locations of the anticipated outfalls described below are shown on Figures 12 and 13. All outfalls will be required to meet water quality limits for specific constituents, and some outfalls may have discharge volume limits where the permit specifies a

loading limit. Not all outfalls will necessarily be active or be permitted in the same permit cycle (e.g., post closure outfall will not be active during operations).

Two IPDES surface water outfalls will be used to discharge treated contact water from active mine pits, the TSF Buttress, pit dewatering, legacy mine materials disturbed by new mining activities, and the plant site and truck shop. One outfall located near the plant site will discharge to the EFSFSR. A second outfall will discharge to Meadow Creek upstream from EFMC to augment streamflow during pit dewatering.

Water from the TSF and TSF Buttress underdrains may be discharged from two outfalls shown on Figure 13, depending on whether IPDES discharge limits are met without treatment of the underdrain water (otherwise, underdrain water will be routed to the plant site for use in processing, to the water treatment plant, or back to the TSF). Discharges from these two outfalls are expected to have a strong seasonal component, with some parts of the year seeing reduced flows, or even no discharge, as contact water is used for ore processing or other mine uses. The expected water treatment and discharge rates range between zero and 2,000 gpm with the largest rates expected in MYs 5 and 6 when dewatering production is greatest and larger than the volume needed for use in the processing plant. Aside from those years, water treatment and discharge rates are zero during the summer months when water use is greatest then range between 200 and 1,000 gpm in the winter months.

An outfall will be permitted on upper EFSFSR for the sanitary wastewater treatment facility at the worker housing facility. That outfall will be active through the operations period and during mine closure until the facility is decommissioned. An additional outfall is expected to be permitted in a future IPDES permit renewal for closure and post-closure discharge of treated TSF process water. That outfall will be on Meadow Creek upstream of EFMC near the TSF Buttress.

Additional permitted outfalls may be necessary during a portion of the operations period for contact water storage pond spillways that could discharge to surface water – although discharge will be very rare or non-existent, only occurring in the event of excessive precipitation or snowmelt. The need for additional outfalls associated with pond spillways and their location will be determined with IDEQ. Each outfall will be permitted through IDEQ and will be required to be monitored, meet discharge limits, and regulate the rate of discharge.

# *1.7.10.2.14. Draining the Yellow Pine Pit Lake*

Draining of the existing YPP lake will be initiated during construction. When the EFSFSR tunnel diversion is ready, stream flows will start being diverted into the tunnel during a period of low flow, most likely in the warmer months, and concurrent with salvaging fish from the pit lake and diverted sections of the EFSFSR. As the EFSFSR water is diverted into the tunnel, the decreased EFSFSR flow into the pit lake will be expected to cause some fish to out-migrate, thereby lessening the number of fishes requiring salvage and creating better conditions for salvaging fish. The period of fish salvage between the start of water diversion to full diversion into the EFSFSR tunnel is expected to be approximately one week. The methods applied for fish salvage are described in Appendix A, Table A-1.

Once fish salvage has occurred in the EFSFSR from the tunnel diversion downstream to the YPP Lake and most of the EFSFSR flow is being diverted into the tunnel, fish salvage in the lake will commence and take approximately one week to complete. The YPP Lake will drain naturally down to the elevation of the outlet of the lake, where the existing rock sill will control the water level, though some leakage and slow lowering via groundwater outflows may occur beyond that level. No erosion or downcutting of the outlet rock sill will be expected because it has endured the full range of EFSFSR flows over decades and both inflow and outflow rates will be minimal during draining due to the river flow being diverted into the tunnel. The drain-down process will naturally convey lake water downstream to the EFSFSR.

After the natural drain down, water remaining in the YPP Lake or entering the pit from groundwater seepage or local stormwater runoff from pre-stripping operations on the highwalls above the pit lake will be managed as mine-impacted water. The water pumped from the pit lake will be used for construction purposes, transferred to the TSF (after it is lined and available) for future use in ore processing, or treated to meet permit limits before being discharged downstream in the EFSFSR via an IPDES permitted outfall.

Sediment remaining in the pit lake bottom will be removed beginning near the end of the final year of construction. Approximately 80 vertical feet of sediment lies on the pit bottom, and the pit walls are too steep to operate equipment without a ramp. Therefore, removal may be staged to coincide with successively lower benches as the pit is mined, and therefore may extend into the first year of operations. During this time, the pit will be used seasonally to capture and store contact water from the adjacent pit walls, and this water will be used or managed as stated above.

The sediment will be removed using an excavator or similar equipment and loaded into trucks and delivered to the TSF. Slurry/dredging methods are not anticipated but will be considered as part of adaptive management if the sediments are too wet to load and/or blend. The truck beds will have flashboards to minimize water leakage from the low-strength, saturated sediments. The loading area will drain back into the former pit lake, preventing off-site discharge of bleed water during loading. If necessary, wet material will be blended with loose dry material (e.g., development rock) from elsewhere on site to enable better loading, transport, and ultimate stability at the destination.

# *1.7.10.2.15. Surface Water Chemistry*

To minimize the volumes of contact water encountering Project disturbance and requiring treatment, the Project will divert upstream non-contact water to prevent it from interacting with SGP facilities during operations. Table 17 provides a summary of the non-contact diversion channels that are considered in the site-wide water chemistry (SWWC) model (SRK 2021a). At closure, the diversion channels will be decommissioned, and non-contact water will follow its natural drainage pathways.

<b>Diversion Channel</b>	<b>Description</b>
North Diversion	Diverts non-contact runoff from the north of the TSF and TSF Buttress to Meadow Creek
South Diversion	Diverts Meadow Creek and its tributaries from the south and west of the TSF around the <b>TSF</b>
<b>Hennessy Diversion</b>	Diverts water from Hennessy Creek away from the YPP to Fiddle Creek
Midnight Diversion	Diverts Midnight Creek away from the YPP to the EFSFSR
<b>West End Diversion</b>	Diverts upper West End Creek around the West End Pit
<b>EFSFSR</b> Tunnel	Diverts the EFSFSR around the YPP downstream of YP-SR-6 to upstream of YP-SR-4

**Table 17. Summary of Diversion Channels included in the Surface Water Chemistry Model.**

Source: SRK 2021a

### **Groundwater Management**

Groundwater will require management to allow mining in the pits and to direct seeps and springs from beneath mine facilities. Groundwater also will provide a portion of the water supply for the SGP. Water supply aspects of the mine operations are described in the Water Use and Water Balance subsection below. Any groundwater used within the SGP will require permitting through IDWR prior to use. Depending on final use or disposal of groundwater, wells drilled on the site could be permitted as domestic use, industrial use, or dewatering wells.

### *1.7.10.3.1. Pit Dewatering*

Lowering the water table in and surrounding the Yellow Pine, Hangar Flats, and West End Pits during operations will increase pit wall stability and provide dry working conditions in the pit bottoms. Development of the Yellow Pine and HFPs will require partial dewatering of the alluvium of portions of the EFSFSR and Meadow Creek valleys, respectively, to limit groundwater inflow to the pits and maintain stability of the pit slopes. Once the WEP is mined below the level of West End Creek, the WEP also will require dewatering.

Dewatering will be accomplished by drilling a series of alluvial and deeper bedrock wells near the pit perimeters to intercept and pump groundwater before the water reaches each pit. Alluvial groundwater at the Yellow Pine and HFPs will be managed using a series of vertical wells (Figures 12 and 13). The WEP is primarily in bedrock with only a thin layer of alluvium in the vicinity of the pit and no alluvial dewatering is planned for that pit. Pumps will be installed in each well and will run as necessary to draw down the groundwater and facilitate mining and backfilling operations. Horizontal drain holes in pit walls may also be considered for depressurizing remnant high pore pressure areas.

Groundwater pumped from pit dewatering will be considered to be contact water and will be managed through forced evaporation or active water treatment when the volume of pumped water exceeds the ore processing facility demand. Treated water will be discharged to either of two IPDES-permitted outfalls, either Meadow Creek or the outfall on the EFSFSR near the water treatment plant, depending on the need for streamflow support in Meadow Creek. The pit

dewatering wells will be permitted as industrial wells in conjunction with a water right application through IDWR.

Groundwater not captured by the pit dewatering, and entering the pits as highwall seepage, will be directed to an in-pit sump in the lowest part of the pit where it will combine with stormwater and snowmelt runoff (i.e., contact water) from precipitation falling within the pit. The water will be used for dust control within the pits, and as needed, pumped to the ore processing facility for use as makeup water. In-pit water that cannot be used will be disposed of through forced evaporation or routed to the water treatment plant then discharged to the EFSFSR or Meadow Creek via IPDES permitted surface outfalls.

# **Climate Change**

Per the definitions utilized by the NOAA Fisheries WCR Guidance to Improve the Resilience of Fish Passage Facilities to Climate Change (NMFS 2023), the Project is a long-term project with a lifespan of more than 10 years with three ESA-listed fish species present (spring/summer Chinook salmon, steelhead, and bull trout). The Project includes potential risk pathways for fish passage including water diversion for consumptive use in accordance with water rights, construction of a fishway, installation of culverts and bridges for stream crossings, and the potential use of trap and haul as a secondary approach for fish passage in the event that primary approaches are not meeting targets. In addition, mine closure and site restoration activities include stream restoration through the mine area with fish passage in a designed channel.

The Project water balances and temperature forecasts utilized by Perpetua and the USFS for the development of the mine and closure designs employed sensitivity analyses to examine the potential range of future site conditions. These forecasts developed ranges for water diversion for consumptive use, contact water management, water treatment, restored stream channel flows, and riparian shading of restored stream segments. These ranges were incorporated into the component designs to allow for variability in stream flows (both high and low flows) and air temperature conditions. Idaho requirements for sizing channels and ponds were also applied so that the facilities will be engineered and constructed to contain high flow storm events with sufficient capacity and freeboard.

Because the operating lifetime of the Project will be approximately 15 years following a threeyear construction period, effects of climate change during the operating lifetime will be limited by that duration. Therefore, there is less potential for climate change to affect Project operational components than components associated with mine closure and restoration. As such, the climate change risks associated with the post-closure components of the Project were prioritized. These post-closure risks were incorporated into the designs of the restored stream channels and their associated riparian shading so that the stream channels will respond to variable climatic conditions in a manner similar to natural drainages. Unlike the restored stream channels, water diversion for consumptive use and trap and haul practices will not continue past the operational period. Furthermore, access roads and stream crossing constructed for the Project will be removed during mine reclamation and closure.

The Project is located with the Columbia River watershed. As such, application of the NMFS guidance (NMFS 2023) calls for use of the Bureau of Reclamation's West-Wide Climate and Hydrology Assessment (BOR 2021) for forecasting future trends in stream flow and temperature as affected by climate change. Within the Columbia River watershed, the assessment makes the following forecasts from the present through 2021:

- Warming air temperatures by 5°F with a range of  $+/-$  5 degrees based on the 10<sup>th</sup> and 90<sup>th</sup> percentiles of forecasts used;
- Consistent total annual precipitation within a range of  $+/-30$  percent;
- Decreasing April 1<sup>st</sup> snow water equivalents by 25 percent within a range of  $+/45$ percent;
- Consistent total annual runoff within a range of  $+/45$  percent;
- Increasing December-March runoff by 10 percent within a range of  $+/-65$  percent;
- Decreasing April-July runoff by 10 percent within a range of  $+/-75$  percent;
- A six month increase in mean drought durations; and,
- An increase in the mean Palmer Drought Severity Index from 1.4 to 1.7.

Based on this assessment (BOR 2021), the effects of climate change will primarily manifest as the following risk pathways identified in the NMFS guidance:

- Drier dry extremes;
- Decreased minimum flows;
- Runoff starting earlier in the year;
- Increased water temperature; and,
- Increased wildfire effects.

These forecasts are qualitatively consistent with Department of Agriculture climate change assessments for the area (Halofsky et al. 2018). Project area flow data collected from U.S. Geological Survey (USGS) gauge 13313000 on Johnson Creek between 1929 and 2017 is also consistent with runoff earlier in the year when comparing recent periods (1988-2017 and 2012- 2017) to earlier periods (1959-1988) and the overall record (1929-2017) (Figure 14).

Instantaneous peak flow events observed by USGS gauge 13313000 indicated lower low flow events and potentially slightly increased high flow events (Rio ASE 2019). This range of potential runoff conditions was utilized in the design of the restored stream channels.

To help inform Project design, Table 18 as a crosswalk to summarize the potential risks and mitigation measures identified by applying the risk pathways from the NMFS guidance (NMFS 2023) to the Project under the BOR assessment (BOR 2021) results.



**Figure 14. Average Monthly Discharge at USGS 13313000 for Various Historic Time Periods, Stibnite Gold Project.**

**Note:** Month 1 is January.

<b>Element</b>	<b>Aspects</b>	Risk Pathways <sup>1</sup>	Mitigation (if/as needed)
<b>Water Diversion</b> for Consumptive Use	<b>Water Rights</b>	Dry extremes - not able to divert at the desired rate Decreased minimum flows – not able to divert at the desired rate Wildfire effects – debris removal and maintenance	Alternative water supplies (i.e., recycled process water, contact water, mine dewatering water, and groundwater pumping) will be employed prior to diversion from stream flow (per the Water Management Plan [WMP]). Inspections and maintenance programs will be used to remove debris from facilities (per the WMP).
	Channel Stability	Wildfire effects – debris removal and maintenance	Inspections and maintenance programs will be used to remove debris from facilities (per the WMP).
	Predation	Dry extremes – more exposure to predators Decreased minimum flows – more exposure to predators Water temperature increase – more exposure to predators	Trap and haul will be employed to relocate fish to lower risk conditions (per the Fishway Operations Management Plan [FOMP]).
	Hyporheic Flow	Water temperature increase – reduced hyporheic flow	Water temperatures will be controlled via design features and mitigation measures including use of pipelines, riparian shading, and mechanical shading (per the WMP, Stream Design, and mitigation measures).
	Screen Design	Water temperature increase – biofouling potential Wildfire effects – debris removal and maintenance	Screen design includes brushing system to remove fouling. Inspections and maintenance programs will be used to remove debris from facilities (per the WMP).
	Screen Cleaning	Runoff Timing Shift - shorter season for maintenance. Water temperature increase - weaker species reduced swimming performance Wildfire effects – debris removal and maintenance	Inspections and maintenance programs will be used to remove debris from facilities (per the WMP). Fish swimming will be monitored and addressed as necessary via trap and haul (per the FOMP).
	Structural	Wildfire effects – debris removal and maintenance	Inspections and maintenance programs will be used to remove debris from facilities (per the WMP).

**Table 18. Summary of Risks Associated with Climate Change.**





 $1$  NMFS 2023.
### **1.7.11. Sanitary Waste Handling Facilities**

Sanitary waste handling facilities will be present at SGP facilities and will be constructed and operated in accordance with Valley County, IDEQ, and Idaho Department of Health and Human Services standards. Sanitary wastewater will be treated using membrane bioreactor (MBR) or similar technology. Early in construction, the currently permitted MBR plant at the existing exploration camp will be used, and treated effluent reused for flushing toilets and urinals (as allowed by Perpetua's existing Reuse Permit M-228-02) or discharged to the existing drain field located in the process plant area, while the worker housing facility and its associated treatment plant is under construction. During operations and closure, sanitary wastewater from the worker housing facility, ore processing facility, and administration buildings will be treated at a new MBR or similar plant and discharged to the EFSFSR via a permitted IDPES outfall. Vaults or portable toilets will be used at off-site facilities and remote locations on site (e.g., TSF, pits, maintenance facility, etc.), and serviced as needed using vacuum trucks. Treatment residuals will be hauled off site to a permitted sanitary landfill. Vault/portable toilet wastewater will be hauled to the on-site sanitary wastewater treatment plant for treatment.

### **1.7.12. On-site Composting Facilities and Solid Waste Collection and Disposal**

On-site composting facilities will be permitted by IDEQ with oversight by the local Health District. Small scale composting associated with organic materials generated at the worker housing facility may be incorporated within the centralized GMS in the Fiddle Valley. These composting facilities will be fenced. Any larger composting facilities deemed necessary to support GM quality or quantity improvements will be located off site.

All construction and demolition waste generated at the SGP will be hauled off site for disposal at a permitted landfill; a landfill will not be constructed or maintained at the SGP. Solid waste from the worker housing facility, shops, and other work areas that cannot be composted or recycled will be collected in wildlife-resistant receptacles and hauled off site for disposal in a municipal waste landfill. Material that meets the classification of a "hazardous waste" will be collected and stored, per the SGP Waste Management Plan at specially designed and operated secured satellite collection sites and a main storage site prior to shipment off-site to a Resource Conservation and Recovery Act certified hazardous waste disposal facility.

#### **1.7.13. Mine Site Borrow Sources**

Various types of earth and rock material will be used from borrow sources for construction, maintenance, closure, and reclamation activities. Most of these materials can be sourced at the SGP from existing development rock dumps, legacy spent heap leach ore, and from development rock removed as part of proposed surface mining and underground exploration activities. These materials will be subject to physical and chemical testing to determine suitability for use.

Native earth materials will be required for some applications. Specific areas within the SGP that have large quantities of high quality native alluvial and glacial granular borrow materials for use include:

- The alluvial and glacial soils in the Meadow Creek valley floor within the footprint of the TSF, TSF Buttress, HFP, and YPP;
- Sand, gravel, and cobbles in the lower EFMC alluvial fan; and,
- Glacial soils in the Fiddle Creek valley walls within the footprint of the Fiddle GMS.

### **1.7.14. Materials, Supplies, Chemical Reagents, and Wastes**

Numerous materials, supplies, and chemical reagents will be used, including fuel, explosives, and ore processing reagents for the SGP. A Spill Prevention, Control and Countermeasure (SPCC) Plan will be developed prior to construction to establish procedures for responding to accidental spills and releases of petroleum products. In addition, a Hazardous Materials Handling and Emergency Response Plan will be developed prior to construction to address procedures for responding to accidental spills or releases of hazardous materials to minimize health risks and environmental effects (see Appendix A, Table A-7 for more details of these measures per the Project's TMP [Perpetua 2022]).

### **Diesel Fuel, Gasoline, and Propane**

Aboveground storage tanks at the SGP will be used for fuels and other fluids, including gasoline, diesel fuel, lubricants, coolants, hydraulic fluids, and propane. Approximately 200,000 gallons of diesel fuel, 10,000 gallons of gasoline, and 30,000 gallons of propane will be stored at the SGP in addition to a variety of materials, supplies, and reagents (Table 19). The aboveground storage tanks will be installed on containments sized to contain 110 percent of the capacity of the tank. Refueling will occur on concrete-paved areas designed to contain refueling spills (i.e., berms around their perimeters). There will be no below ground fuel storage or piping used for refueling. Storage management will be outlined in the SPCC Plan. The storage tank facility for gasoline, diesel fuel, and propane will be located near the maintenance workshop with additional propane storage at the ore processing facility area, the underground portal area, and the worker housing facility.

<b>Common Name</b>	Units	Annual Use	Delivery Form	Typical Vehicle Payload	On-site Storage Capacity	<b>Storage Method</b>	On Site Mine Uses	Estimated <b>Deliveries</b> per Year
Diesel Fuel	Gallons	5,800,00 $\Omega$	<b>Bulk liquid</b>	10,000	200,000	Tanks	Mine Site	580
Lubricants	Gallons	296,000	<b>Bulk liquid</b>	3,000	30,000	Tanks, Totes, Drums	<b>Truck Shop</b>	99
Gasoline	Gallons	500,000	<b>Bulk liquid</b>	5,000	10,000	<b>Tanks</b>	Mine Site	100
Antifreeze	Gallons	40,000	<b>Bulk liquid</b>	3,000	4,000	Tanks, Totes, Drums	<b>Truck Shop</b>	13
Propane - Buildings	Gallons	560,000	<b>Bulk liquid</b>	6,000	30,000	Tanks	<b>Buildings</b>	93
Propane – Lime Plant	Gallons	1,463,00 $\Omega$	<b>Bulk liquid</b>	11,000	30,000	<b>Tanks</b>	Lime Plant	133
Solvents	Gallons	1,000	<b>Bulk liquid</b>	200	1,000	<b>Totes or Drums</b>	<b>Truck Shop</b>	5
<b>Tires</b>	Each	246	<b>Bulk</b> solid	Variable	59	Laydown	Mining	47
<b>Batteries</b>	Units	Variable	Pallets	Variable	500 units	Pallets	Mining	25
<b>Light Ballasts</b>	Pounds	25	Pallets	Variable	1,000	Pallets	General Operations	5
Pesticides/Insecticides	Pounds	250	Pallets	Variable	1,000	Pallets, drums	General Operations	1
Herbicides	Pounds	1,000	Pallets	Variable	2,000	Pallets, drums	Environmental	1
Fertilizer	Pounds	2,500	Pallets	Variable	5,000	Pallets, drums	Reclamation	$\mathbf{1}$
<b>Ammonium Nitrate</b>	Tons	7,300	<b>Bulk</b> solid	24	200	<b>Secured Silos</b>	Open Pits - blasting	304
Explosives	Tons	100	<b>Boxes</b>	5	20	<b>Secured Magazines</b>	Open Pits - blasting	20
Grinding media, SAG Mill	Tons	4,449	<b>Bulk solid</b>	24	200	<b>Bunkers and Bins</b>	Mine Process Area	186
Grinding media, Ball Mill	Tons	3,566	<b>Bulk solid</b>	24	200	<b>Bunkers</b> and Bins	Mine Process Area	149
Grinding media, LS Ball Mill	Tons	34	<b>Bulk solid</b>	24	200	<b>Bunkers</b> and Bins	Mine Process Area	$\overline{c}$

**Table 19. Proposed Materials, Supplies, and Reagents.**





**Source:** Perpetua 2021a.

**Key:** AP = AP 3477 is dialkyl dithiophosphate, a reagent used in the flotation circuit; BM = ball mill; ISO = International intermodal container that is manufactured according to the specifications outlined by the International Organization for Standardization (ISO);  $kg = kilogram$ ;  $LS = limestone$ ;  $SAG = semi-autogenous$  grinding

### **Explosives Storage**

Ammonium nitrate prill will be received in bulk in tanker trucks and transferred into storage silos. Other blasting supplies used for mine blasting operations will include blasting emulsion products, detonating cord, cast primers, and blasting caps. These products will be delivered in boxes or other approved containers on trucks. The explosives storage facility will include two silos containing ammonium nitrate on a concrete pad and two buildings, one for explosives and one for detonators. Components of bulk explosive material will be stored in separate and isolated containers, sized, and designed to meet Bureau of Alcohol, Tobacco, Firearms, and Explosives and MSHA requirements. The explosives storage facility will be fenced and securely gated. An explosives contractor will provide the products and manage the explosives storage facility.

### **Miscellaneous Oils, Solvents, and Lubricants**

Various oils including motor oils, lubricants, antifreeze, and solvents will be shipped to the SGP on trucks. These will be stored in approved containers located within, or directly adjacent to, the maintenance shop and contained within secondary containments to prevent spills into the environment. All used petroleum products, waste antifreeze, and used solvents will be collected in approved containers, transported off site, and disposed or recycled.

### **Miscellaneous Consumables**

Lime will be produced on site and stored in silos at the ore processing facility. Silos will be equipped with air emission controls. Sodium cyanide will be transported as dry cyanide briquettes to the SGP. Nitric and sulfuric acid will be transported in tanks designed to prevent spills even in the event of rollovers. Nitric and sulfuric acids will be stored in specialized noncorrosive, polyethylene‐lined tanks located within the ore processing facility and will have secondary containment.

# **Waste Handling**

Wastes anticipated to be generated at the SGP include fluorescent bulbs, batteries, and empty aerosol containers which will be managed in accordance with the appropriate regulatory standards. Materials that are not consumed will be recycled, to the extent practical, or disposed of in accordance with applicable regulations.

Used petroleum products will be stored on site in approved containers that will be separate from other trash and garbage products. Used petroleum products will be transported off site for recycling or disposal in an approved facility. Other legacy materials could be encountered during construction and operations. If encountered, these materials will be characterized to determine potential for re-processing, reuse, or on-site or off-site disposal.

# **1.7.15. Temporary Closure of Operations**

No periods of temporary or seasonal closure are currently planned; however, a description of temporary closure is required for the SGP cyanidation permit if applicable. In the event of temporary suspension of mining activities, Perpetua will notify the USFS, USACE, IDEQ,

IDWR, IDL, and Valley County in writing with as much advanced warning as possible of the temporary stop of mining activities. This notification will include reasons for the shutdown and the estimated timeframe for resuming production.

During any temporary shutdown, Perpetua will continue to implement operational and environmental maintenance and monitoring activities to meet permit stipulations and requirements for environmental protection. This will include the reclamation success monitoring.

Dewatering of the open pits may continue during temporary closure due to the negative effects that pit lake formation or highwall saturation will have on highwall stability and renewed mine operations. Since ore processing may not be occurring, excess water from the various facilities will need to be managed. The operational plans required by the Cyanidation Permit and other plans developed as part of IDEQ permits will also describe specific activities and provide details on how process water will be managed during a temporary closure. Process water will continue to be managed per IDEQ requirements during any temporary closure including water collection and water treatment of excess water volumes beyond the capacity of the system to store and recycle.

A limited potential exists that unfinished facilities (e.g., haul roads, buttress, open pits, pit backfills, GMSs, etc.) will not have the same protective measures in place (e.g., stormwater collection systems or culverts) as will exist if the facility had been finished. Therefore, Perpetua will identify interim measures that will be taken to manage stormwater, sediment, dust, and other factors while the mining is temporarily stopped. Surface water diversion structures are all proposed to be installed prior to construction of the TSF, open pits, and the TSF Buttress; hence, surface water will be diverted around these facilities regardless of the stage of their completion.

Environmental reports will be submitted per previously agreed upon schedules. Regardless of the operating status of the mine, appropriate monitoring will continue until compliance with permanent regulatory closure requirements is attained, unless modified by the required regulatory authorities.

# **1.8. Surface and Underground Exploration**

Surface and underground exploration including development drilling will occur to evaluate potential mineralized areas outside of the proposed mining areas. New surface and underground exploration activities will be conducted during construction and operations. Any additional future expansion of mining activities will require supplemental permitting and approvals, including additional evaluation under the National Environmental Policy Act (NEPA).

# **1.8.1. Surface Exploration**

A total of approximately 65 acres of exploration drilling disturbance within the operations area boundary is included in the activities associated with the SGP (i.e., 25 acres of temporary roads and 40 acres of drill pads) (Figure 15). With the exception of 11 planned locations, exact locations of the exploration drill pads have not been determined, although general areas for foreseeable exploration have been identified. These areas are displayed on Figure 16:

• Five areas surrounding the WEP.

 $\overline{a}$ 

- Two areas immediately east and west of the YPP.
- An area adjacent to the Fiddle Creek GMS.
- Three areas near the former townsite and electrical transmission line corridor (including the IPA and IPAB areas from the Golden Meadows Exploration Project [USFS 2016]).
- Two areas immediately north and northeast of the HFP.
- An area immediately north of the process plant.
- An area approximately a quarter mile south of the process plant.
- Two areas north and east of the Scout Prospect decline.
- Two areas in the West Rabbit and East Rabbit areas from the Golden Meadows Exploration Project (USFS 2016).
- Nine areas in southeast of the Midnight Pit area, between the pit area and the existing radio communications tower one-half mile to the southeast (including the Broken Hill, Ridgetop, Saddle, Upper Midnight, UM2, Doris K, Garnet, and West Garnet areas from the Golden Meadows Exploration Project [USFS 2016]).

These foreseeable exploration areas will be used for exploration drill holes and for installation of monitoring wells associated with permit monitoring requirements. These foreseeable drill areas are offset from flowing streams with no exploration areas along Sugar Creek.

For this exploration work, Perpetua will use the same or similar drilling methods and environmental protection measures that have been employed for exploration drilling in the past (i.e., exploration under the Golden Meadows Exploration Plan) and will use appropriate drilling equipment (i.e., helicopter-delivered drill rigs, truck or crawler-mounted drill rigs)<sup>4</sup>. Some drill holes will extend to 1,500 ft. or more, but the average drill hole will be approximately 800 ft. long. Drill holes will be both vertical and angled with some holes converted to monitoring wells when completed. Further exploration work will continue under the approved mine plan. The environmental commitments applicable to this exploration work are described in detail in the Decision Notice and Finding of No Significant Impact, Golden Meadows Exploration Project, Attachment A, Section 1.12 (USFS 2016).

<sup>4</sup> Because protection measures associated with the Golden Meadows Exploration Plan were cited for application to the SGP but not detailed in the BA, NMFS has provided additional details on the most notable operating procedures and EDFs for protection of ESA-listed species and DCH from the Golden Meadows Exploration Project in Appendix B.



**Figure 15. Surface Exploration Boundary, Stibnite Gold Project.**



**Figure 16. Foreseeable Exploration Targets, Stibnite Gold Project.**

Reverse-circulation rotary or sonic drills will be used to drill pre-collars for core holes, drilling down to the depth desired for the start or core collection before mobilizing a core drill onto the hole. Pre-collared holes will have surface completions/seals and be capped when completed. In practice, pre-collared holes will only be associated with road accessible drill sites.

Standard drilling procedures utilize crews consisting of a drill operator plus one or two assistants. A geologist oversees drilling activities and compliance with permit requirements, environmental protection measures, and safety procedures.

Drilling support equipment will include helicopters, water trucks, crew trucks, portable mud tanks, pipe trucks or skids, portable toilets, light plants, portable generators, motor graders, excavators, dozers, and product storage pallets. Perpetua will maintain a helipad for exploration and medical evacuations adjacent to the administration offices and warehouse facilities. Helicopter support for exploration activities will occur during daylight hours.

Where practicable, Perpetua will establish drill pads in reclaimed roadbeds and only open temporary roads in the vicinity of authorized mine disturbance in order to access exploration targets. Each drill pad will have between one and five drill holes depending on its location and exploration needs. Placement of drill pads will be guided by exploration requirements, geotechnical studies, geochemical sampling, and groundwater monitoring needs. Perpetua will utilize a rolling maximum of five acres of active temporary road disturbance (10,500 liner ft. of road) and eight acres of active drill pad disturbance (140 pads) within the total 65-acre authorization, and will reclaim road and pad disturbance to remain below that rolling maximum of active disturbance. This reclamation of exploration disturbance will be conducted as soon as practicable following data collection and at least three growing seasons will be needed to establish vegetation and determine reclamation success.

New drill pad disturbance will be kept to the minimum acreage necessary for safe access and working area for equipment and crews. Drill pad sizes will vary depending on the type of drilling work being conducted. Truck-mounted or crawler-mounted drill rigs typically require a 75 to 100-ft.-long by 50 to 60-ft.-wide working area (less than 0.15 acres). Drill pads supported by helicopter require working areas approximately 45 ft. long by 35 ft. wide working areas (less than 0.05 acres). The actual disturbance of each drill pad is dependent on the drill rig utilized, the number and orientation of drill holes on the pad, the steepness of the area topography, and the location of existing access roads.

Water and non-toxic approved drilling fluids will be utilized for all drilling activities. Drilling water will be obtained from currently approved sources plus new approved sources associate with the Project, subject to water rights and appropriations.

Sediment basins and traps (i.e., excavated sumps and/or portable tanks) will be used at each drill site to collect drill cuttings and to manage and circulate drilling fluids. Typical dimensions of road access drill sumps are 16 ft. long, by 8 ft. wide, by 8 ft. deep, with helicopter supported drill sumps being approximately 12 ft. long, by 6 ft. wide, by 3 ft. deep. If needed to manage excess water produced from the exploration drill hole, larger and/or additional sumps will be installed. Upon completion of drilling, sumps will be backfilled and reclaimed.

Exploration drill holes will be abandoned by backfilling holes with drill cuttings, concrete, cement grout, or bentonite grout consistent with IDAPA 20.03.02.060.06(c). Dewatering and monitoring wells will be abandoned with surface completions/seals and be capped consistent with IDAPA 37.03.09 – Well Construction Standards Rules. Pre-collared holes will only be associated with track or truck mounted drilling equipment.

Areas disturbed for exploration will be contoured to blend into surrounding terrain; water bars and surface water channels will be retained to handle flows through the area. Compacted areas will be decompacted as necessary prior to fertilizing and seeding. Previously approved activities (i.e., approved exploration activities and associated reclamation obligations) will continue as well.

# **1.8.2. Burntlog Route Geotechnical Investigation**

The BRGI is a specific case of surface exploration activity that will take place in MY -3 to conduct geophysical investigations to explore and characterize geotechnical conditions along the Burntlog Route to confirm that conditions align with engineering designs for the roadway and stream crossings. The BRGI, will review locations for rock quarries, bridge abutments, and cut slopes. Four geophysical investigation methods will be used: dynamic cone penetrometer testing (DCPT); test pits; drilling with truck and track mounted rigs; and helicopter supported drilling. The geotechnical investigation will assess 24 locations along the Burntlog Route via 40 borings, test pits, or cone penetrometer tests (Figure 17), including:

- Four DCPTs using handheld equipment;
- 14 test pits approximately 3 ft. wide, 10 to 15 ft. deep, and 10 ft. long, using a track mounted excavator;
- Eight boreholes using truck or track mounted hollow stem augur/core rig; and,
- 14 boreholes using a helicopter assisted core rig.

The investigation will result in approximately 0.6 acres of total ground disturbance. Nine locations (B2, B7, B9, B10, B20, and B22), will be located within RCAs, resulting in approximately 0.1 acres of disturbance within the RCAs. Six of the locations selected to characterize geotechnical conditions for stream crossings (B1, B2, B4, B5, B7, and B22) are located within 50 feet of flowing water. Two locations will require access by via off-road travel through areas with wetlands characteristics (601 feet of travel to B3 and 432 feet of travel to B22).

The Burntlog Geotechnical Investigation will be conducted in conformance with all project requirements and EDFs (see Table 21 and Appendix A). In addition to those Project-wide requirements, there are specific standard operating procedures (SOPs) and requirements applied to the geotechnical investigation (Table 20). Additional detail regarding this portion of the proposed action has been excerpted from the BNF's original BA for the project (BNF 2022) and is included below.



**Figure 17. Burntlog Geotechnical Investigation, Stibnite Gold Project, Stibnite, Idaho.**

# **Table 20. Additional Standard Operating Procedures and Requirements for the Burntlog Geotechnical Investigation (Stantec 2024; BNF 2022).**



Following large storm events, the intake pumps will be inspected to determine if stream flow has encroached into the pump area and if the pump needs to be moved so it remains above flowing water. Pump placement will generally be placed above the Ordinary High-Water Mark.

A spill prevention and clean-up kit would be placed at the intake pump site and will consist of absorbent pads and/or boom (which would be sufficient length for a worst-case scenario), drip pan, a shovel, and a fire extinguisher.

**Standard Operating Procedures** Spare fuel for the water intake pump will be stored in approved  $[29 \text{ CFR } 1926.152(a)(1)]$  fuel storage containers placed into a secondary containment vessel capable of holding at least 120% of the volume of the fuel in the fuel container. Fuel would be stored on private property in sealed 55-gallon steel drums, approved double-walled fuel tanks, or in approved single-walled tanks within secondary containment. Fuel would be managed, tanks would be inspected, and any oil release would be responded to in accordance with the SPCC plan. Bulk fuel tanks (storage vessels greater than 55-gallons) will be stored in approved double-walled fuel storage containers sitting inside tertiary containment on private property, outside the riparian conservation areas (RCA). Intake pumps, fuel storage, and containments will be inspected at each refueling and periodically between refueling. Any on-site portable toilets will be located away (generally 100 feet) from any surface water bodies and will be serviced by a state licensed sewerage waste disposal contractor. A portable toilet would be set-up adjacent to select drill sites or at the closest feasible location, serviced by a licensed contractor, and removed upon completion of drilling at each site. Boreholes are promptly abandoned as required by the Idaho Rules Governing Exploration, Surface Mining, and Closure of Cyanidation Facilities (IDAPA 20.03.02) after reaching their total planned depth. Borehole abandonment will generally take place within hours of borehole completion to avoid the need to bring the drilling rig back to the site later. If the annular space of the casing has been sealed with cement (as is the case with boreholes expected to encounter artesian conditions), the casing is left in place. If the annular seal is bentonite, the temporary surface casing is removed before abandonment. Each drill hole will be abandoned from the bottom to the collar by filling the hole with a thickened grout mixture or bentonite chips. Once the drill hole is completely sealed the project geologist or drilling manager will approve the hole-plugging operation prior to the rig being moved off site to ensure integrity of the hole plugging process. The timing for initiating and completing borehole abandonment is as soon as practicable after the geologic information has been interpreted. Abandonment of each borehole will be properly documented. Exploration drilling will not occur when saturated roads SOPs apply. Petroleum products used to support drilling activities would be transported in accordance with Idaho and U.S. Department of Transportation regulations and handled and stored as required by state and federal petroleum product storage and handling laws and regulations. Petroleum products would be kept in containment and spill prevention kits would be available on site. On drilling sites where sumps are necessary but impractical due to slope or soil conditions, a casing diverter and a hose will be used to divert drilling water to a sump located in an adjacent and appropriate site. The following SOPs are designed to minimize the risk of drilling mud discharging at the ground surface: • Drill pad locations within RCAs would require USFS concurrence that no reasonable alternative location exists.

- Drillers would be informed of these locations and would exercise increased vigilance for instances of lost circulation at shallow depths.
- The casing would be advanced simultaneously behind the core drill through the alluvial section of all drill holes.

If drilling fluid should discharge at the ground surface despite the above preventative SOPs, the following new response SOPs have been developed to minimize the risk of drilling fluid subsequently reaching live water.

- Adjacent slopes below the drill rig and stream channels (if drilling in RCAs) in these areas would be regularly monitored during drilling by environmental technicians for any evidence of surface leakage.
- Silt fence, straw wattles, portable sumps, pumps, and hoses would be pre-staged for emergency use. These materials and tools would be used to quickly construct temporary sumps to capture drilling fluid and return it to the drill rig.
- For locations that are deemed to be of sufficient risk to warrant the pre-staging of response materials, a USFS representative would verify that such measures are in place on the ground prior to drilling.
- Prestaging of response materials will occur at all drilling sites.

Section 6 of Idaho Department of Land's (IDL' Best Management Practices for Mining in Idaho (IDL 1992), would be observed, including if water is encountered in exploration holes; it would be sealed off during abandonment to prevent crossflow.

**Standard Operating Procedures** All activities would be conducted in accordance with Idaho environmental anti-degradation policies, including Idaho Department of Environmental Quality (IDEQ) water quality regulations at IDAPA 58.01.02 and applicable federal regulations.

Stormwater monitoring, inspections, and reporting would be conducted in accordance with the National Pollutant Discharge Elimination System Multi-Sector General Permit (MSGP) and the SWPPP.

To minimize the risk of noxious weed infestations or spread of weed seeds, equipment would be inspected and cleaned prior to mobilizing onto the Forest. All access routes, platforms, locations and sump construction sites also would be inspected prior to project-related activities and if they are found to be weed-infested, then the weed infestation would be treated by manually removing infestations using hand tools prior to ground disturbing activity. Herbicide use, where prescribed, would be in accordance with the South Fork Salmon River Sub Basin Noxious and Invasive Weed Management Program (USFS 2007). Infestations within 100 feet of live water would be controlled by hand pulling. Disposal of weeds also would be in accordance with the above plan.

For drill areas in RCAs:

- Pads will be sited to minimize the need to remove any large trees.
- Any tree that is felled will be left in the RCA.
- Silt fencing will be placed around pads and straw bales or wattles placed and staked between the stream and drill pad.
- When applicable, cross drains will be installed within the pad area to ensure drainage away from the RCA and stream.

Sightings of listed or sensitive fish and wildlife species will be reported to the USFS.

Employees and staff will receive training and direction to avoid spawning adult Chinook salmon, bull trout and steelhead.

If USFS administration of this project identifies unanticipated impacts to fish or fish habitat, the surface activity will be suspended by the Cascade or Krassel District Ranger until corrections can be made. A BNF or PNF fisheries biologist will be contacted and consultation will be reinitiated if necessary

If surveys or tracking of noxious weeds and/or rare plants occurs, this information will follow Forest Service protocol and be submitted to the Forest Service botanist and Weed coordinator.

For those drill pads within RCAs of stream channels, visual turbidity monitoring would occur immediately upstream and downstream of active drilling operations every 15 minutes. An annual report would be provided to the Level 1 Team that documents the results of visual observations. If operations are shown to be generating visible turbidity in a stream channel downstream of drilling that is greater than upstream levels, operations would cease until the source of sediment can be identified and mitigated. While actions are taken to stop the turbidity plume, the visual observations of upstream and downstream turbidity would be measured with a turbidity meter at 15-minute intervals until the downstream plume subsides. The USFS will be consulted before drilling resumes at sites B-1, B-2, B-4, B-5, B-7, and B-22. If a turbidity plume occurs due to drilling, the Level 1 Team will be promptly provided with a report that includes an account of the event, measures taken to stop the plume, and turbidity data.

Straw bales and wattles will be used to minimize mobilization of sediment from test pit excavated materials.

No toxic or hazardous substances will be used on site, except for standard petroleum fuel and lubricant products (diesel, gasoline, grease and hydraulic oils), and "over-the- counter" retail products. Use of all chemicals will be in accordance with manufacturer label.

Ground pressure reducing mats will be used when crossing areas delineated as having wetland characteristics. Test pit disturbance will be backfilled and replanted with certified weed-free seed mix.

The operator will immediately report any fuel, oil, or chemical discharges or spills greater than 5 gallons on land, or any spill directly in a stream to IDEQ and BNF/PNF as required by applicable federal and state regulations by phone and/or fax (or as soon as possible after on-site containment efforts are implemented as per the Spill Prevention, Control, and Countermeasure [SPCC] plan), and the BNF or PNF will initiate emergency consultation.

Concurrent reclamation will be conducted where possible and practical to offset potential erosion or sediment release, or as soon as possible.

The instantaneous diversion rate will always be less than or equal to 0.4 percent of the flow or less in Johnson Creek at the point of diversion and the total diversion from Johnson Creek will be less than 6,050 gallons (drilling plus dust abatement water). If drilling (and water use) occurs during the late August through September timeframe, USFS personnel will take flow measurements in Johnson Creek at the water withdrawal site.

**Standard Operating Procedures**

All USFS, county, and state speed limits, road restrictions, and load limits will be observed during travel. If appropriate, during equipment mobilization and demobilization, pilot cars will be used to ensure there are no conflicts or incidents along the narrow access roads leading into the project area.

Staff and contractors will follow speed limits. If significant dust generation is produced, vehicles will be requested to slow down to speeds necessary to minimize the fugitive dust generation or the route will be watered. Up to 2,000 gallons of water would be used for dust abatement.

At project completion, all equipment, supplies, and refuse will be removed from the project site and disposed of according to established solid and liquid waste management practices and applicable local, state and federal laws and regulations. Project activities will not generate materials regulated as "hazardous" or "toxic" waste with the exception of the handling of fuel-related products.

A standard marine-type fuel containment boom (which would be sufficient length for a worst-case scenario), spill prevention kit, and fire kit will be stored at the re-fueling site and will be readily available during off- loading of fuel from the fuel truck or during re-fueling operations.

A spill prevention and cleanup kit consisting of absorbent pads, absorbent booms (which would be sufficient length for a worst-case scenario), shovels and a fire extinguisher will be placed at the fuel storage site (private property), at the core shack (private property), and drill sites or any other areas where fuel and/or petroleum products are present.

After completing operations, all empty fuel and lubricant containers will be removed from the operations area and transported and disposed in accordance with local, state, and federal requirements.

Annual spill awareness/response training will be required for on-site personnel and suppliers/providers.

Two or more stored spill containment/response caches will be placed along each of the fuel delivery routes. Typically, fuel would be delivered to the drill rig in a 100-gallon doubled-wall tank mounted to a pick-up truck, tracked vehicle, or by helicopter.

Fuel containment sites, engines, and other equipment with fuel or lubricants would be periodically (during fueling activities or daily inspection of the drill rigs) checked for leakage or spillage and in accordance with the SPCC plan.

The SPCC plan would be kept at the core shack or office trailer. Staff handling fuel or petroleum products would be trained to successfully implement the SPCC plan. Inspections of fuel storage and handling areas would be conducted as specified in the SPCC plan. Appropriate warning signs would be placed around fuel storage facilities.

All contractors and company staff involved in handling oil and other chemicals would be made aware of the site SPCC plan, spill kit locations, and appropriate emergency response procedures, and would be required to abide by all applicable federal, state, and local laws and regulations pertaining to their respective operations.

Should any oil or chemical discharges or spills occur, the release would be reported to IDEQ, and the USFS (the USFS will contact NMFS and USFWS) and other appropriate agencies as required by applicable federal and state regulations by phone and/or fax immediately (or as soon as possible after on-site containment efforts are implemented as per the SPCC plan). Spill response would be in accordance with the SPCC plan, which includes a trained on-site emergency response team. Spills or discharges would be documented in writing.

Drilling mud and hole plug products would conform to American Petroleum Institute guidelines for ensuring groundwater integrity. Material Safety and Data Sheets for all products would be posted and available on site with the SPCC plan.

A fuel management plan has been created for the project that analyzes measures for minimizing the potential for fuel spills along the main routes into the activity area. The fuel management plan also outlines the times of year and the routes that would be used to deliver fuel into the project area. The fuel plan would be followed for all activities associated with fuel delivery.

# **Dynamic Cone Penetration Test**

The portable DCPT is used to determine underlying soil strength by measuring the penetration of the device into the soil after each hammer blow. The DCPT consists of two ½-in. diameter shafts coupled near the midpoint. The lower shaft contains an anvil and a 0.787-in. cone, which is driven into the soil by dropping a 17.6-pound sliding hammer contained on the upper shaft onto the anvil. The test involves raising and dropping the hammer to the drive cone on the lower shaft

and recording the amount of penetration observed. The DCPT equipment is manually operated, requires no specialized equipment for transport, will not occur in streams, and will not result in ground disturbance.

# **Test Pits**

A total of 14 test pits will be excavated, all located adjacent to Forest Service Road (FR) 447 (Figure 17). The distance of the pits to the closest streams ranges from approximately 50 ft. to approximately 1,680 ft. All of the sites that are within 200 ft. of streams are on the upslope side of FR 447, such that the road will be between the site and the stream. Excavated material will be temporarily placed next to the test pit. Straw bales and wattles will be used to minimize mobilization of sediment. After completion of the test, the excavated material will be placed back into the pit and tamped down with the excavator bucket. Each pit will be active for approximately one day. After the tests are complete, the disturbed ground will be re-planted with certified weed-free seed mixes that are appropriate for the sites.

# **Track/Truck Mounted Drilling**

Drilling will be conducted with track and truck mounted drilling rigs at seven sites (Figure 17). One site (B-3) is more than 500 ft. from the nearest stream, but the other six sites are within 50 ft. of streams. Of the six sites within 50 ft. of streams: two (B-1 and B-2) are at the FR 447 crossing of Johnson Creek; one (B-4) is at the FR 447 crossing of Burntlog Creek; one (B-5) is at the FR 447 crossing of East Fork Burntlog Creek; one (B-7) is at the FR 447 crossing of an unnamed tributary to East Fork Burntlog Creek; and one (B-22) is located on the EFSFSR approximately 3.8 miles upstream from the Yellow Pine Pit (YPP). Sites B-3 and B-22 are the only sites that are not on or immediately adjacent to FR 447, and site B-22 is the only site that is in a wetland area. Access to site B-3 will require a tracked drill rig to travel approximately 600 ft. off road and approximately 400 ft. on an existing closed road. Both the B-3 site and the access routes are entirely outside of riparian conservation areas (RCAs). Access to site B-22 will require a tracked drill rig to travel approximately 250 ft. on an existing closed road and approximately 200 ft. off road through a riparian wetland adjacent to the EFSFSR. Pressure reducing mats will be used at site B-22 to reduce the effects on wetland soils and vegetation. Accessing the B-22 site might require felling of a few small trees.

Drilling will commence with an auger and will transition to a core drill when bedrock is encountered. Standard penetration testing, consisting of driving a thick-walled sampling tube approximately 18 in. into the bottom of the borehole using a slide hammer, will be conducted every 2.5 ft. during augur drilling. When bedrock is encountered, drilling will transition to core drilling. The core drilling assembly will consist of auger flights, drill rods, water trough, water line, water pumps, tools, and ancillary equipment. Additional supplies will include drilling mud, bentonite (clay) hole plug material, and small amounts of biodegradable lubricants (rod grease) that are certified for use in potable drinking water. Drilling muds (if needed) will be recirculated using pumps and above ground troughs, precluding the need for excavating sumps. Auger drilling requires little or no water and core drilling requires approximately ten gallons of water per ft. drilled. Drill holes will be approximately eight in. in diameter and a maximum of 70 ft. deep.

Silt fences and straw bales will be used to minimize the chance of water or sediment exiting the site and reaching streams. Streams adjacent to the drilling sites will be visually monitored for turbidity plumes and drilling will cease if turbidity is detected. If drilling ceases due to detection of turbidity, it will not begin again until the source of the turbidity is identified and addressed, and BNF resource advisor agrees that it is safe to resume drilling.

After the tests are complete, each drill hole will be backfilled to within 3 ft. of the ground surface, capped with concrete from the top of the backfill to a few inches of below the ground surface, and monumented with a steel chain for survey purposes. Native material will be mounded on top of the concrete cap during final reclamation. If artesian or geothermal waters are encountered during drilling, the hole will be sealed with cement grout or quick-setting bentonite to prevent cross-flow and/or mixing of groundwater aquifers.

# **Helicopter Supported Drilling**

Helicopter supported drilling will be conducted at 14 sites (Figure 17). The closest helicopter supported drilling site to a stream is approximately 155 ft. from an unnamed tributary to Riordan Creek. All of the other sites are more than 200 ft. from streams. All drilling will utilize casing advancer/core drilling methods. Drilling equipment will consist of a portable drilling platform, drill rig, drill rods, water trough, water line, water pumps, fly fuel tank (i.e., fuel tank flown in with the helicopter), fuel for the drill rig, water storage tank, tools, and ancillary equipment. The platforms will be supported by four adjustable legs with the base of each leg being approximately one square ft. in area. The platforms will be flown in and set with a helicopter and all of the drilling equipment and supplies will be transported to and from the sites via helicopter. The drill platform will occupy an area of approximately  $400$  ft<sup>2</sup> at each site. Minor brush clearing and tree cutting may be required to clear the areas for the drill platforms and to provide a safety zone around the platform. Silt fences and straw wattles will be used to minimize chance of water or sediment exiting the site and the drill holes will be reclaimed as described for the auger/core drilling.

# **Water Diversion and Use**

Approximately 4,050 gallons of water will be needed for drilling and up to 2,000 gallons may be used for dust abatement. Water will be withdrawn from an existing well at the Stibnite Mine site and will be diverted from Johnson Creek at the FR 447 Bridge using a pump with a rating of up to 150 gallons per minute (0.33 cubic ft./second [cfs]). The pump intake will be screened to NMFS' criteria for mesh size and approach velocity and the screen will be inspected every four hours of pump operation and cleaned as needed. A maximum of 6,050 gallons will be diverted from Johnson Creek.

# **Vehicle Trips**

There will be approximately 145 light vehicle trips and approximately 16 heavy vehicle trips. Most of the travel will be via the Johnson Creek Road (FR 413), Stibnite Road (FR 412) and Burntlog Road (FR 447). However, heavy vehicles traveling to the southern portion of the study area will also use Warm Lake Road (FR 579) to minimize wear on Johnson Creek Road.

# 1.8.2.7. **BRGI Monitoring**

For those drill pads within RCAs of stream channels, visual observations for changes in turbidity will occur immediately upstream and downstream from active drilling operations. An annual report will be provided to the district fisheries biologist or hydrologist that documents the results of the visual observations. If operations are shown to generate a visible increase in turbidity in the water downstream from the drilling activity that is greater than upstream levels, operations will cease until the source of turbidity can be identified and mitigated. While actions are taken to stop the turbidity plume, the upstream and downstream turbidity will be measured with a turbidity meter at 15-minute intervals until the downstream plume subsides. The USFS will be consulted before drilling resumes at sites B-1, B-2, B-4, B-5, B-7, and B-22. If a noticeable increase in turbidity occurs due to drilling activities, the district fisheries biologist or hydrologist will be promptly provided with a report that includes an account of the event, measures taken to stop or reduce the turbidity due to drilling activities, and turbidity data (if collected).

Perpetua will continually monitor the condition of identified BMPs, ensuring they are in place, maintained, and effective. These measures are designed to ensure protection of water quality from increased turbidity or runoff.

### **1.8.3. Underground Exploration**

Underground exploration activities will occur at the newly discovered Scout Prospect, a 1-mile, downward-sloping tunnel (a decline) (see Figure 15). The decline will be used to reach the subsurface mineralized zone known as the Scout Prospect. The decline will be accessed from a portal facility known as the Scout Portal, located south of the planned ore processing facility (Figure 15). Approximately 100,000 tons of rock will be excavated from the decline. Exploration drill holes will be installed at various locations in the decline. Selected drill cuttings or core will be removed from underground for testing.

To construct the portal facility, the hillside will be cut into to develop a flat vertical slope using conventional underground drill and blast operations with mechanized equipment. Explosives will be used in the underground development process to construct the decline. The underground development rock could be used for surface pad construction, hauled to the ore stockpile area, or hauled for storage in the TSF Buttress as appropriate.

Drilling is used in advance of the decline to ensure unexpected or unmanageable water pressures are not intersected. Water will be used in underground drilling or pumped from the collection point to the surface. Upon reaching the surface, this water will be piped to the ore processing facility to be used in the plant.

#### **1.9. Closure and Reclamation**

#### **1.9.1. Overview**

Closure and reclamation at the site will include interim, concurrent, and final closure and reclamation in order to: (1) stabilize Project-related disturbances; (2) mitigate/compensate wetland loss directly related to Project development; (3) comply with applicable water quality standards; and (4) achieve long-term post-mining land uses. Details on reclamation activities to be implemented for the SGP, including appropriate seed mixes to be used are described in the Reclamation and Closure Plan Stibnite Gold Project (Tetra Tech 2021). Interim reclamation is intended to provide shorter-term stabilization to prevent erosion of disturbed areas and stockpiles that will be more fully and permanently reclaimed later.

Concurrent reclamation is designed to provide permanent, low‐maintenance achievement of final reclamation goals on completed portions of the site prior to the overall completion of mining activities throughout the SGP. Approximately 37 percent of the reclamation will be completed concurrent to mining and ore processing; remaining reclamation activities will be completed during closure.

Final closure and reclamation will involve: (1) removing all structures and facilities; (2) reclamation of those areas that have not been concurrently reclaimed such as the TSF and some backfill surfaces; (3) recontouring and improving drainages; (4) creation of wetlands; (5) reconstructing various stream channels; (6) decommissioning of the EFSFSR diversion tunnel; (7) GM placement; (8) planting and revegetation on disturbance areas; and (9) relocating Stibnite Road (FR 50-412) across the backfilled and closed YPP area.

Final reclamation of certain facilities could continue beyond the five-year closure and reclamation period. The Burntlog Route will be needed until the TSF is fully reclaimed, after which the newly constructed portions of the road will be decommissioned and reclaimed (i.e., fully obliterated), and the currently existing portions of the road will be returned to their prior use.

Surface water flow diversion of portions of the EFSFSR, Garnet Creek, Meadow Creek, Midnight Creek, and Hennessy Creek will be reclaimed and incorporated into constructed wetlands (i.e., Garnet Creek) or restored stream channels across the reclaimed TSF (i.e., Meadow Creek) or YPP backfill.

Perpetua will seed and plant stream reaches, riparian areas, and wetlands with native plant species present currently in existing wetlands and riparian areas along streams within the SGP footprint. Seed mixes, live stakes, and nursery-grown container plants and plugs of native graminoids, forbs, shrubs, and trees would be utilized for revegetation (Tetra Tech 2023). The revegetation plan has been developed for specific riparian, wetland, and upland zones to improve long-term bank stability, LWD recruitment, overhead cover, shade, and terrestrial/wetland habitat (Brown and Caldwell 2021e). In an effort to provide shade to action area streams, riparian plantings will be 18 feet wide, with a higher percentage of taller and denser species (e.g., spruce trees) than originally planned (Brown and Caldwell, Rio ASE, and BioAnalysts 2021). The riparian planting plan is described in Appendix F of the Stream Design Report, where local native seed, cuttings, and containerized materials will be used wherever feasible to increase the likelihood of survival (Rio ASE 2021).

Closure and reclamation activities will be intended to achieve post-mining land uses of wildlife and fisheries habitat and dispersed recreation at the SGP under current motorized access requirements and route designations. Dispersed recreation uses will be accessible by the relocated Stibnite Road (FR 50412) through the backfilled YPP that will facilitate recreational

traffic and access to Thunder Mountain. The proposed final reclaimed condition of the site is shown on Figure 18. Concurrent and final closure and reclamation for the SGP are described in greater detail in the following sections.



**Figure 18. Post Closure and Reclamation Condition.**

#### **1.9.2. Decommissioning, Demolition, and Disposal of Facilities**

Perpetua will dismantle or demolish structures and facilities not necessary for post-closure water management (e.g., certain culverts and pipelines). The materials from the dismantling or demolition of structures and facilities will be salvaged or disposed of in permitted off-site landfills. All reagents, petroleum products, solvents, and other hazardous or toxic materials will be removed from the site for reuse or will be disposed of according to applicable state and federal regulations. Concrete foundations will be broken or fractured as required to prevent excessive water retention and covered in-place with a minimum of 2 ft. of cover material (consisting of a minimum of 1.5 ft. of backfill and a minimum of 0.5 ft. GM) or will be broken up and buried in the TSF Buttress or pit backfill prior to installation of a geosynthetic liner cover. Soil/rock beneath fuel storage areas and chemical storage buildings will be tested for contamination and removed or disposed of appropriately if needed.

# **1.9.3. Underground Exploration and EFSFSR Tunnel**

Perpetua will decommission and close underground facilities and underground support facilities, including the portals of the EFSFSR tunnel and Scout decline. To prevent future access to underground workings, the underground portals (i.e., EFSFSR tunnel and Scout decline) will be closed using concrete block bulkheads, rockfills, or a combination of rockfill and lowpermeability foam. The downstream (north) EFSFSR portal and the Scout decline will be closed with bulkheads inside the portals (where overhead cover was at least 3 times the tunnel height) or backfilled with clean rockfill starting inside the portals and working outward, and up against the portal headwalls. Surface swales will be installed to direct surface water around the backfilled portal, and the exterior backfill, and surrounding disturbance will be graded to blend with adjacent topography, covered with GM, and revegetated. At the EFSFSR upstream (south) portal, the control weir will be left in place, and the fishway weir notch raised with concrete, creating an approximately 4-ft.-high sill to exclude river water or alluvial groundwater, and lowpermeability geofoam or similar will be installed inside the portal after the initial backfill or bulkhead, to prevent water entry. Then, the portal area will be filled, regraded, and revegetated as described for the other openings.

# **1.9.4. Yellow Pine Pit**

During MYs 5 through 11, the majority of the YPP backfill material (90 percent) will be WEP development rock. The balance of YPP backfill will include development rock from the HFP (5 percent) and the YPP (5 percent). Backfill will be placed in lifts not exceeding 100 ft. in vertical height with the large equipment, to include selective placement of the top lifts by direct dumping to better control the type of rock that will be placed near the surface. This placement method also will limit subsidence of the backfill and the amount of regrading necessary prior to placement of GM (Tetra Tech 2021). This material will not be compacted beyond that which occurs during placement, subsequent routing of trucks, burial, and consolidation. Portions of the highwalls on the east and west sides of the pit will remain above the backfilled portion of the pit and will not be reclaimed. A sinuous channel will be constructed through the backfilled area for the reconstructed EFSFSR with an average valley gradient approximating the historical, predisturbance river gradient (Tetra Tech 2023). A low permeability geosynthetic liner will be incorporated into the cover over the entire surface of the backfilled YPP, including the

reconstructed channel floodplain corridor to reduce the infiltration of meteoric water into backfill material, which could dewater the restored stream channel and result in additional metal leaching from the underlying backfill. Above the geosynthetic liner in the stream corridor, a layer of relatively fine material will be placed to protect the stream liner from puncture, followed by coarse rock armor to protect from exposure via stream scour, followed by floodplain alluvium at a minimum thickness equal to the maximum estimated scour depth of the proposed stream channel. GM will then be placed and the area revegetated. The lined corridor will be wide enough to accommodate future channel migration, evolution, and over-bank flooding. The cover system outside the stream/floodplain corridor will be similar to that described for the TSF Buttress (Section 1.9.6). Portions of Hennessy and Midnight Creeks will be restored over the backfilled area along with the reconstructed EFSFSR.

Hennessy Creek will cascade over the approximately 275-ft. tall west highwall of the YPP to a restored 0.17-mile section of low-gradient channel on the western edge of the reconstructed EFSFSR floodplain before joining the restored EFSFSR channel. Midnight Creek will be restored across the 0.14-mile southeastern portion of the reconstructed EFSFSR floodplain. After closure of the EFSFSR tunnel, backfilling of the YPP, and restoration of the EFSFSR and Hennessy Creek across the backfill, the Hennessy Creek diversion will be decommissioned and the area reclaimed, along with the adjacent operations-phase public access road.

To accommodate migrating fish, including Chinook salmon and bull trout, step pools will be established within the constructed EFSFSR channel consistent with NOAA 2022 fish passage guidelines (NMFS 2022a). The vertical relief (drop) between successive pools will not exceed published fish passage criteria. Detailed hydrologic and hydraulic analyses will inform the overall channel and floodplain design and construction, with channel bankfull width approximately 25 to 30 ft., and average depth of approximately 2 ft. The lined Stibnite Lake, of similar size to the existing YPP Lake, will be constructed within the lined corridor (Perpetua 2023; Tetra Tech 2023 [Attachment D of the CMP]).

Access through the site to Thunder Mountain Road (FR 50375) will utilize an access road through the backfilled area, replacing the segments of the Stibnite Road (FR 50412) that were removed by mining.

# **1.9.5. West End Pit**

This area includes the WEP, the Midnight Pit, the sidehill pit, and the development rock from legacy mining activity (Figure 4). Reclamation will occur at the conclusion of mining operations. The WEP will not be reclaimed; instead, a pit lake about 400 ft. deep will be allowed to form in the northern portion of the pit below the highwall, which will be about 800 ft. above the pit lake surface. The WEP lake will fill gradually up to 400-ft.-deep, and lake levels will fluctuate seasonally and with longer-term climate variations; however, the lake will not be expected to completely fill with water or spill due to its limited catchment area.

To account for model uncertainty, lake levels will be monitored after closure, as specified in the Environmental Monitoring and Management Plan (EMMP) (Brown and Caldwell 2021c), and a threshold water level will be established, sufficient to contain the predicted runoff volume from a high-snowpack year without discharge. If water levels approach the threshold, either or both

surface water diversion and water treatment could be implemented to prevent the lake from spilling. If needed, a temporary treatment unit will be mobilized to the site to treat and discharge the pit lake water until the lake level falls below the threshold discharge level, thus preventing untreated discharge in potential subsequent wet weather years and enabling gradual and predictable water treatment rather than treatment at higher but variable and uncertain peak spring runoff rates.

The Midnight Pit, the approximately 6-acre, 100-ft.-deep southeastern portion of the overall WEP within the Midnight Creek catchment, will be backfilled during operations with approximately 6 million tons of development rock from the WEP. The backfill will be placed to achieve a mounded final reclamation surface to promote drainage away from the WEP and prevent formation of a pit lake within Midnight Pit. Portions of the backfill will be covered with GM and revegetated, and the remainder covered with talus like development rock to mimic a natural talus slope.

The floor of the sidehill pit southwest of the main WEP will be graded to drain, covered with GM, and revegetated. No backfilling will occur for the main WEP. At closure, the remaining road into the pit and access to highwalls will be blocked with large boulders and/or earthen berms to deter motorized vehicle passage into the pit.

# **1.9.6. Tailings Storage Facility and TSF Buttress**

Perpetua proposes to complete tailings reclamation approximately 9 years after ore processing operations cease. After tailings consolidate sufficiently to use heavy equipment on top of the tailings, predicted to be within 3 to 5 years after the end of deposition, Perpetua will begin with placement of cover material, then construct wetlands and restore Meadow Creek and its tributaries within appropriately sized lined floodplain corridors, place GM, and revegetate the area.

Once ore processing operations have ceased, Perpetua will begin removing the remaining supernatant water pool and ongoing accumulation of meteoric water and consolidation water through a combination of spray evaporators (similar to snowmaking misters) operated within the TSF boundary and an active water treatment that meets IPDES discharge limits, followed by discharge to the EFSFSR or Meadow Creek. Removal of the remaining supernatant water from the TSF will allow the surficial layers of the tailings to dry and gain strength, which will allow equipment to operate on the tailings surface for grading and the placement of the geosynthetic liner, overlain by unconsolidated overburden and GM. Concave areas in the consolidated tailings surface will be filled to create suitable drainage conditions prior to liner and cover installation in the area designed to become restored stream channel. Cover placement and minor grading of tailings will occur, beginning within 3 to 5 years from the end of deposition, as portions of the TSF surface dry enough to allow equipment traffic, working inward from the facility perimeter. The cover material overlying the geosynthetic liner will be sourced from unconsolidated overburden or other appropriate material stored in a GMS on top of the adjacent TSF Buttress.

Perpetua will restore appropriately designed meandering stream channels (Meadow Creek and tributaries) within a stream and floodplain corridor across the top of the lined TSF surface (Rio ASE 2021). Pools and riffles will be constructed within the channel. Measures to create aquatic habitat will include side channels, oxbows, boulder clusters, root wads, and large woody debris (LWD). This will allow for the post-closure development of riparian habitat, convey water off the facility, and minimize potential interaction of surface water with the underlying tailings. Given the nature of the surface of the TSF, the constructed channel will have a shallow gradient.

Detailed hydrologic and hydraulic analyses will inform the overall channel and floodplain design, which will necessitate the construction of defined channels ranging from approximately 5 to 15 ft. in bankfull width, with average bankfull depth reaching approximately 2 ft. A connected floodplain up to 200 ft. wide will convey higher flows during a 100-year flood event.

Consolidation of the tailings will continue after cover placement and surface reclamation, at gradually declining rates, until approximately MY 40. To prevent tailings consolidation water from mixing with surface water on the cover, potentially leading to water quality impacts if discharged to streams, the consolidation water will be collected for treatment, using shallow wells and gravel or geosynthetic drains. Initially, collected flows will be routed to a WTP for treatment and discharge. Treatment will no longer be required after approximately MY 40, at which time the treatment facility will be decommissioned and the WTP site reclaimed.

Final slopes of the TSF Buttress will be variable, to blend with the surrounding terrain to the extent practicable, produce a permanent and stable landform, provide access for future maintenance on the TSF and buttress, and provide for non-erosive drainage across the reclaimed face of the buttress. Upon completion of final grading of the TSF Buttress, a low permeability geosynthetic cover will be placed over the facility, which will be designed to limit infiltration through the underlying development rock. The geosynthetic liner will be overlain by an inert soil/rock layer (non-PAG/metal leaching development rock, fill, or alluvium) and GM and revegetated. Similar to that for the TSF, a channel and floodplain corridor will be established for Meadow Creek across the top of the lined buttress. The channel will have a low gradient and wide floodplain across the top of the buttress, then drop more steeply to the valley floor near the south abutment. The steep channel segment will consist of a boulder chute that will flow through multiple energy-dissipating basins (one mid-slope and one at the toe of the TSF Buttress) before being discharged to a restored Meadow Creek on the valley bottom.

#### **1.9.7. Hangar Flats Pit**

In MYs 6 and 7, HFP will be backfilled up to the valley bottom elevation or slightly higher and no pit lake is anticipated. Following closure, the western pit highwall will remain exposed above the backfill area. The already-established Meadow Creek diversion channel and floodplain corridor will be retained around HFP as the final configuration, and the segment of Meadow Creek between the toe of the TSF Buttress and the entrance to the HFP diversion will be restored along with adjacent riparian wetlands. At closure, the entire surface of the backfilled HFP will be covered with a low permeability geosynthetic liner overlain with seed bank material to establish wetlands. Non-perennial drainages in adjacent upland areas will be routed to facilitate development of the wetland hydrology. Meadow Creek downstream of the HFP diversion, to the confluence with the EFSFSR, will be enhanced during mine operations with LWD, boulder cluster habitat structures, and riparian plantings.

#### **1.9.8. Transmission Line and Electrical Infrastructure**

The Johnson Creek and Stibnite substations will not be decommissioned immediately during mine closure; the transmission line between these substations will remain to provide power for post-closure water treatment. Once there is no longer a need for active water treatment, Perpetua, in coordination with IPC, will disassemble the approximately 9-mile transmission line between the Johnson Creek and Stibnite substations. The substations, switchgear, and power line will be removed. The transmission line ROW and associated access roads will be recontoured to match surrounding topography and revegetated. As part of revegetation, the transmission line structure pads and access roads will be scarified and revegetated. Revegetation will not be required on affected lands, or portions thereof, where planting is not practicable or reasonable because the soil is composed of excessive amounts of sand, gravel, shale, stone, or other material to such an extent to prohibit plant growth (IDAPA 20.02.02).

#### **1.9.9. Burntlog Route**

Once all final mine closure/reclamation work has been completed, Perpetua will reduce the 21 ft.-wide travel way of 19.8 miles of Burntlog Road (FR 447), 1.3 miles of Meadow Creek Lookout Road (FR 51290), and 2 miles along Thunder Mountain Road (FR 375) of the Burntlog Route to their approximate pre‐mining width. The public use status of these existing road segments will be unchanged from the current motor vehicle use map. Returning this 23 miles of existing road to pre-mining condition will entail grading and/or scarification along the outside edges of the road followed by seeding with the species listed in the Reclamation and Closure Plan (Tetra Tech 2021) or as approved by the USFS. Perpetua will remove ditches, cross drains, culverts, safety berms, mile markers, guardrails, and signs on roads if these features are no longer needed. These roads will retain the flatter grades and gentler curves constructed for mine operations.

The approximately 15 miles of Burntlog Route that was newly constructed for the SGP, connecting Burnt Log Road (FR 447) to Meadow Creek Lookout Road (FR 51290) and Thunder Mountain Road (FR 50375) will be fully decommissioned. The road will be decommissioned by pulling back and recontouring road cuts to slopes that are similar to, but not necessarily matching, pre-project conditions, and that will be consistent with the surrounding terrain as practicable. Surface water diversions, cross drains, culverts, safety berms, mile markers, guardrails, and signs will be removed. Soil nail walls, constructed of anchors bolted into the ground with a sprayed concrete surface, will remain to support slopes in areas with soft soils or weathered rock. Water bars or other erosion and sediment control structures, armored stream crossings, and stormwater crossings will be included where necessary. The reclaimed areas will be scarified, and 6 inches of GM will be placed in upland areas, followed by seeding and certified weed-free mulching on slopes over 30 percent. Revegetation will not be required where planting is not practicable or reasonable due to excessive amounts of sand, gravel, shale, stone, or other material to such an extent to prohibit plant growth (IDAPA 20.02.02).

### **1.9.10. Post Closure Public Access**

As mentioned in Section 1.9.4, a service road will be established over the backfilled YPP to allow public access through the reclaimed site and connect Stibnite Road (FR 50412) to Thunder Mountain Road (FR 50375) (Figure 19).

### **1.9.11. Off-site Facilities**

Following mine closure and reclamation, the Burntlog Maintenance Facility buildings will be removed. The sewer system and septic tanks for the Burntlog Maintenance Facility will be decommissioned. All reagents, petroleum products, solvents, and other hazardous or toxic materials will be removed from the site and disposed of according to applicable state and federal regulations. Soil/rock beneath fuel storage areas and chemical storage buildings will be tested for contamination and treated if necessary. After demolition of the buildings and facilities, the site will be graded, revegetated, and drainage restored.

# **1.9.12. Contouring, Grading, Growth Medium Placement, and Seeding**

Except for the HFP highwall above the valley bottom, the WEP, and a portion of the YPP highwall, Perpetua will contour and grade disturbed areas to blend into the surrounding topography and terrain. Compacted areas such as roads, ore stockpile areas, parking lots, fuel storage areas, and building sites will be prepared prior to placement of growth media and revegetation. Haul routes and access roads will be recontoured to establish natural drainage patterns.

Growth media suitability criteria include U.S. Department of Agriculture (USDA) texture, percentage of organic matter, course fragment percentage and acidity (pH). Root zone material suitability guidelines include USDA texture, course fragment percentage, soil acidity (pH), electroconductivity, sodium adsorption ratio, Net Acid Generation pH, bulk density and arsenic, antimony, and mercury levels. Perpetua will manufacture GM material using screened fines from glacial till sources, available mulched vegetation, and off-site composted material from private lands (e.g., composted food waste from the Worker Housing Facility). Off-site sources for composting feedstock materials will be in compliance with USFS requirements.

Planting, seeding, and mulching will be conducted in the fall and early winter to take advantage of snowpack and springtime moisture. Where cover crops are used in lieu of mulch, seeding will occur in the spring or fall followed by seeding of the permanent mixture. The forbs, grass species, seed amounts, and the trees and shrubs planned for planting on reclaimed areas are described in Tetra Tech (2021) and will be approved by the USFS.

# **1.9.13. Weed Treatment**

Perpetua will be responsible for noxious weed control within areas disturbed by proposed activities, and establishment and spread of noxious weeds and invasive plant species on Project areas will be prevented. Perpetua will inspect and remove vegetation material (including noxious weeds) from mechanical equipment and properly dispose of it to minimize the spread of unwanted vegetation. Additionally, Perpetua will limit preconstruction weed treatments, such as

mechanical control and herbicide application, to areas expected to have unavoidable grounddisturbing activities.

As identified below in the Reclamation Maintenance Procedures (Section 1.10.3.3), noxious weed and invasive plant species control will be conducted per USFS-approved methods described in the 2020 Programmatic Activities BA (PNF 2020) and corresponding Opinion from NMFS (WCR-2020-01560) (NMFS 2020). These documents are included in the proposed action and are incorporated by reference, although a brief summary of the approved active ingredients, surfactants, and treatment approaches follows.

The PNF approach to weed treatment allows for: (1) mechanical control, using hand-operated power tools to cut, clear (pull, grub or dig out), mow, or prune weeds; (2) biological control, allowing for release insects or pathogens that are parasitic and "host specific" to target weeds; (3) cultural control, allowing for implementing measures to prevent weed introduction or minimize the rate of weed spread (e.g., clean equipment before and after use, use only weed-free seed and mulch material, etc.); or (4) chemical control, allowing for the use of 14 active ingredients and specific adjuvants.

Active ingredients allowed for use can be found in Appendix C, Table 1; and the list of approved adjuvants are described in Appendix C, Table 2. Appendix C, Table 3 identifies mitigations and BMPs necessary for weed treatment. In addition to complying with these elements of the proposed action, the following key project design features for minimizing effects of herbicide treatment shall also be applied to the SGP:

- Herbicide applications will comply with applicable laws, policy, guidelines, and product label directions.
- Herbicide applications will comply with the buffer restrictions in Appendix C, Table C-1.
- Low pressure and larger droplet size will be used to the extent possible, to minimize herbicide drift during broadcast spraying.
- Mixing/filling will be limited to locations where drainage will not allow runoff or spills to move into live water and in locations where potential contamination of groundwater will not occur.
- No ester formulations of 2,4-D or triclopyr-butoxyethyl ester will be used.
- The polyoxyethylene tallow amine adjuvant will only be used in uplands where there is no potential for movement into aquatic systems.
- Invasive weed treatment with the potential for ground disturbance will be conducted only in areas where a 25-ft. vegetated buffer strip can be maintained. Proper erosion and sediment control BMPs will be employed.

# **1.9.14. Post Closure Water Treatment**

Evaluation of post-closure water treatment is ongoing. For the proposed action, Perpetua has indicated that sources of water that could require treatment during closure and reclamation and through the post-closure period include TSF runoff and tailings consolidation water, plus any TSF Buttress toe seepage. Other development rock will be backfilled into the open pits and closed with synthetic geotextiles, growth media, surface grading, and/or revegetation to preclude contact between the development rock and surface runoff.

As previously described, consolidation water will be withdrawn from beneath the TSF geosynthetic cover using a combination of wells, wicks, and/or gravel drains, and routed to water treatment. Collected flows will be routed to the water treatment plant for treatment and discharge. Once it is determined that treatment is no longer required based upon agency approvals, the treatment facility will be decommissioned and the WTP site reclaimed. Water treatment will be provided during the reclamation and closure and post-closure phases until waters requiring treatment are no longer being generated. Life-of-mine water treatment of the TSF and other facilities is discussed in Section 1.7.10.

As described in Section 1.9.5, if spillage of surface water from the WEP lake becomes imminent, a portable system will be brought to the site to treat and discharge pit lake water to maintain levels below the rim of the lake and prevent uncontrolled release of lake water.

### **1.9.15. Closure and Reclamation Financial Assurance**

As part of the approval for the SGP, Perpetua will be required to post financial assurance to ensure that NFS lands and resources involved with the mine operations are reclaimed in accordance with the approved plan of operations and reclamation requirements (36 CFR 228.8 and 228.13). This financial assurance will provide adequate funding to allow the USFS to complete reclamation and post-closure operation, including continuation of any post-closure water treatment, maintenance activities, and necessary monitoring for as long as required to return the site to a stable and acceptable condition in the event Perpetua was unable to do so. The amount of financial assurance will be determined in collaboration with the USFS and will "address all Forest Service costs that will be incurred in taking over operations because of operator default" (USFS 2004). The financial assurance will be required in a readily available bond or other instrument payable to the USFS. To ensure the bond can be adjusted as needed to reflect actual costs and inflation, there will be provisions allowing for periodic adjustments in the final plan of operations prior to approval. Calculation of the initial bond amount will be completed following the Record of Decision (ROD) when enough information is available to adequately and accurately perform the calculation. In addition to the USFS-required bond, mitigation under Section 404 of the CWA also requires financial assurance.

The IDL will require a bond as part of their permitting authority and IDEQ will require a bond for the cyanidation permit which will then be held by IDL. The IDWR is the state agency responsible for design review and approval of the TSF. IDWR also will require a bond so that the TSF can be placed in a safe maintenance-free condition upon decommissioning or if abandoned by the owner.

#### **1.9.16. Closure and Reclamation Traffic**

Most closure and reclamation traffic will occur May through November. Mine traffic during closure and reclamation is anticipated to result in a total AADT of 27, with 15 being from heavy vehicles and 12 being from light vehicles.

### **1.10. That Monitoring**

Monitoring will be conducted by the Project operator and reviewed by the USFS and other regulatory agencies to ensure compliance with permits and regulations and to manage the impact of the SGP on the environment. Air emissions, groundwater, surface water, aquatic, and other environmental parameters will be monitored during mine construction, operation, closure, and post-closure as described and specified in the EMMP (Brown and Caldwell 2021c). Authorizations from federal and state agencies include monitoring requirements for resources (e.g., air emissions, surface water, and groundwater) during mine construction, operation, closure and reclamation, and post-closure.

Monitoring will be conducted following the completion of closure and reclamation of all facilities and disturbance areas to demonstrate compliance with permit requirements and to measure the success of reclamation and mitigation. Final monitoring requirements and timelines will be outlined in the final permit approval documents and the final EMMP.

The final EMMP will consist of multiple component plans, each of which will be finalized upon issuance of the related permit(s) and will contain monitoring and management requirements from each permit. In some cases, if environmental outcomes may be uncertain, the EMMP will include adaptive management planning which requires identification of performance measures, impact thresholds, and operational adjustment options, all intended to achieve and demonstrate compliance with applicable permitting and/or consistency with the environmental analysis (see also Section 3.9.1).

The EMMP (Brown and Caldwell 2021c) describes the component monitoring and management plans that have been drafted and upon finalization, will be used by Perpetua to manage water resources, manage and monitor mine facilities, and monitor environmental resources. The EMMP includes environmental tasks and lists environmental permits, licenses, authorizations, and corresponding obligations. It also establishes Perpetua's commitments to environmental monitoring and management of mine facilities and environmental resources. The EMMP will provide direction to Perpetua to monitor its operations and environmental commitments, document permit compliance, and reduce potential impacts to environmental resources. Key monitoring requirements of the EMMP are described below.

# **1.10.1. Water Resources Monitoring.**

The Water Resources Monitoring Plan (WRMP) described the planned monitoring of surface and groundwater resources during the construction and operations phases of the SGP (Brown and Caldwell 2021c). Water resources monitoring includes five geographical areas within the Project area. The areas represent portions of the Project with internally similar Project-related activities, hydrology, and potential water quality concerns, i.e.:

- Northern Operations Area (YPP, WEP, Midnight Pit Backfill, EFSFSR diversion, and the confluence with Sugar Creek) where mining activities will expose mineralized rock materials and operations will modify groundwater levels and surface flows.
- Southern Operations Area (TSF, TSF Buttress, TSF surface diversions, HFP, Meadow Creek, and EFMC) where mining operations, development rock placement, and tailings

storage will expose mineralize rock and tailings materials and operations will modify groundwater levels and surface flows.

- Ore Processing Area (processing plant, WTP, truck shop, support facilities, and the EFSFSR below the Meadow Creek confluence) where operations will receive, store, and utilize fuels and reagents (e.g., acids, cyanide) and water treatment discharge will occur.
- Worker Housing Area (employee housing, sanitary water treatment plant) where sanitary water treatment will occur.
- Off-site Facilities (SGLF and Burntlog Maintenance Facility) where operations will store and transfer fuels and reagents for delivery to the on-site areas.

Surface water monitoring locations are described in Table 21 and are shown in Figures 19 through 22. Groundwater monitoring wells are described in Table 3.8-2 of the BA (Stantec 2024) and shown in Figures 4-4, 4-6, and 4-24 of the WRMP (Brown and Caldwell 2021c) and are incorporated here by reference. Most monitoring locations are situated downstream or downgradient from Project components with the potential to contribute constituents to nearby water resources with the remaining locations upstream or upgradient from Project components to characterize baseline conditions influent to the area. Measurements and analyte lists are based on the constituents potentially associated with the Project component and Project activities, i.e.:

- Open pits and haulage areas are focused on potential contributions of sediment and nitrogen species associated with blasting operations.
- Tailings storage, stockpiles, and development rock storage areas are focused on the potential to contribute sediment, major ions (hardness), metals and cyanide associated with leaching of mined materials.
- The processing area is focused on the potential to contribute cyanide, metals, and major ions (hardness) associated with the cyanide ore processing.
- Treatment plant outfalls are focused on collection of monitoring data for compliance with water treatment objectives (i.e., major ions, hardness, metals, temperature, turbidity, and continuous flow measurement).
- The sanitary treatment plant outfall is focused on collection of monitoring data for compliance with sanitary treatment objectives (i.e., *E. coli*, biological oxygen demand, temperature, turbidity, and continuous flow measurement).
- MSGP Stormwater Permits stormwater discharge locations are focused on monitoring data for compliance with those permits (i.e., turbidity, temperature, and metals).

The monitoring locations and parameters identified in Table 21 are proposed at this time and may change to reflect requirements by federal and state agencies through their permitting processes. Sampling frequency (and subsequent reporting frequency) will vary by permit and by monitoring location and parameter within permits. At this time, the frequency of sampling is not known for the monitoring proposed in the WRMP, but some parameters are more likely to be required to have more frequent monitoring (continuous or weekly) at some locations, such as flow, temperature, and pH. Continuous measurements are expected to be required for effluent flow for IPDES outfalls. Other physical parameters, such as temperature, may also be required to be measured continuously using a sonde at outfalls and downstream locations. A few parameters, if required, may only be monitored on an annual basis at a limited number of locations, such as

WET (Whole Effluent Toxicity) or nutrient loading. Most other parameters are likely to be monitored on a quarterly or monthly basis (Brown and Caldwell 2021c).

Each final federal or state permit/regulatory document will specify the terms of monitoring requirements, including the sampling locations, sampling frequency, sample type, water quality parameters, and other details such as laboratory methods and total versus dissolved metals. As stated on page 4-1 of the WRMP, "When all resource monitoring requirements are known, this plan will be updated to ensure that all required monitoring information is obtained." Surface and groundwater monitoring for closure and post-closure are not addressed at this time, but will be addressed in an update to the WRMP toward the end of the operations period based on site conditions at that time.



# **Table 21. Surface Water Monitoring Locations (listed from downstream to upstream by area).**


















**Figure 19. Northern Operations Area Surface Water Monitoring Locations.**



**Figure 20. Ore Processing Area Surface Water Monitoring Locations.**



**Figure 21. Southern Operations Area Surface Water Monitoring Locations During Operations.**



**Figure 22. Worker Housing Area Surface Water Monitoring Locations During Operations.**

Instantaneous flow and temperature measurements will be collected at each surface water location for each sampling. The USFS will review monitoring results for comparison to USFS requirements, the approved mine plan, modeling forecasts, and water quality standards.

The USFS will also require the following mitigation measures regarding water resources monitoring and monitoring results as follows.

### *1.10.1.1. Monitoring Measure - WRMP Implementation, Water Quantity*

Because construction, operation, and closure of the SGP has the potential to impact surface or groundwater resources, a focused Water Resources Monitoring Plan (WRMP) will be implemented by Perpetua. As the mine owner/operator, Perpetua will be responsible for the implementation of the plan focused on confirming the predicted groundwater drawdown within allowance for model uncertainty and its relationship to discharges at proximal surface water resources. The plan will include surface water, groundwater, and meteorological monitoring requirements for the Project. Water quantity measurements will include diversion rates from groundwater pumping, water levels in groundwater monitoring wells and piezometers located within the Operations Area Boundary, and flow rates of streams and springs at USGS monitoring stations as well as spring locations characterized in the baseline program within the predicted 10 foot drawdown contour. Monitoring results will be provided to the USFS on a quarterly basis and summarized in an annual report. The mine owner/operator will be responsible for continued monitoring and reporting of changes in groundwater levels and surface water flows prior to, and during, operation and for a period of time in the post-reclamation period. The Plan will be reviewed and approved by USFS and implemented prior to the commencement of mining. State authorizations may also have monitoring requirements and these requirements along with monitoring already conducted or proposed could be applied to satisfy the needs of this mitigation measure.

### *1.10.1.2. Mitigation Measure – Groundwater Discharge to Surface Water*

Impacts to groundwater discharge to surface water resources are predicted by the numerical groundwater flow model and are accounted for in the authorization. However, if monitoring results indicate a different nature or extent of impacts that are outside of model uncertainty and associated with Project water management, additional compensatory mitigation will be implemented to mitigate for the effects of that reduced flow on the use of the affected surface water resource. Any additional compensatory mitigation will need to be performed in compliance with applicable regulations, including 404 CWA permitting and 401 Water Quality Certification.

### *1.10.1.3. Monitoring Measure – WRMP Implementation, Water Quality*

Because construction, operation, and closure of the proposed Project has potential to impact surface or groundwater resources, a focused WRMP for the Project will be developed by Perpetua. As the mine owner/operator, Perpetua will be responsible for the implementation of the plan for the Project incorporating the confirmation of predicted surface water and groundwater

chemistry plus surface water temperature. The plan will include mined development rock and ore, surface water, groundwater, and meteorological monitoring requirements. Monitoring results will be provided to the USFS on a quarterly basis and summarized in an annual report. Perpetua will be responsible for continued monitoring and reporting of surface and groundwater chemistry and temperature prior to, during, and after operations for a period of time in the post-reclamation period. The plan will be reviewed and approved by the USFS and implemented prior to the commencement of mining. State authorizations may also have monitoring requirements and these requirements along with monitoring already conducted or proposed could be applied to satisfy the needs of this measure.

# *1.10.1.4. Monitoring Measure – Higher Frequency Water Quality Sampling and Analyses*

In scenarios where there is a demonstrated reason for concern that water sources and discharges around Project components could have rapidly changing analyte concentrations, water quality samples will be collected and analyzed more frequently than the regular monitoring program frequency for key parameters for a limited time until monitoring parameters stabilized (e.g., weekly sampling compared to monthly or quarterly sampling). The higher frequency data collected, which may coincide with requirements under other state and federal permits, will be reviewed and compared to previously collected data, baseline concentrations, and other permit conditions. Higher frequency water quality sampling and analyses will be applied to:

- Discharges from the start-up or resumption of mine water treatment plants following an extended shut-down (pH, specific conductivity, weak acid dissociable [WAD] cyanide, organic carbon, arsenic, antimony, and mercury) until results meet IPDES permit limits or the results of monitoring are considered sufficient based on USFS review in consultation with applicable state regulatory agencies.
- Monitoring of spill indicators in affected receiving monitoring wells, contact water collection ponds, and surface waters (pH, specific conductivity, spilled material indicators) until the results of the monitoring are considered sufficient based on USFS review in consultation with applicable state regulatory agencies.

# *1.10.1.5. Mitigation Measure – Contingency Plan for Long-term Power Interruption*

While IPDES permitting requires contingency planning for power interruption associated with discharging water treatment plants, other water management activities not associated with discharges are not required to have contingency plans under that permit. Perpetua will develop and maintain water management contingency plans for a long-term power interruption of longer than 24 hours to prevent unauthorized discharge from the following water management facilities: (1) contact water collection ponds; (2) TSF; (3) dewatering; (4) process plant water containments; and (5) any water pumping associated with a spill response.

# *1.10.1.6. Monitoring Measure - Updated Geochemical and Temperature Modeling*

Geochemical modeling and/or temperature modeling will be updated as necessary (at the request of the USFS) if monitoring results obtained from the WRMP or other data collection indicate a change in water quality conditions that will significantly influence prediction and recognition of

potential mine impacts. The USFS' review of quarterly and annual monitoring results compared to predicted conditions will provide early warning of potentially unanticipated, undesirable impacts to water resources to allow for implementation of appropriate mitigation measures. Implementation of these mitigation measures will reduce or eliminate potential impacts to water quality.

## *1.10.1.7. Mitigation Measure - Contingent Stream Temperature Reduction Measures*

Due to inherent limitations in modeling and forecasting stream flow temperatures over a multidecade period, effectiveness of the actual performance of TSF consolidation, stream channel restoration, riparian plantings, and other temperature reduction measures implemented may differ from forecast.

Ditches and pipelines utilized to divert water around the TSF during operations are expected to result in maintaining cooler water temperatures for downstream reintroduction into the mainstream system. In addition, these diversions will not be affected by TSF consolidation or implementation of stream channel restoration. Therefore, these surface flow diversions will not be removed/reclaimed and continue to be utilized to divert flows in part of in whole until:

- TSF consolidation appropriate for stream channel restoration could be verified via consolidation monitoring and re-modeling for the as-built tailings facility;
- Stream restoration design and implementation could be reassessed prior to construction by resurveying the as-built and partially consolidated TSF surface to determine whether design stream gradients could be achieved or whether the stream channel design will need adjustment to accommodate the gradients of the post-consolidation TSF surface; and,
- Achievement of design shading effects of riparian plants on stream temperatures could be reassessed prior to construction by measuring the success of establishing riparian plantings at locations outside the TSF footprint (e.g., HFP diversion corridor, TSF Buttress, across the YPP backfill or others) or a TSF-analogous test plot location utilizing the design cover materials and thicknesses.

Operational period maintenance practices for the diversions will remain in effect into the closure and post-closure period to prevent sedimentation and other factors from impairing the effective use of the diversions. Upon verification of the items above with any associated design adjustments, stream water temperature monitoring data in the constructed restored stream channel will be collected to confirm the performance of the temperature reduction measures. In an event where monitoring data indicated that acceptable stream temperatures will not be attained, the ditch and pipeline diversions will be recommissioned and utilized to convey surface flows in part or in whole until an effective planting design will be developed and implemented.

### *1.10.1.8. Mitigation Measure – Streamflow Temperature Adjustment*

In the event that riparian shading does not provide sufficient shade to maintain Summer Maximum Weekly Maximum Temperature (MWMT) at or below those included in the closure plan, adaptive management in the areas of concern will be used to identify the issues and

implement improvement measures. Depending on the degree and spatial extent of the mitigation needed, these measures could include supplemental plantings with larger, container plants along stream reaches, leaving low-flow diversion pipes in place for longer periods while vegetation is established, installation of temporary shade structures, storing and covering snowpack along reaches to allow melt water into the system, or retrofitting additional pond features for mixing day and night time flows to lower maximum daily stream temperatures.

In order for the USACE to issue a permit under Section 404 of the CWA and authorize dredge or fill placement in WOTUS, all unavoidable impacts to jurisdictional WOTUS must be mitigated. Perpetua proposes to accomplish compensatory mitigation for impacts to wetlands through a combination of mitigation bank credits in the North Fork Payette subbasin<sup>5</sup> and permitteeresponsible on-site mitigation within the SFSR subbasin, plus some additional off-site mitigation outside the SFSR subbasin to account for temporal impacts (Tetra Tech 2023).

The Project includes activities that will result in permanent impacts to WOTUS including wetlands. Perpetua submitted a draft CMP (Tetra Tech 2023) that addresses compensation for lost wetland areas and functions, in addition to addressing mitigation proposed for impacted streams, to the USACE for approval on April 7, 2023. The CMP addresses compensatory mitigation for permanent impacts that will be accomplished through a combination of mitigation bank credits and the creation of new wetlands, streams, and enhancing and reclaiming existing wetlands and streams in the general vicinity of the impact areas. The CMP also addresses compensatory mitigation to reduce the temporal loss of aquatic functions and potential risks associated with actions described in the CMP. Temporal loss of functions and values is discussed further below.

The CMP describes an accounting process for tracking the various wetland impacts (losses) and associated wetland mitigation (gains), in addition to streams. The CMP uses the Montana Wetland Assessment Method (MWAM) functional assessment tool to determine functional units for each affected wetland assessment area. These units are based on a combination of MWAM scores and acres of wetlands. When these functional units will be lost due to development in the associated wetland, those losses are considered "debits." Conversely, the creation of new wetlands can result in "credits" by assessing and estimating the predicted functional scores and area of proposed wetlands that will be created, restored, or enhanced. Using this system of accounting for wetland credits and debits, the CMP provides a ledger that itemizes debits throughout the construction and operating phases and proposed credits for conceptual wetland creation actions. This system of accounting for losses and compensatory gains is intended to demonstrate a means of ensuring that adequate mitigation will be provided regardless of the final impact area. The ledger can be scaled up or down to identify the appropriate wetland credits needed to compensate for the final determination of wetland debits, which will be documented in the CWA 404 permit. The ledger system also provides a way to track and assess temporal effects which are the effects that come from the loss of wetland functions during the period between impacts and compensatory mitigation.

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<sup>&</sup>lt;sup>5</sup> Mitigation proposed for the North Fork Payette River subbasin is not described or discussed in detail in this opinion as it is a subbasin not occupied by ESA-listed Chinook salmon and steelhead.

Based on the CMP ledger of debits and credits, the amount of time associated with the temporal impacts related to wetlands is approximately 20 years, during which time as many as 620 functional units are outstanding (Tetra Tech 2021; 2023). These temporal effects will only occur within the Salmon River drainage. Coordination with the USACE for approval of existing and predicted wetland functional assessment scores is ongoing. The USACE may have changes to the methodology for the functional assessment evaluation which could result in changes in the final CMP to ensure mitigation reflects any resulting changes. Final impact acreages will be determined as part of the CWA Section 404 permit application process and will be agreed upon by the USACE.

The CMP describes a plan to locate the compensatory wetland mitigation sites within the same subbasins as the associated wetland impact sites. Temporal lag between effects on stream functional units and their mitigation will be addressed via off-site stream improvements located in subbasins outside the Project vicinity (Tetra Tech 2023). The proposed compensatory wetland mitigation within the Project area subbasin will be located around the mine site area where the majority of wetland impacts will occur, with no mitigation sites proposed along the access roads and the transmission line routes. The current location and configuration of mitigation sites identified in the CMP were selected based on suitable hydrology and compatibility with watershed-scale features and on the likelihood that compensatory mitigation wetlands will be sustainable within five years (Tetra Tech 2023). At the conclusion of the USFS process, final wetland impacts will be assessed, any agreed upon off-site compensatory mitigation projects will be finalized, and a final mitigation plan will be prepared, including a final assessment of functional units lost and created, and then the final credits/debits will be documented in the CWA Section 404 permit.

Monitoring associated with implementation of the CMP includes:

- Annual monitoring of stream restoration will be conducted during the low-flow period of the first five years following completion of the restoration actions with the understanding that attainment of performance standards may take a longer monitoring period. Monitoring activities and reporting will vary by year for the five-year period as follows:
	- o During the first year following construction, the restored stream reaches will be assessed to determine if they had been constructed within design tolerances with any changes between design and construction noted along with the reasons for that change in an As-Built Report as part of the first annual monitoring report.
	- o During the first, third, and fifth years following restoration, physical channel conditions and aquatic habitat will be monitored by measuring stream channel characteristics including bankfull width, maximum pool depth, mean pool depth, bankfull width/depth ratio (W:D), and average channel depth along with three to five perpendicular transects per stream reach coinciding with riffle and pool features. In addition, to document bedform complexity, average channel slope and pool frequency/depth, a longitudinal profile of the channel thalweg will be conducted for a representative sub-reach. During each of the five years, streambanks will be visually inspected for eroding banks and to determine LWD and rock structures remain intact.
- o Riparian vegetation will also be surveyed to characterize the number of different native herbaceous, shrub, and tree species within the stream restoration reach (years 1 through 5), the percent cover of native herbaceous, shrub, and tree species observed within the restoration reach (years 2 through 5), and the percent cover of Idaho-listed noxious weed species (years 1 through 5).
- Following the fifth year of monitoring, a stream functional assessment for restoration design reaches will be completed for comparison to baseline stream functional assessments to determine whether restoration design reaches are functioning appropriately. The stream functional assessment will utilize USFS Watershed Condition Indicators (WCI), including the following parameters:
	- o water temperature;
	- o fine sediment;
	- o chemical contaminants;
	- o physical barriers located downstream;
	- o substrate embeddedness;
	- o LWD frequency;
	- o LWD recruitment potential;
	- o pool frequency;
	- o pool quality;
	- o off-channel habitat with cover;
- o W:D;
- o Streambank condition;
- o floodplain connectivity;
- o Change in peak/base flows;
- o change in drainage network;
- o road density and location;
- o RCAs;
- o disturbance history;
- o occupancy potential;
- o critical habitat;
- o fish use presence/absence.
- Annual monitoring of wetland restoration areas will be conducted for the first five years following completion of restoration with the understanding that attainment of performance standards, primarily for palustrine forested (PFO) wetlands may take a longer monitoring period. During the first annual monitoring, the wetland mitigation area will be assessed to determine whether site grading achieved design elevations for adequate hydrology to sustain plant communities with subsequent monitoring assessing whether restored areas are receiving adequate flow to support wetland vegetation. Monitoring will employ shallow piezometers, well points, and/or 12-inch test pits along defined monitoring transects to examine subsurface conditions in addition to surface observations of water marks, drift lines, sediment deposits, and drainage periods. Wetland vegetation will also be surveyed to characterize the number of different native submerged aquatic, herbaceous, shrub, and tree species within the restored palustrine aquatic bed (PAB), palustrine emergent marsh (PEM), palustrine scrub-shrub (PSS), and PFO wetland classifications (years 1 through 5), the percent cover of native submerged aquatic, herbaceous, shrub, and tree species observed within the restoration reach (years 2 through 5), and the percent cover of Idaho-listed noxious weed species (years 1 through 5). Soil monitoring in conjunction with vegetation monitoring along established transections will assess development of organic wetland soil by examining organic matter on soil surfaces, monitoring soil profiles to determine whether soil color meets USACE hydric soil indicated criteria (i.e., chroma less than two on the Munsell color chart) and assessing development of redoximorphic features by checking for oxidized and reduced soil zones.
- Following the fifth year of monitoring, the MWAM will be used to conduct a functional assessment of restored wetlands and the results will be compared to the results of the functions and values assessments performed prior to project construction.
- Annual monitoring reports will be submitted for stream and wetland restoration monitoring results. The first annual monitoring report will include as-built drawings of each stream and wetland restoration project completed describing site condition, topography, and planted areas, site dimensions, and water supply/water control features. Any deviations from the design will be documented. Each annual report will present monitoring results organized into the following sections: Monitoring Requirements and Performance Standards, Summary Data, Maps and Plans, and Conclusions.

Performance standards for water resources include:

- Comparison of water quality constituent concentrations to regulatory criteria;
- Comparison of predicted water quality constituent concentrations to observed conditions with further action per the project mitigation measures;
- Comparison of predicted project effects on groundwater levels and stream flows to observed conditions with further action per the project mitigation measures;
- Comparison of predicted stream temperatures to observed conditions with further action per the Project mitigation measures; and,
- Functional assessment of restored streams and wetlands compared to CMP targets for functional replacement.

# **1.10.2. Fisheries and Aquatic Resources**

Perpetua has submitted a draft FMP (Brown and Caldwell, Rio ASE, and BioAnalysts 2021) describing the proposed monitoring of instream and riparian habitat conditions and instream biological communities. The purpose of the monitoring is to confirm proper project implementation and to document physical and biological characteristics to assess whether they are achieving the desired effectiveness. Tables 7-2 and 7-3 in the FMP summarize the monitoring proposed by Perpetua, and are incorporated here by reference. A brief summary of the habitat and biological monitoring is summarized below.

# **Stream and Riparian Habitat Monitoring.**

Monitoring of stream restoration would occur annually in the low-flow period for the first 5 years following completion of restoration actions. Assessment of stream design for the restoration reaches would occur in the first year of monitoring after implementation and physical channel conditions, aquatic habitat, and riparian vegetation would be monitored annually for 5 years.

According to the Compensatory Stream and Wetland Mitigation Plan (Tetra Tech 2023), riparian vegetation performance standards include:

• Year 2: Cover of native herbaceous species at least 50 percent and cover of native woody species at least 10 percent.

- Year 3: Cover of native herbaceous species at least 60 percent and cover of native woody species at least 25 percent.
- Year 5: Cover of native herbaceous species at least 80 percent and cover of native woody species at least 50 percent.
- Years 1-5: At least five native herbaceous species and 4 native shrub or tree species present in riparian buffer.
- Years 1-5: Less than 10 percent cover of Idaho-listed noxious weed species.

#### **Biological Community Monitoring.**

Both fish and macroinvertebrate surveys would be performed annually to provide information about the status and trends in biological responses during mining and restoration as well as providing indicators of aquatic community health that contribute to effectiveness monitoring. Monitoring will follow protocols established and carried out during baseline monitoring studies (MWH 2017; Stantec 2019; Stantec 2020; and Watershed Solutions 2021).

Perpetua will continue to monitor aquatic communities on an annual basis (or as permit conditions dictate) during construction, closure, and restoration and enhancement phases of the project; however, Perpetua will periodically review the monitoring results and determine if the frequency and/or spatial scale of monitoring may be reduced or expanded in the future. Any revisions to the program would be discussed with appropriate stakeholders prior to implementation. Perpetua proposes to use snorkeling methodologies for monitoring.

Snorkel surveys would be performed at a subset of the 30 survey locations monitoring between 2012 and 2018 (subset locations yet to be determined). During operations, Perpetua will monitor fish use of the EFSFSR tunnel by video and with PIT-tag arrays placed at appropriate locations in and near the tunnel. Chinook salmon redd and carcass surveys will be performed annually according to Copeland et al. (2019) within the watershed upstream from the fishway. Surveys will occur at least once every 10 to 12 days during the spawning season and will include all stream reaches where there is suitable spawning habitat.

Stream macroinvertebrate community composition will be monitored annually throughout construction, closure, and restoration phases of the Project as an indicator of water quality and overall stream habitat condition. Monitoring will occur at a subset of the 11 locations previously surveyed within Meadow Creek, EFSFSR, Sugar Creek, and Tamarack Creek. The metrics and indices of the macroinvertebrate community that will be used to assess conditions are listed below and are based on those used in the baseline aquatic biological surveys (MWH 2017).

Total Taxa Richness Ephemeroptera Taxa Richness Plecoptera Taxa Richness Percent Plecoptera Trichoptera Taxa Richness Hilsenhoff Biotic Index Percent 5 Dominant Taxa Scraper Taxa Richness Clinger Taxa Richness

Long-lived Taxa Richness Metals Tolerance Index Intolerant Taxa Richness Percent Tolerant Individuals Shannon-Weaver H' (log e) Percent Filterers Percent Gatherers Percent Predators Percent Scrapers

Percent Shredders Percent Unclassified Filterer Richness Gatherer Richness Predator Richness

Scraper Richness Unclassified Richness PIBO Observed/Expected Index

Perpetua will prepare annual summary reports of the stream habitat and biological monitoring for submittal to the USFS, permitting agencies, and other stakeholders. The specifics of monitoring reports and their details remain to be further refined.

### **1.10.3. Dust Monitoring**

In addition to air quality monitoring requirements associated with the IDEQ's Air Quality Permit to Construct, the USFS will require off-site dust monitoring to determine the effectiveness of dust control measures in protecting USFS vegetation and visual resources that include the following:

### **Monitoring Measure - Fence-Line Dust Control Monitoring Plan Implementation.**

Because dust emissions from the Project may impact air quality, a dust monitoring plan was developed by Perpetua. As the mine owner/operator, Perpetua will be responsible for the implementation of the dust monitoring plan, including installation of dust monitors at two locations near the mine operations boundary. One location will be south of the mine boundary close to the Burntlog Route. The other location will be between the eastern mine boundary and wilderness areas. The plan will include dust and meteorological monitoring during operations and quarterly reports to the USFS; monitoring and reporting will occur during non-winter periods and be implemented prior to commencement of mining. After five years of monitoring and every three years thereafter, the USFS and Perpetua will review this plan to determine if sufficient information was acquired and the monitoring may be removed.

### **1.10.4. Reclamation Monitoring**

Prior to reclamation monitoring and maintenance programs, the USFS and IDL will agree to specific quantitative and qualitative reclamation monitoring plans and standards. Reclamation monitoring will begin during concurrent reclamation at SGP facilities. Quantitative and qualitative monitoring of reclamation success will begin the first growing season after concurrent or final reclamation is completed and will continue until success criteria are satisfied. The Reclamation and Closure Plan (RCP) (Tetra Tech 2021) presents the quantitative and qualitative reclamation monitoring that will be conducted and the performance standards that will be used (with USFS and IDL approval) to determine when maintenance activities are necessary, or reclamation is complete. These monitoring requirements are summarized below.

### **Erosion and Sediment Control Monitoring**

Soil stability will be estimated for all reclaimed areas using qualitative descriptors examining soil movement, surface rock, pedestaling, flow patterns, and rilling/gullying. A reclamation specialist will observe each reclaimed area and assign qualitative descriptors. Soil stability

monitoring will be completed twice annually for erosion control purposes, once in the spring and once in the fall, during the period when reclamation activities are being implemented. Once reclamation activities are completed soil stability observations will be made as part of performance monitoring after three years and will recur every three years until stabilization objectives have been met. For performance monitoring, the observations will be made at the same time the vegetation success observations are made. The monitoring results will be used to aid in determining the cause of any failures that are encountered and to locate problem areas before erosion becomes widespread enough to affect reclamation success.

# **Slope Stability Monitoring**

Slope stability will be monitored during the erosion observations. Qualified staff will look for signs of slope movement, cut slope and rock face failures, and other indications of slope instability. The location and dimensions of significant surface cracks and fill slope bulges will be monitored. This information will be used to determine if surface cracks are the result of differential settling of fill material or slope instability. The appropriate regulatory agency will be notified, and corrective plans will be developed.

### **Reclamation Maintenance Procedures**

Details pertaining to reclamation maintenance practices appear in Appendix A, Table A-6. To maintain normal conditions for reclaimed areas per their designs and/or if the performance of reclaimed areas is not satisfactory, appropriate maintenance activities will be implemented. Maintenance activities may include one or more of the following:

- Sediment removal from sediment basins, stormwater drainage channels, and diversions as necessary to maintain their design capacity;
- Diverting surface water away from reclaimed areas where erosion jeopardizes attainment of reclamation standards;
- Stabilizing rills, gullies, and other erosion features or slope failures that have exposed development rock;
- Noxious weed and invasive plant species control (per the USFS-approved methods and the 2020 Programmatic Activities Biological Opinion); and,
- Re-seeding or re-applying reclamation treatments in areas where it is determined through monitoring and agency consultation that reclamation will not meet standards.

### **Annual Report**

Perpetua will submit an annual report to the USFS and the other federal and state agencies that are responsible for issuing authorizations applicable to reclamation for the preceding calendar year. The annual report will contain descriptions of the reclamation activities completed during the previous year, a summary of areas reclaimed, a discussion of the results of the reclamation monitoring conducted, and corrective actions implemented.

Perpetua's proposed performance standards for reclamation include:

- Physical stability of reclaimed facilities free from erosion features that will affect revegetation and risks for mass slope failures;
- Comparison of predicted revegetation of riparian shade areas to observed conditions and the shading effects on stream temperatures with further action per the project mitigation measures; and,
- Achievement of 70 percent of the pre-existing vegetation cover for general revegetation areas (i.e., areas not associated with the establishment of riparian shading to achieve target stream temperatures).

## **1.11. Environmental Design Features**

The SGP must comply with all laws and regulations that apply to the proposed activities with prominent requirements relevant to the effect's analysis summarized in Table 22. This table has been modified in this opinion to only include EDFs relevant to minimizing or avoiding adverse effects to ESA-listed salmonids. For the complete table, please refer to BA Table 3.9-1. Standards and guidelines in the Boise and Payette Forest Plans (USFS 2003; 2010a) that are designed to reduce or prevent undesirable impacts resulting from proposed management activities are incorporated into the Proposed Action by reference. In addition, BMPs outlined in the Best Management Practices for Mining in Idaho (IDL 1992) will be implemented where appropriate and applicable for operations to minimize site disturbance from mining and drilling activities.

Perpetua will implement the mitigation measures/EDFs described in this section. Based on the application of permits and regulatory compliance requirements to the Project, regulatory requirements, standards and guidelines, BMPs, and permit conditions are listed in Table 22. The EDFs beyond regulatory requirements that have been proposed and committed to by Perpetua are listed in Appendix A. In particular, activities associated with the CMP for impacts on wetland resources are summarized in Table 23. The effects analysis in the BA takes these EDFs as well as regulatory requirements into consideration, such that the identified potential impacts of the SGP will be those that remain after their application. These EDFs and regulatory requirements will be applied to reduce and minimize impacts to resources from the SGP.

# **Table 22. Prominent Regulatory and Forest Plan Requirements.**

















#### **Table 23. Overview of Specific Activities in the Compensatory Mitigation Plan.**













**Figure 23. Stream and Wetland Restoration Areas, Stibnite Gold Project.**


**Figure 24. Perpetua Lemhi River Restoration Project Overview, Stibnite Gold Project.**

### **1.11.1. Agency Requirements**

Mitigation measures required by the USFS will represent reasonable and effective means to reduce the impacts identified in the resource analysis or to reduce uncertainty regarding the forecasting of impacts into the future. If environmental impacts are inevitable, certain regulatory programs may require compensatory mitigation of the impacts. Any mitigation measures are in addition to the regulatory requirements (Table 22) and Project EDFs (Appendix A) accounted for in the impact analysis.

The Project contains a number of EDFs, operational activities, best practices, mitigation measures, and monitoring plans intended to reduce impacts to the environment. An IARB will be formed to provide oversight for the Project's environmental-related activities including adaptive management. The IARB will consist of all permitting agencies including IDEQ, IDWR, NMFS, USFWS, EPA, Valley County, and the USFS<sup>6</sup>.

Member agencies on the IARB will have access to Project design reports, Project as-built drawings, monitoring reports, model updates required by mitigation measures, and any environmental action plans. These agencies will also have the opportunity to provide input where appropriate on Project documentation. Specific construction stage documentation subject to IARB review upon their completion include:

- Construction design of the water treatment plants;
- Construction design of the TSF;
- Construction design of the processing plant facility components;
- Construction design of the Burntlog Route;
- Construction design of the fish tunnel;
- Construction design of the EFSFSR water intake;
- The final Stormwater Pollution Prevention Plan (SPPP);
- The final CMP:

 $\overline{a}$ 

- Monitoring and mitigation plans under the EMMP (including adaptive management); and,
- Engineering as-builts for completed facilities.

# **1.11.2. Stibnite Gold Mitigation Plan**

The basis of the Perpetua's proposed EDFs are impact avoidance and minimization up front or as part of operations. The potential impacts of the SGP remaining after applying the avoidance and minimization measures were addressed by Perpetua on a resource-basis by further avoidance, minimization, and/or compensatory mitigation described in proponent-proposed specific resource mitigation plans. The following mitigation plans have been developed for the SGP:

<sup>6</sup> The proposed action refers in several places to IARB "approval" of several plans and designs. NMFS role in that process will be to provide verification prior to finalization and implementation of plans and designs that will affect ESA-listed species that the effects remain consistent with the effects analysis and extent of take provided in this opinion.

- Stibnite Gold EMMP (Brown and Caldwell 2021d):
- Fisheries and Aquatic Resources Mitigation Plan (Brown and Caldwell, Rio ASE, and BioAnalysts 2021);
- Fishway Operations and Management Plan (FOMP) (Brown and Caldwell, McMillen Jacobs Associates, and BioAnalysts 2021);
- Compensatory Mitigation Plan (Tetra Tech 2023).
- Snow Avalanche Hazard Assessment for Access Roads (DAC 2021);
- Development Rock Management Plan (Brown and Caldwell 2022);
- Environmental Legacy Management Plan (Perpetua 2021c);
- Reclamation Closure Plan (Tetra Tech 2021);
- Transportation Management Plan (Perpetua 2022);
- Water Management Plan (Brown and Caldwell 2021b);
- Water Resources Monitoring Plan (Brown and Caldwell 2021c); and,
- 404 permit application including a draft CMP (Perpetua 2023).

Below is a brief discussion of the FMP, FOMP, CMP, and the Lemhi Restoration Project. The Water Resources Monitoring Plan was described in section 1.10.1.1.

Perpetua will integrate all required USFS and USACE requirements and mitigation commitments into the current draft EMMP (Brown and Caldwell 2021c). This EMMP consists of a program framework and appendices containing component monitoring and management plans. Perpetua will use the EMMP to guide monitoring, document permit compliance, implement impact reduction procedures, and address adaptive management thresholds and responses where impacts and mitigation effectiveness carry substantial uncertainty.

### **Fisheries and Aquatic Resources Mitigation Plan**

Perpetua's FMP (Brown and Caldwell, Rio ASE, and BioAnalysts 2021) describes the measures that Perpetua has proposed to minimize adverse impacts on fisheries and aquatic resources. Additional details regarding how these plans address specific activities and impacts are also summarized in Appendix B of the BA (Stantec 2024).

The FMP actions will begin during construction and continue throughout mine operations and into closure and reclamation. The FMP includes water quality protection; fish protection, salvage, and relocation during diversions and dewatering activities; a process of protection and salvage for draining of the YPP; measures to avoid impacts during blasting; monitoring streamflow; restoring passage in stream channels with fish passage impediments; and monitoring of fish and aquatic biota. The FMP and its components will continue to be refined in consultation with natural resource and regulatory agencies.

### **Fishway Operations and Management Plan**

Perpetua has proposed a fishway for upstream and downstream passage of anadromous and migratory fish in the EFSFSR during construction and mine operations, to be part of the tunnel that diverts the EFSFSR around the YPP. Additional details regarding how these plans address specific activities and impacts are also summarized in Appendix B of the BA (Stantec 2024).

Perpetua's FOMP (Brown and Caldwell, McMillen Jacobs, and BioAnalysts 2021) outlines the operation of the fishway and monitoring for effective fish passage as well as an adaptive approach to provide for fish trap and haul operations as an alternative, using the same facilities consistent with 2022 NMFS guidelines for fish passage. Fish protection measures for the EFSFSR tunnel and YPP dewatering are outlined as well, such as a temporary fish barrier downstream of the YPP during tunnel construction carefully sequenced dewatering of the YPP, and start of fishway operations.

Measures to avoid and minimize impacts to fish habitat are detailed in the FMP and FOMP (Brown and Caldwell, Rio ASE, and BioAnalysts 2021; Brown and Caldwell, McMillen Jacobs, and BioAnalysts 2021). These measures include the following:

- Water quality protection measures designed on managing contact and non-contact water to maintain and improve water quality while supplying sufficient water for mining and ore processing. Diversions, ditches, and other mine facilities will be lined and/or water collected and treated to protect water quality. Riparian corridors will be restored and enhanced, and certain diversions piped, to reduce stream temperatures. Water treatment will continue during both operations and the post-closure phase.
- Fish protection, salvage, and relocation during dewatering and diversions measures for screening or excluding of fish from diversion channels, water withdrawals, low-flow pipes, and the YPP dewatering to exclude and protect fish. Work windows have been developed based on fish periodicity to account for the different life stages of the targeted fish species. During diversions and dewatering activities in fish bearing streams, fish handling and salvage protection measures have been identified to safely isolate, collect, handle, and transport the fish.
- Instream work window has been established to protect spawning and incubation. The instream work window is from May 1 to August 1, providing there are no incubating eggs (i.e., redds) within the construction area. If incubating eggs are present, the work window will be from June 15 to August 1. September 15 to April 30 could be an alternate work window that will avoid spawning adults and minimize impacts to juveniles as long as there is no documented spawning and therefore no incubation occurring in the affected stream section.
- Trap and haul protocols at the fishway (if needed) the primary goal is operating and maintaining the EFSFSR fishway during construction and operations, and later in the mine life by restoring the EFSFSR stream channel over the backfilled YPP to provide permanent, volitional upstream and downstream fish passage and access to important stream habitats of the upper EFSFSR and portions of Meadow Creek. If fish are not able to use the fishway during any period, trap and haul procedures have been developed to safely collect, handle, and move fish upstream of the fishway. Details regarding the proposed use of trap and haul were described in Appendix C of the FOMP (Brown and Caldwell, McMillen Jacobs, and BioAnalysts 2021), operating as follows:
	- o Trap and haul will only be used when deemed necessary to avoid further delay of adult passage near or during the spawning period. At one week prior to the spawning period, if adults present in the resting pool below the fishway have not entered the fishway over a 48-hour period and proceeded up the fishway they will be collected and transported upstream without further delay.
- o The migration period (i.e., Chinook Salmon: Jul. 7 Sep. 15; Steelhead: Apr 1. May 31); will be monitored via fishway video and Passive Integrated Transponders (PIT) tag detections in the fishway. In addition, environmental staff will also provide input on visual observations of adults near the fishway entrance. Arrivals at the fishway and redd surveys will be conducted to refine migration and spawning periods of fish passing the EFSFSR tunnel. Information will be presented annually to the IARB and periods of migration and spawning will be adjusted as needed.
- o Fish will be crowded within the first fishway pool with removable picket panels and netted with a large sanctuary dip net. Once in the sanctuary dip net, fish will be removed and placed into a wetted transfer boot to move fish from the capture pool to the transport tank. Water-to-water transfer is the goal.
- o If necessary, trap and haul is only expected to occur once per day but may occur twice if target species need to be separated (i.e., bull trout and Chinook salmon) or there is sufficient fish onsite that require more than one transport.
- Avoidance measures during blasting activities measures to largely avoid or minimize the potential effects from blasting activities using appropriate setback distances from aquatic habitats to limit blast-related air overpressure and ground vibrations to harmless levels. Other additional blasting techniques can also be used to reduce these levels, and BMPs and site-specific modification of methods can further minimize or prevent damage to fish and the aquatic environment.
- Monitoring streamflow activities for maintaining, to the extent practicable, appropriate streamflows and streamflow monitoring in natural or restored channels where fish are present.
- Stream restoration and enhancement design elements for stream restoration and enhancement based on natural channel design principles intended to restore permanent fish passage at YPP, improve fish habitat site-wide for spawning and rearing salmonids, and provide a net ecological benefit relative to current conditions.
- Restoring passage in stream channels removing existing passage barriers within the mine site to allow for fish movement between streams and areas of the mine site where access is currently blocked or impeded within the SGP footprint as well as along the Burntlog Route.
- Monitoring fish and aquatic biota provide the data necessary to evaluate how the various mitigation and protection measures are implemented, and to assess the status and trends and ongoing effectiveness. To address the potential for variances in the outcome of these measures, an adaptive management approach is outlined that will provide the mechanism to modify or adjust these measures or approaches in response to monitoring and evaluation as well as new information or technologies that may become available over the more than 20 years of construction, mining, reclamation, and restoration.

When diverting and restoring stream channels, cofferdams will isolate portions of the stream channel slated for restoration within the existing ordinary high-water mark (OHWM) to keep water and fish out of the new channel until construction is completed. Once the new channel is completed (including prewashing the substrate), water will be slowly reintroduced into the new channel (one-third of the flow initially), with seine block nets keeping fish from entering the new channel. Seine block nets will be placed in the upstream end of the original channel, which will

then be electrofished to remove all fish before all flow can be rerouted into the new channel. Any fish captured will be moved upstream of the seine block net. Once the original channel is cleared, two-thirds of the flow will be released into the new channel, and then ultimately all flow will be released into the new channel and the seine block net to the new channel removed. The original channel will be permanently blocked from the new channel and then filled with clean native alluvium as the new floodplain. Steps for isolating the stream channel include:

- Temporary cofferdams placed between the actively flowing river surface water and all active work areas. Temporary cofferdams may be placed at additional locations to achieve required water quality standards, or to simplify construction determined by the contractor.
- Fill material for bulk bags or 'super sacks," if used, shall be clean, washed, and rounded material similar in gradation to existing channel substrate, and not contain fines. Material must be approved before use.
- Cofferdams and diversion dams must be built in a manner to meet turbidity limits as defined in the project specifications. Use of gravel and soil to build a pushup type cofferdam or flow diversion dam are acceptable at all locations not connected to surface water flow but will not be allowed in the actively flowing channel.

When reintroducing water to dewatered areas and newly constructed channels, a staged rewatering plan will be applied. The following will be applied to all rewatering efforts:

- Turbidity monitoring protocol will be applied to rewatering effort.
- Pre-wash the area before rewatering. Turbid wash water will be detained and pumped to the floodplain or sediment capture areas rather than discharging to fish-bearing channels.
- Install seine nets at upstream end to prevent fish from moving from downstream until 2/3 of the total flow is restored to the channel.
- Starting in early morning, introduce 0.33 of new channel flow over period of 1 to 2 hours.
- Introduce second third of flow over the next 1 to 2 hours and begin fish salvage of bypass channel if fish are present.
- Remove upstream seine nets once  $2/3$  flow in rewatered channel and downstream turbidity is within acceptable range (less than 40 NTU) or less than 10 percent of the background condition).
- Introduce final third of flow once fish salvage efforts are complete, and downstream turbidity verified to be within acceptable range.
- Install plug to block flow into old channel.
- Follow same steps when rewatering mainstem.

Turbidity monitoring will include:

• Record turbidity reading, location, and time for background reading approximately 100 ft. upstream from the project area using a recently calibrated turbidimeter or via visual observation.

- Record the turbidity reading, location, and time at the measure compliance location point.
	- o 50 ft. downstream for streams less than 30 ft. wide;
	- o 100 ft. downstream for streams between 30 and 100 ft. wide; and,
	- o 200 ft. downstream for streams greater than 100 ft. wide.
- Turbidity will be measured (background location and compliance point) every 4 hours while work is being implemented.
- If exceedances occur for more than two consecutive monitoring intervals (after 8 hours), the activity will stop until the turbidity level returns to background. The Idaho Office of Species Conservation (OSC) will be notified for all exceedances and corrective actions at project completion.
- If turbidity controls (cofferdams, wattles, fencing, etc.) are determined ineffective, crews will be mobilized to modify, as necessary. Occurrences will be documented in the project daily reports.

### **1.11.2.3. Compensatory Mitigation Plan**

Construction of the SGP will permanently impact wetlands and other WOTUS subject to regulation under Section 404 of the CWA and requires a USACE permit pursuant to Section 404. Perpetua's CMP (Tetra Tech 2023) provides detailed descriptions of proposed restoration, establishment, enhancement, and/or preservation of aquatic resources to compensate for unavoidable impacts to WOTUS associated with activities that will be authorized by a USACE permit. Additional details regarding how these plans address specific activities (Table 24) and impacts, including protection measures, are summarized in BA Appendix B. The CMP can be updated and revised until the USACE has determined all mitigation requirements. The CMP demonstrates the feasibility of achieving the amount and types of mitigation to offset the impacts in a manner consistent with the 2008 Mitigation Rule. The CMP provides detailed descriptions of proposed restoration, establishment, enhancement, and/or preservation of aquatic resources to compensate for unavoidable impacts to WOTUS associated with activities that will be authorized by a USACE permit (Tetra Tech 2023).



### **Table 24. 404 Permit Activities.**



The CMP describes mitigation to address the requirements of the USACE and EPA under the Compensatory Mitigation for Losses of Aquatic Resources under CWA Section 404 (Final Rule) (USACE and EPA 2008). The CMP includes the 12 required elements of compensatory mitigation plans (33 CFR 332.4(c)/40 CFR 230.94(c)): objectives; maintenance plan; site selection; performance standards; site protection; monitoring requirements; baseline information; long-term management plan; determination of credits; adaptive management plan; mitigation work plan; and financial assurances.

The activity authorized under the 404 permit is summarized in Table 24. The CMP will be revised as appropriate by the USACE Regulatory Division—Walla Walla District, Boise Field Office, in compliance with the CWA Section 404/DA permit, stream and wetland delineations and jurisdictional determinations, development of the stream functional assessment for USACEapproved stream functional analysis, wetland and stream credits and debits determinations, and compliance with USACE's 404(b)(1) Guidelines (40 CFR Part 230).

# **Perpetua Lemhi River Restoration Project**

Perpetua has proposed the LRLT's Little Springs Conservation Easement and Restoration Project in an effort to use offsite mitigation to offset temporal losses to fish habitat in the SFSR drainage (Figures 2 and 23). The LRLT is negotiating a perpetual conservation easement for the property to ensure the restoration benefits and associated mitigation credits persist for at least the length of the predicted temporal loss for the SGP (Tetra Tech 2023). This project reach is located on private ranch lands on the upper Lemhi River, approximately 12 miles northwest of Leadore, Idaho. The project reach extends approximately 7,000 ft. from RM 42.63 to RM 41.32 of the mainstem Lemhi River on the west side of SH 28 (Rio ASE 2023).

The primary goals of the Lemhi Project (Figure 24) are to improve habitat for limiting life stages of ESA-listed fish species (i.e., pre-smolts (over-winter rearing), parr (summer rearing), adult (spawning and holding), and parr (high flow refugia), and to restore natural stream channel processes to maintain diverse habitat over time (Rio ASE 2023). Improving stream habitat conditions are intended to help increase fish population abundance, productivity, and spatial structure. Specific Lemhi Project targets are:

- Increase habitat quality and complexity (especially for juvenile life stages) by creating multi-threaded channels and connected off-channel habitat.
- Reduced W:D where the channel is over-widened to increase hydraulic complexity, floodplain connection, pool scour potential, shade, cover, and natural channel-forming processes.
- Increased frequency, duration, and area of floodplain connection to provide high flow refugia for rearing juveniles and to improve fine sediment distribution, groundwater recharge, floodwater storage, and nutrient cycling.
- Increased instream structure, hydraulic diversity, and more variable instream velocity.
- Increased pool quantity, frequency, and complexity.
- Surface/groundwater interchange to moderate instream temperature and provide areas of localized temperature refuge.
- Increased instream cover and interstitial space along margins for rearing life stages and for adult holding and cover leading up to and during spawning.
- Creation of a riparian corridor to increase shade, provide overhead cover, stabilize banks, provide instream structure, and increase LWD recruitment potential.

Lemhi Project elements targeting these objectives are presented below and illustrated on Figure 24 and Appendix D, Figures D-1 to D-22:

• Develop a multi-threaded channel network of 12,426 ft. (2.35 mi.) perennial and nonperennial side channels and 4,965 ft. (0.94 mi.) of non-perennial tertiary channels through excavation of new channels and pilot channels to target flow into existing low areas. Tertiary channels are low depressions in the ground surface or relic channels disconnected from the mainstem that will largely exist or be constructed to convey surface water seasonally, and as such, will involve little to no excavation and/or treatment resulting in natural evolution with variable outcomes (i.e., some tertiary channels may develop into perennial side channels while the remainder become abandoned).

- Install large and small woody material (e.g., beaver dam analogues, robust beaver dams, apex log and bleeder jams, etc.) to promote in-channel complexity, force hydraulic response (scour, deposition, split flow, floodplain connection, sediment sorting, and overall hydraulic diversity), and provide concealment cover for juvenile salmonids.
- Add floodplain roughness structures (e.g., logs, willow clumps, and slash) to provide high flow refugia, accommodate future channel dynamics, and promote lateral channel migration while maintaining a multi-threaded and sinuous channel character.
- Increase frequency of floodplain activation through channel constriction, blocking, and appropriate channel sizing of new channels and resizing of existing channel(s).
- Revegetation by means of planting native species within the riparian zone and transplanting local vegetation harvested near the Lemhi Project site; existing, mature riparian vegetation is limited within the Lemhi Project vicinity and will be preserved and used as floodplain roughness and/or bank roughness where available and appropriate.

This project will reconnect and/or create a series of interconnected perennial and non-perennial side channels and wetland complexes within a broad floodplain area that has been previously impacted by land use alterations, riparian vegetation clearing, levees, and grazing. The existing, single-threaded Lemhi River channel will be bifurcated and obstructed using natural materials at multiple locations to force flow into relic channels on the floodplain (Tetra Tech 2023). The amount of excavation is expected to range from full channel excavation similar to the mainstem, to small pilot channels connecting adjacent relic channels, to no excavation where channel geometry and conditions allow. Tertiary channels (i.e., low depressions or relic channels disconnected from the mainstem) that largely exist or will be constructed to convey surface water seasonally (primarily non-perennial). Little to no excavation and/or treatment is proposed for these channels, which are expected to evolve naturally over time with variable outcomes. Some tertiary channels may develop into perennial side channels while others may become abandoned. (Rio ASE 2023).

Riparian vegetation in conjunction with relic channel scars throughout the floodplain provide the framework upon which the multi-threaded channel network will be restored. The strategy for restoration therefore includes removing levees and areas of high ground to reconnect relic flow paths, while also narrowing and raising the existing mainstem channel to reduce its conveyance and force flow into the newly reconnected secondary channels and floodplain. The strategy employs both active and passive restoration, whereby excavation would be used (active restoration) to cut areas of high ground and fill portions of the existing channel. Likewise, woody debris and beaver dams will be actively installed to obstruct flow and provide instream habitat. The newly activated secondary channels and floodplain will also scour and create new channels over time (i.e., passive restoration) with riparian vegetation providing structure and habitat. The elevated groundwater table will also facilitate more rapid and expansive riparian vegetation reestablishment, which will improve channel evolution and associated development of complex

habitat over time, including the maintenance of a narrow W:D, bank stabilization, and promotion of scour pool development. Improved riparian vegetation will also provide a food source for upland species and beaver and improve nutrient cycling in the river (Rio ASE 2023).

Large woody material structures are proposed in the main channel, side channels, and floodplain to provide roughness and habitat throughout the project area. There are a variety of proposed structures that will consist of key log members that act as the skeleton of the structure. The structure will then be completed with the addition of woody racking material and slash. These structures are intended to emulate natural log jams, and therefore these materials will not be permanently locked or fixed in place; pieces may move throughout the project reach as flows peak (Rio ASE 2023).

Construction is scheduled to be completed by MY -2 (Tetra Tech 2023), and will be staged so that the new channel will be created first. Once the new channel is completed, the channel will be 'pre-washed' into a reach equipped with sediment capture devices prior to the introduction of flow into the new channel. After this stage is completed, the Lemhi River bank will be breached and flow will be rerouted into the side channel. A cofferdam will be installed to block the flow in the mainstem so work can occur in the mainstem with no active flow. As flows are reintroduced to the Lemhi and other created channels, gravels will be 'pre-washed' in an effort to control turbidity (Stantec 2024).

Any topsoil and native channel material displaced by construction will be stockpiled away from the channel for use during site restoration. When construction is finished, all streambanks, soils, and vegetation will be cleaned and restored as necessary using stockpiled topsoil and native channel material to renew ecosystem processes that form and maintain productive habitats (Stantec 2024).

Fish will not have access to the construction area during construction activities. The Lemhi project will be conducted so that no work will occur in the flowing channel to minimize or avoid impacts caused by sediment and turbidity, changes to water quality (spills risk or chemical contaminants), and noise and vibration. There will be no physical barriers created or removed in the Lemhi Analysis Area other than the cofferdam to block or redirect flows depending on the construction stage. Staging construction so side channels are created before restoring the main channel is key to reducing the impacts. The text below describes the process that will be taken to avoid working in-water.

Work areas will be isolated for performing work within the in-water work window (first quarter of July (Q1) through the third quarter of August (Q3). Cofferdams will be placed within the existing Lemhi River channel to isolate areas for excavated connections between the existing and new channel. The new channel will then be activated, and additional cofferdams will be required to complete filling of the existing Lemhi River channel and installation of wood habitat structures.

Cofferdams will isolate portions of the proposed channel within the existing OHWM to keep water and fish out of the new channel until construction is completed. Once the new channel is completed (including prewashing the substrate), water will be slowly reintroduced into the new channel (one-third of the flow initially), with seine block nets keeping fish from entering the new channel. Seine block nets will be placed in the upstream end of the original channel, which will then be electrofished to remove all fish before all flow can be rerouted into the new channel. Any fish captured will be moved upstream of the seine block net. Once the original channel is cleared, two-thirds of the flow will be released into the new channel, and then ultimately all flow will be released into the new channel and the seine block net to the new channel removed. The original channel will be permanently blocked from the new channel and then filled with clean native alluvium as the new floodplain. The steps for staged dewatering include:

- Temporary cofferdams placed between the actively flowing river surface water and all active work areas. Temporary cofferdams may be placed at additional locations to achieve required water quality standards, or to simplify construction determined by the contractor.
- Fill material for bulk bags or 'super sacks", if used, shall be clean, washed, and rounded material similar in gradation to existing channel substrate, and not contain fines.
- Cofferdams and diversion dams must be built in a manner to meet turbidity limits as defined in the project Specifications. Use of gravel and soil to build a pushup type cofferdam or flow diversion dam are acceptable at all locations not connected to surface water flow but will not be allowed in the actively flowing channel.

Staged Re-watering – when reintroducing water to dewatered areas and newly constructed channels, a staged re-watering plan will be applied. The following will be applied to all rewatering efforts:

- Turbidity monitoring protocols will be applied to the re-watering effort.
- The area will be pre-washed before re-watering. Turbid wash water will be detained and pumped to the floodplain or sediment capture areas rather than discharging to fishbearing channels.
- Seine nets will be installed at the upstream end to prevent fish from moving from downstream until 2/3 of the total flow is restored to the channel.
- Starting in early morning, 1/3 of the new channel flow will be introduced over a period of 1 to 2 hours.
- The second third of flow will be introduced over the next 1 to 2 hours and fish salvage of the bypass channel will begin if fish are present.
- Upstream seine nets will be removed once 2/3 flow in the re-watered channel and downstream turbidity is within an acceptable range (less than 40 NTU or less than 10 percent of the background condition).
- The final third of flow will be introduced once fish salvage efforts are complete, and downstream turbidity verified to be within acceptable range.
- A plug will be installed to block flow into the old channel.
- The same steps will be followed when re-watering the mainstem.

Turbidity monitoring will include:

- Record turbidity reading, location, and time for background reading approximately 100 feet upstream from the project area using a recently calibrated turbidimeter or via visual observation.
- Record turbidity reading, location, and time at the measure compliance location point, located as follows.
	- o 50 feet downstream for streams less than 30 feet wide;
	- o 100 feet downstream for streams between 30 and 100 feet wide;
	- o 200 feet downstream for streams greater than 100 feet wide.
- Turbidity will be measured (background location and compliance point) every 4 hours while work is being implemented.
- If exceedances occur for more than two consecutive monitoring intervals (after 8 hours), the activity will stop until the turbidity level returns to background. The OSC will be notified for all exceedances and corrective actions at project completion.
- If turbidity controls (cofferdams, wattles, fencing, etc.) are determined ineffective, crews will be mobilized to modify, as necessary. Occurrences will be documented in the project daily reports.

Additional details regarding the Lemhi Project are described in Appendix D.

We considered, under the ESA, whether or not the proposed action would cause any other activities and determined that it would not.

### **2. ENDANGERED SPECIES ACT: BIOLOGICAL OPINION AND INCIDENTAL TAKE STATEMENT**

The ESA establishes a national program for conserving threatened and endangered species of fish, wildlife, plants, and the habitat upon which they depend. As required by section  $7(a)(2)$  of the ESA, each federal agency must ensure that its actions are not likely to jeopardize the continued existence of endangered or threatened species, or adversely modify or destroy their DCH. Per the requirements of the ESA, federal action agencies consult with NMFS and section 7(b)(3) requires that, at the conclusion of consultation, NMFS provide an opinion stating how the agency's actions would affect listed species and their critical habitats. If incidental take is reasonably certain to occur, section 7(b)(4) requires NMFS to provide an ITS that specifies the impact of any incidental taking and includes RPMs and terms and conditions to minimize such impacts.

The PNF and USACE determined the proposed action is not likely to adversely affect SRKW and its DCH. Our concurrence is documented in the "Not Likely to Adversely Affect" Determinations section (Section 2.13).

# **2.1. Analytical Approach**

This opinion includes both a jeopardy analysis and an adverse modification analysis. The jeopardy analysis relies upon the regulatory definition of "jeopardize the continued existence of" a listed species, which is "to engage in an action that reasonably would be expected, directly or indirectly, to reduce appreciably the likelihood of both the survival and recovery of a listed species in the wild by reducing the reproduction, numbers, or distribution of that species" (50) CFR 402.02). Therefore, the jeopardy analysis considers both survival and recovery of the species.

This opinion also relies on the regulatory definition of "destruction or adverse modification," which "means a direct or indirect alteration that appreciably diminishes the value of critical habitat as a whole for the conservation of a listed species" (50 CFR 402.02).

The designations of critical habitat for SR spring/summer Chinook salmon and SR Basin steelhead use the term primary constituent element (PCE) or essential features. The 2016 final rule (81 FR 7414; February 11, 2016) that revised the critical habitat regulations (50 CFR 424.12) replaced these terms with physical or biological features (PBFs). The shift in terminology does not change the approach used in conducting a "destruction or adverse modification" analysis, which is the same regardless of whether the original designation identified PCEs, PBFs, or essential features. In this opinion, we use the term PBF to mean PCE or essential feature, as appropriate for the specific critical habitat.

The ESA Section 7 implementing regulations define effects of the action using the term "consequences" (50 CFR 402.02). As explained in the preamble to the final rule revising the definition and adding this term (84 FR 44976, 44977; August 27, 2019), that revision does not change the scope of our analysis, and in this opinion, we use the terms "effects" and "consequences" interchangeably.

We use the following approach to determine whether a proposed action is likely to jeopardize listed species or destroy or adversely modify critical habitat:

- Evaluate the rangewide status of the species and critical habitat expected to be adversely affected by the proposed action.
- Evaluate the environmental baseline of the species and critical habitat.
- Evaluate the effects of the proposed action on species and their critical habitat using an exposure–response approach.
- Evaluate cumulative effects.
- In the integration and synthesis, add the effects of the action and cumulative effects to the environmental baseline, and, in light of the status of the species and critical habitat, analyze whether the proposed action is likely to: (1) directly or indirectly reduce appreciably the likelihood of both the survival and recovery of a listed species in the wild by reducing the reproduction, numbers, or distribution of that species; or (2) directly or indirectly result in an alteration that appreciably diminishes the value of critical habitat as a whole for the conservation of a listed species.
- If necessary, suggest a reasonable and prudent alternative (RPA) to the proposed action.

This opinion relies on a series of assumptions for this analysis. Deviation from these assumptions may fall outside the scope of this consultation and may trigger subsequent ESA review. NMFS' assumptions include:

- Exploration for additional ore reserves is anticipated to continue through the life of the SGP operations. If additional ore tonnage is identified and defined, the production life of the SGP may be extended beyond the currently proposed mine life schedule. Any such extension and the associated operations and effects are beyond the scope of this consultation.
- The proposed action states that Perpetua will use the same or similar drilling methods and environmental protection measures that have been employed for exploration drilling in the past (i.e., exploration under the Golden Meadows Exploration Plan). Therefore, NMFS assumes that all environmental protection measures from the Golden Meadows Exploration Plan including, but not necessarily limited to, those specified in Appendix B will be employed.
- Noxious weed and invasive species plant control will be conducted according to methods approved in the 2020 Programmatic Activities Biological Opinion (PNF 2020). Therefore, NMFS assumes that all environmental protection measures associated with herbicide use including, but not necessarily limited to, those specified in Appendix C will be employed.
- Because Perpetua will alternatively use trap and haul to get fish upstream of the YPP cascade should tunnel passage not work as designed, NMFS assumes that ESA-listed salmonids will have access and be present in stream segments upstream of the cascade beginning in MY -1, whether the tunnel functions as intended or not.
- No untreated contact water at the ground surface (e.g., pond overflow or failures in contact water collection) will be discharged directly into surface waters.
- The SWWC model outputs represent the best available information regarding the predicted contaminant concentrations in surface waters resulting from project implementation.
- Mixing zones for IPDES discharges of pollutants will not be authorized; IPDESauthorized discharges will be required to meet end-of-pipe limits.
- A rigorous effectiveness monitoring program will be finalized with input from the IARB and implemented as designed. This program will have sufficient monitoring components to provide for early detection of unanticipated environmental impacts in surface waters.
- An effective adaptive management plan will be finalized with input from the IARB and adequate financial commitments (financial bonding) will be adhered to.
- The percent change in stream flow information presented in Table 4.1-35 of the BA are accurate and are based on the average net effect of the proposed action on flow rates in affected stream reaches.
- Riparian revegetation efforts will be successful, shade reductions assumed in the SPLNT model will be achieved, and predicted maximum weekly maximum stream temperatures will not be exceeded. If shade reductions are not realized or if actual stream temperatures are greater than predicted at any time during project implementation (including closure and post-closure), effective adaptive management actions will be implemented to reduce stream temperatures.

# **2.2. Rangewide Status of the Species and Critical Habitat**

This opinion examines the status of each species that would be adversely affected by the proposed action. The status is determined by the level of extinction risk that the listed species face, based on parameters considered in documents such as recovery plans, status reviews, and listing decisions. This informs the description of the species' likelihood of both survival and recovery. The species status section also helps to inform the description of the species' "reproduction, numbers, or distribution" for the jeopardy analysis. The opinion also examines the condition of critical habitat throughout the designated area, evaluates the conservation value of the various watersheds. The Federal Register notices and notice dates for the species and critical habitat listings considered in this opinion are included in Table 25.

**Table 25. Listing status, status of critical habitat designations and protective regulations, and relevant Federal Register decision notices for ESA-listed species considered in this opinion.**

<b>Species</b>	Listing Status <sup>1</sup>	Critical Habitat <sup>2</sup>	Protective Regulations		
Chinook salmon (Oncorhynchus tshawytscha)					
Snake River spring/summer-run	T 4/22/92; 57 FR 14653	12/28/93; 58 FR 68543	6/28/05; 70 FR 37160		
Steelhead (O. mykiss)					
Snake River Basin	T 8/18/97; 62 FR 43937	9/02/05; 70 FR 52630	6/28/05; 70 FR 37160		

Note: Listing status 'T' means listed as threatened under the ESA.

<sup>1</sup>The listing status for Snake River spring/summer Chinook salmon was corrected on 6/3/92 (57 FR 23458).

<sup>2</sup>Critical habitat for Snake River spring/summer Chinook salmon was revised on 10/25/99 (64 FR 57399).

#### **2.3. Status of the Species**

This section describes the present condition of the SR spring/summer Chinook salmon evolutionarily significant unit (ESU), and the SR Basin steelhead distinct population segment (DPS). NMFS expresses the status of a salmonid ESU or DPS in terms of likelihood of persistence over 100 years (or risk of extinction over 100 years). NMFS uses McElhany et al.'s (2000) description of a viable salmonid population (VSP) that defines "viable" as less than a 5 percent risk of extinction within 100 years and "highly viable" as less than a 1 percent risk of extinction within 100 years. A third category, "maintained," represents a less than 25 percent risk within 100 years (moderate risk of extinction). To be considered viable, an ESU or DPS should have multiple viable populations so that a single catastrophic event is less likely to cause the ESU/DPS to become extinct, and so that the ESU/DPS may function as a metapopulation that can sustain population-level extinction and recolonization processes (ICTRT 2007). The risk level of the ESU/DPS is built up from the aggregate risk levels of the individual populations and major population groups (MPGs) that make up the ESU/DPS.

Attributes associated with a VSP are: (1) abundance (number of adult spawners in natural production areas); (2) productivity (adult progeny per parent); (3) spatial structure; and (4) diversity. A VSP needs sufficient levels of these four population attributes in order to: safeguard the genetic diversity of the listed ESU or DPS; enhance its capacity to adapt to various environmental conditions; and allow it to become self-sustaining in the natural environment (ICTRT 2007). These viability attributes are influenced by survival, behavior, and experiences throughout the entire salmonid life cycle, characteristics that are influenced in turn by habitat and other environmental and anthropogenic conditions. The present risk faced by the ESU/DPS informs NMFS' determination of whether additional risk will appreciably reduce the likelihood that the ESU/DPS will survive or recover in the wild.

The following sections summarize the status and available information on the species and DCHs considered in this opinion based on the detailed information provided by the *ESA Recovery Plan for Snake River Spring/Summer Chinook Salmon & Snake River Basin Steelhead* (NMFS 2017); *Biological Viability Assessment Update for Pacific Salmon and Steelhead Listed Under the Endangered Species Act: Pacific Northwest* (Ford 2022); *2022 5-Year Review: Summary & Evaluation of Snake River Spring/Summer Chinook Salmon (NMFS 2022b)*; and *2022 5-Year Review: Summary & Evaluation of Snake River Basin Steelhead (NMFS 2022c)*. These documents are incorporated by reference here. Additional information that has become available since these documents were published is also summarized in the following sections and contributes to the best scientific and commercial data available.

### **2.4. Snake River Spring/Summer Chinook Salmon**

A summary of the current status of the Snake River spring/summer Chinook salmon ESU can be found on NMFS' publicly available intranet site ( [https://www.fisheries.noaa.gov/s3/2024-](https://www.fisheries.noaa.gov/s3/2024-08/status-species-snake-river-spring-summer-chinook-salmon-july-2024.pdf) [08/status-species-snake-river-spring-summer-chinook-salmon-july-2024.pdf\)](https://www.fisheries.noaa.gov/s3/2024-08/status-species-snake-river-spring-summer-chinook-salmon-july-2024.pdf), and is incorporated by reference here (NMFS 2024a). Overall, the species is at a moderate-to-high risk of extinction within the next 100 years. The SFSR and Upper Salmon River are two of the five MPGs that make up the ESU. Populations that may be affected by the proposed action include: EFSFSR, SFSR, and Lemhi River (Table 26) (3 of 28 extant populations in the ESU). To achieve recovery, all MPGs in the ESU must be viable. In order for the MPG to be viable, it must be comprised of multiple, relatively nearby, highly viable, viable, and maintained populations. Neither the SFSR MPG nor the Upper Salmon River MGP are viable. To support MGP viability and to achieve recovery of the species, the EFSFSR population must be maintained, and the SFSR and Lemhi River populations must reach viable status.





 $1R$  isk ratings are defined based on the risk of extinction within 100 years: High = greater than or equal to 25 percent; Moderate = less than 25 percent; Low = less than 5 percent; and Very Low = less than 1 percent.

<sup>2</sup>There are several scenarios that could meet the requirements for ESU recovery (as reflected in the proposed goals for populations in Oregon and Washington). What is reflected here for populations in Idaho are the proposed status goals selected by NMFS and the State of Idaho.

### **2.5. Snake River Basin Steelhead**

A summary of the current status of the SR Basin steelhead DPS can be found on NMFS' publicly available intranet site ( [https://www.fisheries.noaa.gov/s3/2024-08/status-species-snake-river](https://www.fisheries.noaa.gov/s3/2024-08/status-species-snake-river-basin-steelhead-july-2024.pdf)[basin-steelhead-july-2024.pdf\)](https://www.fisheries.noaa.gov/s3/2024-08/status-species-snake-river-basin-steelhead-july-2024.pdf), and is incorporated by reference here (NMFS 2024b). Overall, available information suggests that SR Basin steelhead continue to be at a moderate risk of extinction within the next 100 years. The Salmon River MPG is one of six MPGs that make up the DPS. To achieve recovery, all MPGs in the ESU must be viable. In order for the MPG to be viable, it must be comprised of multiple, relatively nearby, highly viable, viable, and maintained populations. The Salmon River MPG is comprised of twelve of the 24 extant populations in the DPS. Populations that may be affected by the proposed action include: the SFSR and the Lemhi River populations (Table 27) (2 of 24 extant populations in the DPS). To support MPG viability and to achieve species recovery, the SFSR and Lemhi River populations must reach viable status.

**Table 27. Summary of Viable Salmonid Population (VSP) parameter risks and overall current status and proposed recovery goals for populations of Snake River Basin steelhead that may be affected by the action.**

		<b>VSP Risk Rating</b> <sup>1</sup>		<b>Viability Risk Rating</b>	
<b>MPG</b>	<b>Population</b>	Abundance/ <b>Productivity</b>	<b>Spatial</b> <b>Structure/Diversity</b>	2022 <b>Assessment</b>	<b>Proposed</b> <b>Recovery</b> Goal <sup>2</sup>
Salmon River	South Fork Salmon River	Moderate	Low	Maintained	Viable
	Lemhi River	Moderate	Moderate	Maintained	Viable

 $1R$  isk ratings are defined based on the risk of extinction within 100 years: High = greater than or equal to 25 percent; Moderate = less than 25 percent; Low = less than 5 percent; and Very Low = less than 1 percent.

<sup>2</sup>There are several scenarios that could meet the requirements for ESU recovery (as reflected in the proposed goals for populations in Oregon and Washington). What is reflected here for populations in Idaho are the proposed status goals selected by NMFS and the State of Idaho.

# **2.6. Status of Critical Habitat**

In evaluating the condition of DCH, NMFS examines the condition and trends of PBFs which are essential to the conservation of the ESA-listed species because they support one or more life stages of the species. Proper function of these PBFs is necessary to support successful adult and juvenile migration, adult holding, spawning, incubation, rearing, and the growth and development of juvenile fish. Modification of PBFs may affect freshwater spawning, rearing or migration of ESA-listed Chinook salmon and steelhead. Generally speaking, sites required to support one or more life stages of the ESA-listed species (i.e., sites for spawning, rearing, migration, and foraging) contain PBFs essential to the conservation of the listed species (e.g., spawning gravels, water quality and quantity, side channels, or food) (Table 28).

Table 29 describes the geographical extent of critical habitat within the Snake River basin for SR spring/summer Chinook and SR Basin steelhead. Critical habitat includes the stream channel and water column with the lateral extent defined by the ordinary high-water line, or the bankfull elevation where the ordinary high-water line is not defined. In addition, critical habitat for Chinook salmon includes the adjacent riparian zone, which is defined as the area within 300 feet of the line of high water of a stream channel or from the shoreline of standing body of water (58

FR 68543). The riparian zone is critical because it provides shade, streambank stability, organic matter input, and regulation of sediment, nutrients, and chemicals.

<b>Site</b>	<b>Essential Physical and Biological Features</b>	<b>Species Life Stage</b>		
Snake River Basin steelhead <sup>a</sup>				
Freshwater spawning	Water quality, water quantity, and substrate	Spawning, incubation, and larval development		
Freshwater rearing	Water quantity and floodplain connectivity to form and maintain physical habitat conditions	Juvenile growth and mobility		
	Water quality and forage <sup>b</sup>	Juvenile development		
	Natural cover <sup>c</sup>	Juvenile mobility and survival		
	Free of artificial obstructions, water quality	Juvenile and adult mobility		
Freshwater migration	and quantity, and natural cover <sup>c</sup>	and survival		
Snake River spring/summer Chinook salmon				
Spawning gravel, water quality and quantity, Spawning and juvenile rearing cover/shelter, food, riparian vegetation, space, and water temperature.		Juvenile and adult		
Migration	Substrate, water quality and quantity, water temperature, water velocity, cover/shelter, food <sup>d</sup> , riparian vegetation, space, safe passage	Juvenile and adult		

**Table 28. Types of sites, essential physical and biological features (PBFs), and the species life stage each PBF supports.**

a Additional PBFs pertaining to estuarine areas have also been described for Snake River steelhead. These PBFs will not be affected by the proposed action and have therefore not been described in this opinion.

**b** Forage includes aquatic invertebrate and fish species that support growth and maturation.

<sup>c</sup> Natural cover includes shade, large wood, log jams, beaver dams, aquatic vegetation, large rocks and boulders, side channels, and undercut banks.

<sup>d</sup>Food applies to juvenile migration only.

Critical habitat throughout the SR Spring/summer Chinook salmon and SR Basin steelhead designations varies from excellent in wilderness and roadless areas to poor in areas subject to intensive human land uses (NMFS 2017). Critical habitat throughout much of the Interior Columbia (which includes the Snake River and the Middle Columbia River) has been degraded by intensive agriculture, alteration of stream morphology (i.e., channel modifications and diking), riparian vegetation disturbance, wetland draining and conversion, livestock grazing, dredging, road construction and maintenance, logging, mining, and urbanization. Reduced summer streamflows, impaired water quality, and reduction of habitat complexity are common problems for critical habitat in non-wilderness areas. Human land use practices throughout the basin have caused streams to become straighter, wider, and shallower, thereby reducing rearing habitat and increasing water temperature fluctuations.

**Table 29. Geographical extent of designated critical habitat within the Snake River basin for ESA-listed salmon and steelhead.**

<b>Evolutionarily</b> Significant Unit (ESU)/ <b>Distinct Population</b> <b>Segment</b> (DPS)	<b>Designation</b>	<b>Geographical Extent of Critical Habitat</b>
<b>Snake River</b> spring/summer Chinook salmon ESU	58 FR 68543: December 28, 1993 64 FR 57399; October 25, 1999	All Snake River reaches upstream to Hells Canyon Dam; all river reaches presently or historically accessible to Snake River spring/summer Chinook salmon within the Salmon River basin; and all river reaches presently or historically accessible to Snake River spring/summer Chinook salmon within the Hells Canyon, Imnaha, Lower Grande Ronde, Upper Grande Ronde, Lower Snake-Asotin, Lower Snake- Tucannon, and Wallowa subbasins.
<b>Snake River Basin</b> steelhead DPS	70 FR 52630; September 2, 2005	Specific stream reaches are designated within the Lower Snake, Salmon, and Clearwater River basins. Table 21 in the Federal Register details habitat areas within the DPS's geographical range that are excluded from critical habitat designation.

In many stream reaches designated as critical habitat in the Snake River basin, streamflows are substantially reduced by water diversions (NMFS 2017). Withdrawal of water, particularly during low-flow periods that commonly overlap with agricultural withdrawals, often increases summer stream temperatures, blocks fish migration, strands fish, and alters sediment transport (Spence et al*.* 1996). Reduced tributary streamflow has been identified as a major limiting factor for SR spring/summer Chinook and SR Basin steelhead in particular (NMFS 2017).

Many stream reaches designated as critical habitat for these species are listed on the CWA 303(d) list for impaired water quality, such as elevated water temperature (IDEQ 2022). Many areas that were historically suitable rearing and spawning habitat are now unsuitable due to high summer stream temperatures, such as some stream reaches in the Upper Grande Ronde. Removal of riparian vegetation, alteration of natural stream morphology, and withdrawal of water for agricultural or municipal use all contribute to elevated stream temperatures. Water quality in spawning and rearing areas in the Snake River has also been impaired by high levels of sedimentation and by heavy metal contamination from mine waste (e.g., IDEQ and USEPA 2003; IDEQ 2001).

The construction and operation of water storage and hydropower projects in the Columbia River basin, including the eight run-of-river dams on the mainstem lower Snake and lower Columbia Rivers, have altered biological and physical attributes of the mainstem migration corridor. Hydrosystem development modified natural flow regimes, resulting in warmer late summer and fall water temperature. Changes in fish communities led to increased rates of piscivorous predation on juvenile salmon and steelhead. Reservoirs and project tailraces have created opportunities for avian predators to successfully forage for smolts, and the dams themselves have created migration delays for both adult and juvenile salmonids. Physical features of dams, such as turbines, also kill out-migrating fish. In-river survival is inversely related to the number of hydropower projects encountered by emigrating juveniles. However, some of these conditions have improved. The BOR and USACE have implemented measures in previous Columbia River System hydropower consultations to improve conditions in the juvenile and adult migration

corridor including 24-hour volitional spill, surface passage routes, upgrades to juvenile bypass systems, and predator management measures. These measures are ongoing and their benefits with respect to improved functioning of the migration corridor PBFs will continue into the future.

### **2.7. Climate Change Implications for ESA-listed Species and their Critical Habitat**

One factor affecting the rangewide status of SR salmon and steelhead, and aquatic habitat at large is climate change. As observed by Siegel and Crozier (2019), long-term trends in warming have continued at global, national, and regional scales. The five warmest years in the 1880 to 2019 record have all occurred since 2015, while 9 of the 10 warmest years have occurred since 2005 (Lindsey and Dahlman 2020). The years 2020 - 2022 were also hot years in national and global temperatures; all ranked in the top ten hottest years in the 141-year record of global land and sea measurements and followed the warmest decade on record. The year 2023 was even warmer, marking the hottest ever year on record

[\(https://www.ncei.noaa.gov/access/monitoring/monthly-report/global/202013\)](https://www.ncei.noaa.gov/access/monitoring/monthly-report/global/202013). Events such as the 2014-2016 marine heatwave (Jacox et al. 2018) are likely exacerbated by anthropogenic warming, as noted in the annual special issue of Bulletin of the American Meteorological Society on extreme events (Herring et al. 2018). The U.S. Global Change Research Program (USGCRP) reports average warming in the Pacific Northwest of about 1.3ºF from 1895 to 2011, and projects an increase in average annual temperature of 3.3ºF to 9.7ºF by 2070 to 2099 (compared to the period 1970 to 1999), depending largely on total global emissions of heat-trapping gases (predictions based on a variety of emission scenarios including B1, RCP4.5, A1B, A2, A1FI, and RCP8.5 scenarios). The increases are projected to be largest in summer (USGCRP 2018).

Climate change generally exacerbates threats and limiting factors, including those currently impairing salmon and steelhead survival and productivity. The growing frequency and magnitude of climate change related environmental downturns will increasingly imperil many ESA-listed stocks in the Columbia River basin and amplify their extinction risk (Crozier et al. 2019, 2020, 2021). This climate change context means that opportunities to rebuild these stocks will likely diminish over time. As such, management actions that increase resilience and adaptation to these changes should be prioritized and expedited. For example, the importance of improving the condition of and access and survival to and from the remaining functional, highelevation spawning and nursery habitats is accentuated because these habitats are the most likely to retain remnant snowpacks under predicted climate change (Tonina et al. 2022).

Climate change is already evident. It will continue to affect air temperatures, precipitation, and wind patterns in the Pacific Northwest (ISAB 2007; Philip et al. 2021), resulting in increased droughts and wildfires and variation in river flow patterns. These conditions differ from those under which native anadromous and resident fishes evolved and will likely increase risks posed by invasive species and altered food webs. The frequency, magnitude, and duration of elevated water temperature events have increased with climate change and are exacerbated by the Columbia River hydrosystem (EPA 2021a; 2021b; Scott 2020). Thermal gradients (i.e., rapid change to elevated water temperatures) encountered while passing dams via fish ladders can slow, reduce, or altogether stop the upstream movements of migrating salmon and steelhead (e.g., Caudill et al. 2013). Additional thermal loading occurs when mainstem reservoirs act as a heat trap due to upstream inputs and solar irradiation over their increased water surface area

(EPA 2021a, 2021b, 2021c). Consider the example of adult sockeye salmon in 2015, when high summer water temperatures contributed to extremely high losses of Columbia River and Snake River stocks during passage through the mainstem Columbia and Snake River (Crozier et al. 2020), and through tributaries such as the Salmon and Okanogan rivers, below their spawning areas. Some stocks are already experiencing lethal thermal barriers during a portion of their adult migration. The effects of longer or more severe thermal barriers in the future could be catastrophic. For example, Bowerman et al. (2021) concluded that climate change will likely increase the factors contributing to prespawn mortality of Chinook salmon across the entire Columbia River basin.

Columbia River basin salmon and steelhead spend a significant portion of their life-cycle in the ocean, and as such the ocean is a critically important habitat influencing their abundance and productivity. Climate change is also altering marine environments used by Columbia River basin salmon and steelhead. This includes increased frequency and magnitude of marine heatwaves, changes to the intensity and timing of coastal upwelling, increased frequency of hypoxia (low oxygen) events, and ocean acidification. These factors are already reducing, and are expected to continue reducing, ocean productivity for salmon and steelhead. This does not mean the ocean is getting worse every year, or that there will not be periods of good ocean conditions for salmon and steelhead. In fact, near-shore conditions off the Oregon and Washington coasts were considered good in 2021 (NOAA 2022). However, the magnitude, frequency, and duration of downturns in marine conditions are expected to increase over time due to climate change. Any long-term effects of the stressors that fish experience during freshwater stages that do not manifest until the marine environment will be amplified by the less-hospitable conditions there due to climate change. Together with increased variation in freshwater conditions, these downturns will further impair the abundance, productivity, spatial structure, and diversity of the region's native salmon and steelhead stocks (ISAB 2007; Isaak et al. 2018). As such, these climate dynamics will reduce fish survival through direct and indirect impacts at all life stages (NOAA 2022).

All habitats used by Pacific salmon and steelhead will be affected by climate dynamics. However, the impacts and certainty of the changes will likely vary by habitat type. Some changes affect salmon at all life stages in all habitats (e.g., increasing temperature), while others are habitat-specific (e.g., stream-flow variation in freshwater, sea-level rise in estuaries, upwelling in the ocean). How climate change will affect each individual salmon or steelhead stock also varies widely, depending on the extent and rate of change and the unique life-history characteristics of different natural populations (Crozier et al. 2008; Crozier and Siegel 2023). The continued persistence of salmon and steelhead in the Columbia basin relies on restoration actions that enhance climate resilience (Jorgensen et al. 2021) in freshwater spawning, rearing, and migratory habitats, including access to high elevation, high quality cold-water habitats, and the reconnection of floodplain habitats across the interior Columbia River basin.

### **2.8. Action Area**

"Action area" means all areas to be affected directly or indirectly by the Federal action and not merely the immediate area involved in the action (50 CFR 402.02). The proposed action occurs on private (820 acres) (including approximately 535 acres of patented mining claims owned or controlled by Perpetua), public lands administered by the USFS (2,372 acres) and BOR (37

acres), the State of Idaho (62 acres) in Valley County, Idaho (Figure 1), and Lemhi County, Idaho (Figure 2). Figure 25 displays the Stibnite component of the action area, which reflects activities directly associated with the SGP. The action area is the area of overlap between the effects of the action and the distribution or presence of ESA-species and/or designated critical habitat, which includes portions of the SFSR subbasin (HUC 17060208) and Zeph Creek-Lemhi River subwatershed (HUC 170602040702).

The action area is comprised of two discontiguous areas. Within the SFSR subbasin, the action area includes the access routes and associated infrastructure such as borrow pits and maintenance facilities (Figure 3), the Stibnite project area (Figure 4), OSV route and associated equipment storage and parking areas (Figure 7), transmission line and associated infrastructure (Figure 8) and the BRGI investigation sites and access routes (Figure 17). The major streams in the action area include the SFSR (downstream from the Warm Lake Highway Bridge to its confluence with the Salmon River); Johnson Creek (downstream from the FR447 Bridge); the EFSFSR (from its headwaters downstream to its confluence with the SFSR), and Sugar Creek (from just upstream of West End Creek to its mouth). Tributaries to these major streams that are either directly impacted by ground disturbing activities (e.g., stream crossing work) or receive stormwater runoff are also part of the action area. Because of the flow- and water quality-related (contaminants) effects, the action area extends from the project area downstream to the mouth of the SFSR. The downstream extent of the action area ends at the mouth of the SFSR because water quality-related effects causally linked to the proposed action are not expected to occur downstream of the confluence with the Salmon River given increased flows, distance from the project area, contaminant fate and transport mechanisms, and existing variability in contaminant concentrations (resulting from other natural and anthropogenic sources, such as atmospheric deposition mercury).

The second component of the action area reflects the Lemhi restoration portion of the proposed action, and is located entirely on private lands on the Upper Lemhi River approximately 12 miles northwest of Leadore, Idaho. The reach of the Lemhi Project extends approximately 7,000 ft. from river mile (RM) 42.62 to RM 41.32 on the mainstem of the Lemhi River on the western side of SH 28. Figure 2 displays the action area for this portion of the project, defined as the Lemhi Project reach plus 100-ft. buffers upstream and to the southwest of the project, 328 ft. (100 meters) downstream of the project, and out to SH 28 to the northeast of the project) for the Proposed Action.



**Figure 25. Stibnite Action Area, Stibnite Gold Project.**

### **2.9. Environmental Baseline**

The "environmental baseline" refers to the condition of the listed species or its DCH in the action area, without the consequences to the listed species or DCH caused by the proposed action. The environmental baseline includes the past and present impacts of all federal, state, or private actions and other human activities in the action area, the anticipated impacts of all proposed federal projects in the action area that have already undergone formal or early section 7 consultations, and the impact of state or private actions which are contemporaneous with the consultation in process. The impacts to listed species or DCH from federal agency activities or existing Federal agency facilities that are not within the agency's discretion to modify are part of the environmental baseline (50 CFR 402.02).

The action area is used by all freshwater life history stages of SR spring/summer Chinook salmon and SR Basin steelhead. Streams within the action area are DCH for each of these species. The condition of the listed species and DCH in the action area are described further below. Because climate change has already had impacts across the Snake River basin, discussions of the status of the species, status of critical habitat, and environmental baseline within the action area incorporate the current effects of climate change.

# **2.9.1. Condition of Species in the Action Area**

All life stages of SR spring/summer Chinook salmon and SR Basin steelhead have potential to be exposed to the effects of the proposed action. The following sections provide a summary of the current status and importance of populations within the action area to the recovery of these species.

# **SR Spring/summer Chinook Salmon**

2.9.1.1.1. SFSR and EFSFSR spring/summer Chinook salmon in the SFSR basin are part of the SFSR MPG. Two populations, the SFSR and EFSFSR populations, occur within the action area. Viable populations within the MPG should include some populations classified as "Very Large'" or "Large," and "Intermediate" reflecting proportions historically present. The SFSR population is classified as a large-size population, has hatchery influence (hatchery supplementation began in the mid-1970s), and is proposed to achieve a viable status in order to support recovery of the ESU (NMFS 2017). The EFSFSR population is also a large-size population, has hatchery influence (hatchery supplementation began in the 1998), and is proposed to achieve a maintained status in order to support recovery of the species (NMFS 2017). Both populations are currently at a high risk of extinction within the next 100 years based on information available for the 2022 5 year review (NMFS 2022b) and the viability assessment completed for Pacific salmon (Ford 2022) as part of the 5-year review effort.

Population-specific habitat concerns for the SFSR MPG, which includes the EFSFSR and SFSR populations, include: fine sediment; high stream temperatures; passage barriers; and wildfires affecting riparian zones (PNF 2020, as cited in NMFS 2022b). Although levels of fine sediment in the action area are often influenced by wildfires and rain-on-snow landslides, the PNF has not observed recent spikes in sediment levels at long-term monitoring sites.

Some of the recommended recovery strategies for this MPG include maintaining wilderness protection; providing or improving fish passage to and from areas with high intrinsic potential through barrier removal, screening and other projects; reducing or preventing sediment delivery by improving road systems, riparian communities, and rehabilitating abandoned mine sites; and improving riparian and floodplain health and function (NMFS 2017). Excess sediment, floodplain connectivity, poor water quality, and high-water temperatures are limiting factors that both of these populations share. Other limiting factors include passage barriers (EFSFSR population), channel alteration (SFSR population), and degraded riparian habitat (EFSFSR population).

Within the action area, Chinook salmon are known to spawn and rear in the SFSR, the EFSFSR, Johnson Creek, Cabin Creek, Sugar Creek, and lower reaches of Burntlog Creek, and Meadow Creek (when adults are outplanted). Chinook salmon also use the lower 500 ft. of EFMC, but no Chinook salmon occur in Hennessy, Midnight, Fiddle, or Garnet Creeks. Juvenile Chinook salmon have also been found in the lower reaches of Trapper and Riordan Creeks.

Chinook salmon in the upper EFSFSR (upstream from the YPP) were extirpated by diversion of the EFSFSR into a bypass tunnel from mining operations in the late 1930s. After cessation of mining and abandonment of the bypass tunnel, a high gradient and impassable riffle/cascade on the EFSFSR flowing directly into the YPP continued to prevent fish passage into the upper watershed. Both the riffle/cascade and the YPP were created as a result of mining operations. Chinook salmon use the SFSR and the mainstem of the EFSFSR downstream from the YPP as a migratory corridor and for spawning and rearing. Adult Chinook salmon and salmon redds have been observed in Sugar Creek (48 redds counted from 2014-2022) and in the EFSFSR as far upstream as the inlet of the YPP (i.e., downstream from the impassable cascade).

Adult Chinook salmon have been periodically introduced into Meadow Creek and the upper EFSFSR, upstream from the YPP, by the Idaho Department of Fish and Game (IDFG) in cooperation with the Nez Perce Tribe (NPT) when there is sufficient overstock from the local hatchery. These releases began in 2011, although no releases occurred between 2018 and 2021. However, in August and September, 2022, the NPT and IDFG released 387 adult Chinook salmon into the EFSFSR, just below the Meadow Creek confluence.

Since supplementation began, some adult Chinook salmon have returned to spawn in the EFSFSR, but are not able to migrate beyond the cascade that exists upstream of the YPP. The Chinook salmon transported upstream from the YPP, although introduced, are part of the ESAlisted population. The juveniles migrate downstream naturally and mix with the juveniles spawned downstream from the YPP.

The NPT surveyed stream segments above the YPP where redds were historically documented and identified 29 redds in Meadow Creek and three redds in the EFSFSR between Fiddle Creek and Meadow Creek. The PNF surveyed the EFSFSR reach between the box culvert and Meadow Creek and identified one redd. The NPT and IDFG have also conducted annual Chinook salmon redd surveys in the EFSFSR and its tributaries below the YPP. Over a 20-year period, 38 redds have been observed between the YPP and the road crossing upstream from Sugar Creek. Johnson Creek had the highest numbers of Chinook salmon redd counts in the Upper EFSFSR watershed, ranging from 193 (2008, 2011) to 376 (2014), with an average count of 207 redds per year.

In 2018 and 2019, a total of 55 juvenile Chinook salmon were captured during sampling efforts in the YPP (Brown and Caldwell 2020b). Forty of these fish were PIT tagged, of which 27 had associated observation details in the Columbia Basin PIT Tag Information System (PSMFC 2024). Fish were documented migrating downstream in April and May the following year. Two fish were detected at McNary Dam, and one fish was detected in the Lower Columbia River estuary. None of these tagged fish were detected as returning adults (PSMFC 2024).

The 2017 NMFS Recovery Plan (NMFS 2017) identified recovery strategies for SR spring/summer Chinook salmon for the Lower EFSFSR and Upper EFSFSR watersheds (mine site location) including:

- Maintain current wilderness protection and protect pristine tributary habitat;
- Provide/improve passage to and from areas with high intrinsic potential through barrier removal;
- Reduce and prevent sediment delivery to streams by improving road systems and riparian communities, and rehabilitating abandoned mine sites; and,
- Manage risks from tributary fisheries according to an abundance-based schedule.

# *2.9.1.1.2. Lemhi*

All life stages of Chinook salmon use the Lemhi River in the action area. Most spawning occurs in the Lemhi River upstream from Hayden Creek (which includes the restoration site) and in the Hayden Creek drainage. Historically there were at least seven other tributaries providing spawning habitat for Chinook salmon, but now spawning has been reduced to just Hayden Creek and the mainstem.

Within the Upper Salmon River MPG, the Lemhi population of SR spring/summer Chinook salmon occurs within the Lemhi portion of the action area. The Lemhi population is a very largesize population, has low hatchery influence, and is proposed to achieve a viable status in order to support recovery of the ESU (NMFS 2017). A high spatial structure/diversity risk rating for the Lemhi River population is driven by a substantial loss of access to tributary spawning and rearing habitats, and the associated reduction in life-history diversity.

Habitat concerns for the Upper Salmon River MPG include low flows from water diversions; degraded riparian conditions (losses from agriculture and overgrazing by livestock); sediment input from grazing and agricultural practices; and high-water temperatures (NMFS 2022b). Since 2016, summer instream flows have improved, tributary connectivity to the mainstem has improved, floodplain and habitat complexity has increased, and grazing management has been altered to protect riparian habitat. Overall, work in the Lemhi River basin between 2007 and 2019 has increased the summer rearing capacity for parr by 62 percent, and researchers have reported an increase in juvenile Chinook salmon productivity (Uthe et al. 2017; Haskell et al. 2019).

Juvenile survival and production data indicate that spawning and rearing PBFs will have to improve for the Lemhi River Chinook salmon population to achieve maintained status (NMFS 2017). The population is currently at a high risk of extinction within the next 100 years based on information available for the 2022 5-year review (NMFS 2022b) and the viability assessment completed for Pacific salmon (Ford 2022) as part of the 5-year review effort, remaining at high risk because of spatial structure loss.

An important opportunity to advancing the recovery of the Upper Salmon River MPG is to increase habitat complexity by creating multi-threaded channels, increase rearing habitat by increasing floodplain connectivity, reducing the width-to-depth ratio, stabilize banks, and improve willow-dominated riparian areas, as well as maintaining and improving instream flows and tributary connectivity.

### **SR Basin Steelhead**

### *2.9.1.2.1. SFSR*

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The proposed action would affect individuals in the Salmon River MPG, specifically the SFSR steelhead population. This population is one of the few that has never been supplemented with hatchery fish and has high proportions of B-run<sup>7</sup> individuals. The SFSR population is currently at a moderate risk of extinction within the next 100 years based on information available for the 2022 5-year review (NMFS 2022c) and the viability assessment completed for Pacific salmon (Ford 2022) as part of the last 5-year review effort. The SFSR population is targeted to achieve a viable status (low risk of extinction). The 5-year geometric mean for the SFSR and Secesh River populations (combined) has steadily decreased since 2010, decreasing by 57 percent over the two most recent five-year periods (Ford 2022). Limiting factors include excess sediment, migration barriers, and degraded riparian conditions. The recovery strategy emphasizes reducing and stabilizing disturbed areas and improving and rehabilitating roads as actions for reducing sediment delivery to spawning and rearing stream reaches.

The SFSR population spawns, rears, and migrates through the action area. Steelhead spawning overlaps many of the mainstem areas used by Chinook salmon, and steelhead redds have been observed in smaller tributaries such as Camp and Fitsum Creeks (Thurow 1987). In the EFSFSR, Snake River Basin steelhead occur up to YPP, where the impassable falls/cascade (30 percent gradient) blocks upstream migration. Due to their spawn timing, spawning surveys are not typically performed; therefore, where spawning occurs in the watershed is not well documented. Steelhead redds and adults were identified in 2004 downstream from the town of Yellow Pine. Most of the observed spawning sites were in small pockets of suitable substrate, often in marginal positions rather than in well-developed spawning riffles (Nelson 2004). Some steelhead also spawn in the EFSFSR upstream from the town of Yellow Pine. Within the action area, steelhead spawning occurs in Johnson Creek, Burntlog Creek, the SFSR, Sugar Creek, and the EFSFSR below the YPP - coincident with the upstream endpoint for DCH in the EFSFSR. There

<sup>&</sup>lt;sup>7</sup> Steelhead referred to as "B-run" are larger ( $> 78$  cm long), spend two years in the ocean, and appear to begin their upriver freshwater migration later in the year than steelhead referred to as "A-run." Steelhead referred to as "A-run" are smaller (usually 58 to 66 cm long), spend one year in the ocean, and begin their upriver freshwater migration earlier in the year than steelhead referred to as "B-run" (NMFS 2017).

is no recreational fishery for steelhead in the SFSR nor is the population supplemented with hatchery-produced fish.

Within the action area, steelhead are known to spawn and rear in the SFSR, the EFSFSR, Johnson Creek, Cabin Creek, and Sugar Creek. No steelhead occur in Meadow, East Fork Meadow, Hennessy, Midnight, Fiddle, or Garnet Creeks.

Although the BA indicates that streams along the Burntlog route do not support steelhead, eDNA samples indicate the presence of *O. mykiss* in upper Burntlog, East Fork Burntlog, Peanut, Trapper, and Riordan Creeks. Stantec (2024) postulated that steelhead may reach the headwaters in Johnson Creek, recognizing that there is no data documentation to verify if they are steelhead or resident *O. mykiss*. They further suggested that it is unlikely that steelhead occur in upper Trapper Creek, its tributaries, or in upper Riordan Creek due to suspected passage barriers in the lower reaches of these creeks. NMFS agrees with this characterization for Riordan and Trapper Creeks. However, because no barriers are suspected to occur in upper Burntlog Creek, East Fork Burntlog, or Peanut Creek, NMFS considers *O. mykiss* in these remaining streams as steelhead in our analysis.

Fish tissue samples and eDNA have been sampled from various locations upstream from the YPP from 2014 to 2016, and two fish tested positive for *O. mykiss* DNA, one in Meadow Creek Lake and one in the EFMC. The BA authors believe that the rainbow trout genetics detected from these locations are, in fact, California golden trout (*O. mykiss aguabonita*), a subspecies of rainbow trout that were stocked in Meadow Creek Lake by IDFG that are not native to the region. NMFS agrees with this characterization. Otherwise, no *O. mykiss* have been observed in the EFSFSR upstream from the YPP during the aquatic baseline surveys conducted since 2012, but some have been observed in the YPP and downstream from Sugar Creek (BA Table 3-4) (Brown and Caldwell 2019; MWH 2017).

The 2017 NMFS Recovery Plan for Snake River Idaho Spring/Summer Chinook Salmon and Snake River Basin Steelhead Populations included recovery strategies for Salmon River steelhead. Priorities for the SFSR steelhead population, specific to the EFSFSR watershed include:

- Collect and analyze population-specific data to accurately determine population status;
- Maintain wilderness protection and protect pristine tributary habitat;
- Eliminate artificial passage barriers and improve connectivity to historical habitat;
- Reduce and prevent sediment delivery to streams by rehabilitating roads and mining sites; and,
- Manage risks from tributary fisheries through updated Fisheries Management Evaluation Plans and Tribal Resource Management Plans according to an abundance-based schedule.

Based on NMFS' 2022 5-Year Review (NMFS 2022c), recommended actions specific to SR Basin steelhead in the EFSFSR include improving water quality by reclaiming abandoned mine sites such as the Cinnabar Mine; continuing to conduct appropriate road maintenance, road obliteration, road relocation, and road resurfacing; improving riparian conditions in disturbed areas; eliminating passage barriers; restoring floodplains; and improving planning for potential

climate change effects by continuing to monitor stream temperature and validate fish distribution in modeled cold water refugia.

### *2.9.1.2.2. Lemhi*

All life stages of SR Basin steelhead use the Lemhi River within the action area. Steelhead within the Lemhi portion of the action area are in the Salmon River MPG, Lemhi population. The Lemhi population is a very large-size population, has hatchery influence, and is proposed to achieve a viable status in order to support recovery of the ESU (NMFS 2017). The population is currently at a moderate risk of extinction within the next 100 years based on information available for the 2022 5-year review (NMFS 2022c) and the viability assessment completed for Pacific salmon (Ford 2022) as part of the 5-year review effort.

Recommended actions for the Lemhi River population include improving riparian conditions in disturbed areas, eliminating passage barriers, and increasing winter juvenile rearing habitat by increasing floodplain connectivity and complex habitat structure; reducing W:D, increasing lowto zero-velocity pool habitat with cover, providing more side channel and multi-threaded channel habitat, and reducing the fine sediment delivery to streams (BioMark et al. 2019; NMFS 2022c). Specifically, in the upper Lemhi River, recommended actions to improve conditions for steelhead include increasing habitat complexity by creating multi-threaded channels, narrow the W:D, stabilize banks, increase willow-dominated riparian areas, maintain and improve instream flow, and improve tributary stream connections to the to the mainstem Lemhi River (Biomark et al. 2019; NMFS 2022c).

### **2.9.2. Condition of Critical Habitat in the Action Area**

The SGP portion of the action is located in headwater tributaries of the SFSR, with proposed mining activities to take place primarily in areas previously disturbed by mining beginning around 1919. Excavation of the YPP began in 1938, and volitional passage of Chinook salmon and steelhead to the EFSFSR and Meadow Creek was eliminated at that time. Additional mining continued at the site in the 1940s, 1970s, 1980s, and 1990s. Mining, milling, smelting, and leaching activities left behind impacts including underground mine workings, multiple open pits, development rock dumps, mill tailings deposits, cyanidation heap leach pads, neutralized (spent) heap leach ore piles, a mill and smelter site, three town sites, camp sites, a washed-out earthen dam (with its associated erosion and downstream sedimentation), haul roads, an abandoned water diversion tunnel, an airstrip, and other disturbances.

Past mining activities have resulted in ongoing releases of contaminants to surface water and groundwater at the site including elevated concentrations of antimony, arsenic, copper, lead, mercury, and cyanide. Most notable are elevated concentrations of arsenic and antimony. Past mining activities have also caused alterations to stream configurations and habitat including formation of the YPP lake, creation of a fish passage barrier immediately upstream of the YPP lake, sediment and tailings deposits, development rock dumps, and channel diversions.

#### **South Fork Salmon River**

Streams within the action area are DCH for both SR spring/summer Chinook salmon and SR Basin steelhead. The SFSR and its tributaries offer a large amount of suitable spawning and rearing habitat. The majority of land in the lower SFSR and EFSFSR watersheds is federally managed. Historically, the area was impacted by logging, mining, grazing, and road building. Grazing no longer occurs in the action area, and prior to the proposed action, mining in the action area has not been as prevalent as it once was. Logging rarely occurs, and has most recently been performed as post-fire salvage or when reducing hazard fuels. In more recent times, wildfire has become the largest disturbance mechanism in the SFSR subbasin. Recreation and use of the existing road system are the primary human activities in the action area, although some private inholdings and associated homesteads exist. The existing network of roads and trails continue to impact aquatic habitat conditions.

Within the Stibnite project area, steelhead DCH includes the SFSR, Cabin Creek, the EFSFSR upstream to the YPP, Sugar Creek, Johnson Creek, Burntlog Creek, and lower Riordan and Trapper Creeks. Chinook salmon DCH overlaps steelhead critical habitat in each of these rivers and streams, but also extends further upstream of the YPP to include reaches of the EFSFSR and Meadow Creek.

Habitat conditions within the action area have been significantly impacted by mining activities. Open pit mining activities began upstream of the EFSFSR and Sugar Creek confluence in 1938 and continued for 14 years. Upstream fish passage was eliminated when the EFSFSR was initially diverted around the mining area. In order to expand and deepen the YPP, the EFSFSR was diverted through the Bradley Tunnel to Sugar Creek in 1943 (Hogen 2002; Midas Gold2016). When mining ceased in 1952, the Bradley Tunnel was abandoned and the EFSFSR was allowed to flow into the abandoned pit, which was 450 ft. deep. Over time, sediment transported downstream from the watershed settled in the YPP Lake. Now, the YPP Lake is approximately 5 acres in size and averages about 30 ft. deep. It is predominantly surrounded by steep, unnatural shorelines created by historical mining operations. Very little vegetation exists on the hillside and shoreline. A long, steep riffle/cascade, with a gradient of about 30 percent, exists at the inlet of the lake. This cascade is an impassible barrier to Chinook salmon, and is likely an impassible barrier to steelhead under most flows. Overtime, an alluvial fan has formed at the base of this cascade and Chinook salmon have been documented spawning at the inlet to the lake.

The BA uses the Southwest Idaho Ecogroup Matrix of Pathways and Watershed Condition Indicators (Matrix) (USFS 2003 and USFS 2010a) as a tool for assessing environmental baseline conditions and evaluating the potential effects of an action on WCIs. A WCI is a particular aquatic, riparian, or hydrologic measure that is relevant to the conservation of ESA-listed salmonids. In some instances, a WCI is synonymous to a PBF, temperature being a prime example. In other instances, many WCIs comprise a PBF. For example, the LWD, pool frequency and quality, large pools/pool quality, and off-channel habitat WCIs provide insight into the natural cover and cover/shelter features of spawning, rearing, and migration areas.

The WCIs are described in terms of their functionality, that is, functioning appropriately (FA), functioning at risk (FAR), or functioning at unacceptable risk (FUR). A watershed comprised of WCIs that are FA is considered to be meeting the biological requirements of listed anadromous species (whereas WCIs that are FAR or FUR suggest that the relevant PBF is not in a condition that is suitable for conservation).

Over the past 20 years, various fish and aquatic habitat studies have been conducted in the SFSR subbasin which have provided a better understanding of aquatic resource baseline conditions within the analysis area. Studies have been conducted by federal, state, local, and tribal agencies (e.g., PNF, BNF, IDFG, and the NPT), as well as private entities (e.g., Perpetua).

The Stibnite portion of action area includes all of the watercourses (i.e., streams and rivers) and waterbodies (i.e., lakes, reservoirs) in the 12-digit hydrologic unit code (HUC) subwatersheds that overlap with the effects of the SGP. Because the majority of the activities and disturbance will occur at the mine site, which is located in the EFSFSR subbasin, the baseline condition within this subbasin is a primary focus. Relevant habitat conditions in other watersheds, and subwatersheds that may be impacted by SGP activities are also described, as appropriate. Chinook salmon use only the lower 500 ft. of EFMC, but no anadromous fish occur in Hennessy, Midnight, Fiddle, or Garnet Creeks. Baseline conditions for these subwatersheds can be found in section 4.1.2 of the BA (Stantec 2024), which is incorporated by reference here.

### **East Fork South Fork Salmon River**

The EFSFSR watershed covers approximately 250,000 acres and enters the mainstem SFSR near the confluence of the Secesh River. Most of the watershed is administered by the USFS and managed by the PNF and BNF. Private land in the watershed includes small parcels of land along Johnson Creek, large legacy mines in the headwater drainages (e.g., Stibnite and Cinnabar mines), and the town of Yellow Pine. Predominant historical land uses occurring in this watershed include timber harvest and large-scale mining (Wagoner and Burns 2001). Extensive cattle grazing also historically occurred in the Johnson Creek watershed, but federal grazing allotments have now been retired and grazing has been reduced to private lands.

Large-scale historical mining altered stream channel conditions in the Upper EFSFSR watershed. The USFS and mine operators have since undertaken restoration work. However, habitat for migratory salmonids in the EFSFSR upstream from the YPP lake is inaccessible because historical mining excavation of the stream channel has created a gradient barrier (YPP lake cascade). While mining in the area has not occurred since about 1950, there are still significant legacy effects that continue to impact channel conditions and fish populations.

Although not formally identified as a CERCLA Superfund site, some mining operators at the site conducted activities to reduce the release of hazardous substances before 2001. Notable work included diverting Meadow Creek and stabilizing the Bradley Tailings/SODA disposal area, which was completed in 1999. In addition, the USFS began using its CERCLA authorities to address legacy mining impacts at the site since 2001. In 2002, the USFS removed tailings from a pond and soils located at the former smelter stack area, and the Meadow Creek floodplain was reconstructed in the former pond area. In 2004 and 2005, the USFS reconstructed Meadow Creek directly downstream of Smelter Flats. This included the removal of tailings from the channel and depositing this material in a new containment cell located on the SODA. The new channel banks were revegetated with willow plants and the old channel was backfilled and reclaimed. In 2009,

the USFS regraded and covered a portion of the remaining tailings at Smelter Flats to prevent further erosion and exposure risk. It is in these restored stream reaches, that surplus hatchery Chinook salmon have periodically been placed by the NPT and IDFG and have successfully spawned and reared for the first time in decades.

The EFSFSR drainage has the lowest quality habitat for sensitive and protected fish in the SFSR subbasin (NPCC 2004). Primary habitat limitations in the EFSFSR drainage are reduced riparian habitat and decreased streambank stability due both to road design and the extent of the existing road system; secondary limitations include reduced instream LWD, water quality degradation, and fish passage barriers resulting from legacy mining in the area (NPCC 2004).

All IDEQ-inventoried waterbodies at the proposed mine site (except for West End Creek) are listed under Section 303(d) of the federal CWA as "impaired" due to water quality (IDEQ 2018). The causes for listing of these waters are associated with elevated concentrations of arsenic, antimony, and mercury. Each of the 303(d)-listed waterbodies has designated beneficial uses of "cold water communities," "salmonid spawning," and "primary contact recreation," and all (except Sugar Creek) have designated beneficial uses of "drinking water supply."

Wildfires have eliminated much of the tree canopy at the SGP mine site and vicinity, which has resulted in erosion and landslides. In addition, the failure of a dam on the EFMC in 1965 resulted in extensive erosion, both upstream and downstream from the former dam and reservoir site, which in turn has led to extensive and ongoing deposition of sediment in the lower reaches of Meadow Creek and downstream in the EFSFSR. Currently, while concentrations of total suspended solids and turbidity are low during some months, there is seasonal variation in these concentrations associated with high flow periods when concentrations can reach moderate to high levels.

The upper EFSFSR subwatershed is considered a priority subwatershed to the Forest's Aquatic Conservation Strategy (ACS). According to the PNF Land and Resource Management Plan, the ACS is intended to provide guidance towards long-term maintenance and restoration of characteristics found in healthy, functioning watersheds, riparian areas, and associated fish habitat. The ACS provides a scientific basis for protecting and restoring aquatic ecosystems; assisting towards the short- and long-term recovery of listed fish species; delisting of water quality impaired waters; and resource management. Priority subwatersheds are those that provide a pattern of protection/restoration across the Forest that will help lead towards the ACS goals (e.g., recovery of listed fish species, delisting of water quality impaired waters).

Habitat is generally "FAR" in the EFSFSR watershed (Table 30). Only intragravel quality, interstitial sediment deposition, LWD, W:D, and streambank condition are "FA" in the watershed. The remaining WCIs currently range from "FAR to FUR." Habitat conditions most likely to be affected by the proposed action are described in more detail below for action area watersheds used by ESA-listed Chinook and steelhead. More detail can be found for other WCIs and other subwatersheds in Appendix C of the BA (Stantec 2024).

Pathway	<b>Indicators</b>	<b>Upper EFSFSR</b> Watershed Functionality*	<b>Upper Johnson</b> <b>Creek Watershed</b> Functionality*
	Temperature	FA-FUR	FA-FAR
Water Quality	Intragravel Quality	FA	FA-FUR
	Chemical Contamination/Nutrients	<b>FUR</b>	FA
<b>Habitat Access</b>	Physical Barriers	<b>FUR</b>	FA-FUR
	Interstitial Sediment Deposition	FA	FA
<b>Habitat Elements</b>	<b>LWD</b>	FA	<b>FA</b>
	Pool Frequency	FA-FUR	FA
	Pool Quality	FAR	FA.
	Off-channel Habitat	<b>FAR</b>	<b>FA</b>
	Refugia	<b>FAR</b>	<b>FAR</b>
Channel Condition and Dynamics	W∶D	FA	FAR
	Streambank Condition	FA.	FAR
	Floodplain Connectivity	FAR	FAR
Flow/Hydrology	Change in Peak/Base Flows	<b>FAR</b>	Unknown
	Increase in Drainage Networks	FAR	<b>FUR</b>
Watershed Conditions	Road Density and Location	<b>FAR</b>	<b>FAR</b>
	Disturbance History	<b>FUR</b>	<b>FUR</b>
	RCAs	<b>FUR</b>	<b>FAR</b>
	Disturbance Regime	<b>FUR</b>	<b>FAR</b>

**Table 30. Summary Matrix of Watershed Condition Indicators in the Upper East Fork South Fork Salmon River and Upper Johnson Creek Watersheds.**

Functioning Appropriately = (FA), Functioning at Risk = (FAR) and Functioning at Unacceptable Risk = (FUR). \*See Southwest Idaho Matrix of Pathways and Indicators for explanation of functionality ratings (USFS 2010a and 2010b).

Habitat is generally "FAR" in the Upper Johnson Creek watershed (Table 30). Only chemical contamination/nutrients, interstitial sediment deposition, LWD, pool frequency, pool quality, and off-channel habitat are "FA" in the watershed. The remaining WCIs currently range from "FAR to FUR." Habitat conditions most likely to be affected by the proposed action are described in more detail below for action area watersheds used by ESA-listed Chinook and steelhead. More detail can be found for other WCIs and other subwatersheds in Appendix C of the BA (Stantec 2024).

# *2.9.2.2.1. EFSFSR*

The EFSFSR is a tributary to the SFSR. The EFSFSR between its confluence with Sugar Creek upstream to the YPP lake is 0.75 mile, and upstream to the confluence with Meadow Creek is 3.8 miles. This stream reach includes the YPP lake, immediately upstream from which is the high gradient cascade that is a complete barrier to upstream passage for all fish species including migrating Chinook salmon and steelhead. As previously mentioned, Chinook salmon also spawn and rear in the stream reach upstream from the lake in years that they have been introduced there by the IDFG. Downstream from the YPP lake, this stream reach is accessible to all life stages of Chinook and steelhead.

Between Meadow Creek and the YPP Lake, the EFSFSR widens and has larger streambed material (including abundant cobble and boulders), relative to the upper EFSFSR. This stream reach has moderate to high stream gradients (approximately 2 to 8 percent) (HDR 2016). Moving downstream to the confluence with Sugar Creek, the EFSFSR is similar in width, gradient, and substrate material as upstream, but many of the larger boulders and cobble are sharp and more angular. Based on field surveys conducted by Rio ASE (Rio ASE 2019), there are more and deeper pools upstream from the YPP Lake. The EFSFSR generally supports a healthy riparian corridor, with the exception of areas near the YPP Lake and areas of legacy mine waste dumps along the banks upstream and downstream from the YPP Lake.

The EFSFSR between Meadow Creek and Sugar Creek has been heavily impacted by legacy mining activities. In addition to the YPP Lake, a remnant of legacy mining activities, these impacts include waste rock dumps in and adjacent to the stream channel, tailings washed down from Meadow Creek valley, roads and infrastructure within and adjacent to the EFSFSR channel, dam construction across the EFSFSR main channel, and other legacy impacts (Midas Gold 2016). There are efforts currently underway by EPA to remove contaminated soils caused by historic mining on the EFSFSR, with one site just downstream from the YPP Lake, and two sites upstream of the YPP Lake. The uppermost site, which is located just downstream from the Meadow Creek confluence, was restored during the summer of 2023. Restoration included removal of contaminated tailings and mine waste located within channels and floodplain of the EFSFSR and its tributaries, and the diversion of surface water around mine wastes that were sources of metals to the stream, and improving both water quality and fish habitat complexity (NMFS No: WCRO-2022-03035).

Benthic macroinvertebrates (BMI) in the EFSFSR have been sampled for five years at five locations. All sample locations (one site downstream from Sugar Creek confluence, one site immediately upstream from Sugar Creek confluence, two sites below Meadow Creek, and one site upstream from Meadow Creek) are inhabited by a BMI community that is intolerant of organic pollution and generally poor water quality conditions (EcoAnalysts, Inc. and MWH Americas, Inc. 2017). Results also indicate there is little to no impact of metals on the macroinvertebrate community (i.e., relatively intolerant of metals contamination) at the sites. The macroinvertebrate community at the site below the YPP appears to be impacted by historic mining in the area, as the community composition at this site includes a larger proportion of filterers (e.g., blackfly larvae) and higher metals tolerant index (MTI) values. Other sites exhibiting potential decreases in water quality include an increase in MTI values in the EFSFSR below Sugar Creek (EcoAnalysts, Inc. and MWH Americas, Inc. 2017). *2.9.2.2.2. Yellow Pine Pit Lake*

During mining activities during the 1930s through the 1950s, the nearly 5-acre YPP lake was created by open-pit mining while the EFSFSR was diverted through the Bradley Tunnel to Sugar Creek (Hogen 2002). After mining ceased in 1952, the EFSFSR was allowed to flow through the abandoned mine pit. The pit currently has a maximum depth of approximately 36 ft. Diverting the EFSFSR back into the stream channel and pit created a long cascade with a high gradient cascade that precluded fish passage upstream into the upper watershed. Therefore, all streams upstream from the YPP Lake are inaccessible to anadromous Chinook salmon and steelhead without human intervention. The YPP Lake is used by both fish and mammals, including Chinook salmon, bull trout, and river otters. Mountain whitefish are abundant in the lake (Brown and Caldwell 2019; Brown and Caldwell 2020b).

The YPP Lake is the largest feature that affects flow rates in the EFSFSR; however, because of its small area, it affects low flows only slightly and does not affect high flows at all (Kuzis 1997). The lake also displays thermal stratification (i.e., order).

#### *2.9.2.2.3. Meadow Creek*

Meadow Creek, located within the upper EFSFSR subwatershed, is a major tributary to the EFSFSR that flows through a flat-bottomed valley surrounded by steep mountains. Elevations range from 3,937 ft. above sea level in the lower reach to over 7,546 ft. in the headwaters. Meadow Creek has been heavily impacted by legacy mining-related activities, including deposition of tailings and spent heap leach ore, ore processing facilities, heap leach pads, and other infrastructure, stream relocation into a straightened riprap channel, and construction of an airstrip (Midas Gold 2016). The downstream end of the valley shows remnant effects from early mining activities, along with a large outwash feature created by a dam failure in the EFMC. Portions of the creek have been modified over the years to improve conditions caused by past mine operations, including the regrading and revegetation of the 2 percent gradient lower reach of the creek in 2004 and 2005.

The middle reach of Meadow Creek is an engineered channel that was constructed to bypass the SODA. The channel was lined with riprap over geotextile fabric and is confined between reinforced/engineered slopes with a gradient of less than 2 percent. This reach has a short section with a 9 percent gradient, shallow depths, and few pools, which may be a partial fish migration barrier at low flows. The channel includes low-gradient riffles, glides, and runs. There is no sidechannel development or potential LWD recruitment.

Upper Meadow Creek encompasses the headwaters downstream to the location of proposed TSF Buttress. Upper Meadow Creek is confined and high gradient at the most upstream extent and low gradient and unconfined immediately upstream from the SODA in lower Meadow Creek, transitioning from a gradient of 4 to 8 percent to 2 to 4 percent. Habitat is composed of riffles, step runs, and pools. The presence of side channels in some portions provide potential for lateral channel movement in the less confined sections. Immediately upstream from the SODA, Meadow Creek is unconfined, with a gradient less than 1 percent. The reach is composed of lowgradient riffle, step run, and pool habitat. The floodplain is active with oxbow cutoffs, side channels, and backwater features.

BMI in Meadow Creek have been sampled for five years at three locations. All sample locations (two sites downstream from EFMC and one site in the rock engineered reach upstream from EFMC) are inhabited by a BMI community that is generally intolerant of organic pollution and generally poor water quality conditions; however, the BMI at these locations may be slightly
more stressed than those in the EFSFSR. Results also indicate there is little to no impact of metals on the macroinvertebrate community at the sites.

## *2.9.2.2.4. EFSFSR - Headwaters*

Upstream from the Meadow Creek confluence, the EFSFSR is characterized by narrower channels with moderate gradient (2 to 4 percent), transitioning to higher-gradient (4 to 8 percent) step-pool habitat farther upstream. Overall substrate size is generally smaller than downstream reaches, with sand, gravel, smaller cobble, and boulders. This reach of the EFSFSR has relatively abundant riparian vegetation and large amounts of LWD.

Kuzis (1997) found that the Headwaters EFSFSR displays evidence of a high sediment load, such as streambed aggradation (deposition of material), channel splitting, pool filling, and overbank deposits of fines. The combination of low-gradient, relatively wide valley, plentiful wood supply, and a high sediment supply have resulted in current channel conditions.

## *2.9.2.2.5. EFSFSR - between Sugar Creek and Profile Creek*

The EFSFSR downstream from Sugar Creek is adjacent to the SGP mine site in the No Mans-EFSFSR subwatershed. Stibnite Road (CR 50-412) closely parallels the EFSFSR along its entire length of the action area. The EFSFSR ranges from low-gradient habitat with pools to high gradient habitat with cascades. Substrate throughout the reach is variable, and dependent on the gradient, with the lower-gradient sections dominated by gravel and cobble, while the highergradient units are dominated by large cobble and boulders. Avalanches in 2014 have resulted in high concentrations of LWD in the EFSFSR downstream from Sugar Creek (MWH 2017). In April 2019, a series of avalanches and related landslides caused extensive damage to Stibnite Road (CR 50-412), and pushed snow, timber, and other debris into the EFSFSR (Midas Gold 2019a). These events were naturally occurring in burn areas and were related to rain-on-snow events.

### *2.9.2.2.6. Sugar Creek*

The Sugar Creek subwatershed is a priority ACS subwatershed. Sugar Creek flows into the EFSFSR downstream from the YPP. It is a relatively low-gradient stream. A gated maintenance USFS road closely parallels Sugar Creek for nearly 2 miles; however, this is a ML1 road, a road closed to public motorized use. Much of Sugar Creek has large aggregates of LWD. The dominant substrates are sand, gravel, and cobble. Sugar Creek has widened channels, and excessive medial and lateral bar formation in response to past sediment inputs. In the 1940s, approximately 1 million cubic yards of glacial overburden was removed from the EFSFSR channel and placed in both Sugar Creek and other parts of the EFSFSR (Kuzis 1997).

BMI in Sugar Creek have been sampled for five years at two locations. All sample locations (one site upstream from and one site downstream from West End Creek) are inhabited by a BMI community that is intolerant of organic pollution and generally poor water quality conditions. While data gathered by MWH (2017) suggests macroinvertebrate communities can be considered to be in good condition based on community composition metrics, Kraus et al. (2022) reported decreased insect biodiversity in Sugar Creek as a result of upstream historic mining impacts and associated mercury loading to the stream network.

## **Johnson Creek**

A portion of Johnson Creek flows through the Porcupine Creek-Johnson Creek subwatershed, which has been identified as an ACS priority subwatershed. Johnson Creek flows from its origin near Deadwood Summit approximately 32 miles to its confluence with the EFSFSR near the town of Yellow Pine. It is the largest tributary stream of the EFSFSR. The Johnson Creek watershed encompasses 213 miles. Major tributaries to Johnson Creek include Riordan, Trapper, Burntlog, Ditch, Halfway, Rustican, Sheep, Lunch, Landmark, Rock, and Boulder Creeks.

Roads are the foremost cause of soil erosion in the Johnson Creek drainage – much of Johnson Creek Road is next to Johnson Creek. Roads have contributed up to 90 percent of the management induced sediment. Landslides occur naturally, resulting in the addition of sediment, substrate, and LWD; however, the frequency and magnitude of landslide events have greatly increased due to the construction of roads (Rabe and Nelson 2010).

### *2.9.2.3.1. Burntlog Creek*

Burntlog Creek flows through the Lower and Upper Burntlog Creek subwatersheds, which have been identified as ACS priority subwatersheds. Burntlog Creek, a tributary of Johnson Creek, is a moderate-gradient stream that parallels Johnson Creek in the lowest reaches and occupies a steep valley floor in the upper reaches of the drainage. It is approximately 14.1 miles long, with approximately 3.2 miles upstream from the Burnt Log Road crossing. Tributaries of Burntlog Creek that cross the Burnt Log Road include East Fork Burntlog Creek and Peanut Creek. Burntlog Creek is characterized by a granitic watershed geology. Substrate samples measured at two locations sampled for six years show relatively low cobble embeddedness levels (around 8 percent) and multi-year estimates of free matrix estimates ranging from 51 to 56 percent (Stantec 2020); both cobble embeddedness and free matrix are considered FA at these survey sites. The upper reaches have moderate amounts of LWD from extensively burned areas, and minimal overhead canopy. The dominant substrates are sand, gravel and cobble.

### *2.9.2.3.2. Trapper Creek*

A portion of Trapper Creek flows through the Porcupine Creek-Johnson Creek subwatershed, which has been identified as an ACS priority subwatershed. Trapper Creek, a tributary of Johnson Creek, is approximately 9.2 miles long, with approximately 4.1 miles upstream from the existing Burnt Log Road crossing. Trapper Creek has a high gradient passage barrier approximately 0.75 miles upstream from its confluence with Johnson Creek. The downstream portion consists of large boulders and cascades. Trapper Creek is characterized by a granitic watershed geology. The lower reach of Trapper Creek has a low cobble embeddedness level (around 3 percent) and free matrix values are FA (Stantec 2020). The mid-section of Trapper Creek has an abundant LWD count (MWH 2017). The stream's dominant substrates are gravel and cobble.

### *2.9.2.3.3. Riordan Creek*

Riordan Creek, a tributary of Johnson Creek, is approximately 9.8 miles long with less than 0.12 miles upstream from the Burntlog Road crossing. Riordan Creek is a relatively low-gradient stream. Roughly halfway down the length of the stream is Riordan Lake. Downstream from the lake, Riordan Creek has a slightly higher gradient, particularly just before it enters Johnson Creek. A trail with bridges that are open to small off-road vehicles crosses the creek several times above and to the north side of the lake. The dominant substrates are sand and gravel.

# **Watershed Condition Indicators**

Not all WCIs are equal in terms of evaluating the potential impacts of the SGP within the mine site. Some baseline WCIs are of historical interest, some will not be affected by the SGP. For these reasons, NMFS selected WCIs that have the greatest potential to accurately identify potential impacts of the SGP on the ability of the PBFs to support all anadromous salmonid life stages. The five WCIs selected for detailed analysis in this opinion include:

- Water Temperature;
- Sediment/Turbidity;
- Chemical Contaminants;
- Physical Barriers; and,
- Change in Peak/Base Flows.

The conditions of each of these WCIs will be discussed in more detail below. Baseline conditions for all WCIs are presented in Appendix C of the BA.

### *2.9.2.4.1. Water Temperature*

Stream temperatures at the mine site are believed to be influenced by a combination of past mining activities, loss of shade from forest fires, as well as climate change (Baldwin and Etheridge 2019). Continuous temperature data has been collected since 2011 (Figure 26) (some gages have a longer period of record) at the following five USGS gages:

- Meadow Creek (upstream of historic mining impacts): USGS Gage 13310850
- EFSFSR (upstream of Meadow Creek): USGS Gage 13310800
- EFSFSR (downstream of Meadow Creek): USGS Gage 13311000
- EFSFSR (upstream of Sugar Creek): USGS Gage 13311250
- Sugar Creek (upstream of EFSFSR): USGS Gage 13311450

As shown in Figure 26, and based on data recently collected at the five gaging stations above, baseline water temperatures in the upper EFSFSR and Meadow Creek are fully supportive of spawning, rearing, and migration for Chinook salmon and steelhead. Water temperatures in the lower sites are generally supportive of juvenile rearing for Chinook salmon and steelhead. Stream temperatures in the EFSFSR below Meadow Creek typically don't drop below 13°C until after early September. Temperatures are slightly cooler downstream in the EFSFSR, dropping below 13°C slightly sooner – in late August. Sugar Creek exhibits similar temperature patterns as the EFSFSR below the YPP.

Climate change is expected to impact stream temperatures into the future. The NorWeST summer stream temperature model, produced by the USFS Rocky Mountain Research Station (Isaak et al. 2016), provides a variety of scenario-based predictions for mean August stream temperatures. These are presented alongside average August temperatures obtained at the USGS gaging stations at the mine site to provide information regarding the possibility of changing climate conditions in the analysis area (Table 31). Under baseline conditions, August mean stream temperatures may warm by about 1.5°C.





**Source:** Isaak et al. 2016; USGS 2024

**Key:** °C = degrees Celsius;

<sup>1</sup>This is the A1B warming trajectory.



**Figure 26. Seven Day Average Daily Maximum (7DADM) Temperatures in Degrees Celsius at Five USGS Gaging Stations at The Mine Site. Temperature Data Collected From 2011 Through 2024.**

#### *2.9.2.4.2. Sediment and Turbidity*

Substrate monitoring has occurred since 2010, with sites spread out across the Stibnite Analysis Area (BA Figure 4.1-1). Stantec (2024) measured these sites annually, so the data are best interpreted as a measure of annual, watershed scale conditions and trends, rather than sitespecific effects from point sources of sediment. Generally high embeddedness relative to reference conditions could indicate degraded conditions in a watershed, while low embeddedness indicate favorable conditions in a watershed.

Nelson and Burns (2005), Nelson et al. (2006), and Zurstadt et al. (2016), describe a method to rate the interstitial sediment deposition indicator. The rating system is used in this analysis to describe the current condition of the interstitial sediment deposition WCI in the analysis area. Background levels of sediment deposition are generally lower in areas dominated by non-granitic lithology (Nelson et al. 2006); however, the watershed geology within the analysis area is granitic.

The current existence, use, and maintenance of the Stibnite Road, Quartz Creek Road, and historical mining disturbance in the Stibnite area continue to be a source of existing and potential anthropogenic sediment to the EFSFSR. Because they occur in the same geology and have experienced similar weather and management activity, analysis area tributaries that lack data are expected to have embeddedness levels comparable with those measured in other tributaries.

The floods of 2008 and landslides in 2018 and 2019 deposited sediment into the EFSFSR and the sediment accumulations behind log jams and debris fans that were created are evident. However, it also may be that the influx of diverse particle sizes and LWD were more beneficial than deleterious because the system was deficient in LWD, and anadromous salmonid spawning sites were limited downstream of the town of Yellow Pine.

A variety of past disturbances at the SGP mine site affected streambank stability and erosion, and the proximity to existing roads have affected the sediment levels in the streams. However, substrate/sediment monitoring between 2012 and 2019 showed most monitoring locations within the mine site area at a FA level. Currently, following storm events, high levels of sediment are released from the EFMC, the largest single source of sediment in the watershed, resulting in increased turbidity.

### *2.9.2.4.3. Passage Barriers*

Fish passage barriers were identified and described within the SGP mine site (BioAnalysts 2021). BioAnalysts (2021) identified fish passage barriers, with five artificial barriers and one natural barrier in fish-bearing streams (BA Figure 4.1-2) (Stantec 2024). The status of these barriers was identified as either complete, meaning no fish species can pass at any time of year, or partial, meaning some or all fish may pass at moderate or high flows, but not at low flows. Artificial barriers can be attributed to various actions, for example, construction of culverts and stream alteration (BioAnalysts 2021). Table 32 presents the amount of total potential habitat upstream from each barrier for Chinook salmon and steelhead; those that do not have potential habitat for these species upstream from the barrier are not included in the table.

BioAnalysts (2021) identified major barriers to fish movement in the SGP mine site area including: (1) the high gradient cascade in the EFSFSR upstream from the YPP Lake; (2) EFSFSR box culvert; and (3) the high gradient cascade in Meadow Creek upstream from the confluence with the EFMC (at the downstream end of the engineered channel). The high gradient cascade in the EFSFSR upstream from the YPP Lake is a complete barrier to natural fish passage. The other two major barriers, the EFSFSR box culvert and Meadow Creek barriers, are flow-dependent partial barriers that can block seasonal migration, and only hinder migration of fish that reside in or were stocked upstream from the YPP Lake cascade barrier (i.e., translocated Chinook salmon).





Notes:

<sup>1</sup>·Not·all·of·the·Total·Habitat·is·considered·usable/accessible·habitat.¶

<sup>2</sup>·Results·based·on·accessible·Intrinsic·Potential·Habitat·for·early·life·stages·of·Chinook·salmon·and·steelhead. <sup>3</sup> Results based on usable/accessible Critical habitat for all life stages modeled for Chinook salmon.

Key: EFSFSR = East Fork South Fork Salmon River; YPP = Yellow Pine Pit

### *2.9.2.4.4. Chemical Contaminants*

Baseline water quality at the mine site is influenced by both natural mineralization and historical mining activity (Baldwin and Etheridge 2019). Locally, remnant features from historical mining include underground mine workings; multiple open pits; development rock dumps, piles, and tailings deposits; heap leach pads and spent heap leach ore piles; contaminated soils from the former mill and smelter sites; former surface water diversions, dams, townsites, and roads; and an abandoned water diversion tunnel (Midas Gold 2016). Detailed descriptions of existing chemical contaminant conditions in streams, sediments, and biota within the action area are

available in Midas Gold (2019b), HDR (2017), and MWH (2017). This section summarizes baseline contaminant conditions in the water column, sediments, and biota. Because the toxicity of many contaminants is dependent upon other water chemistry conditions, we have included a brief summary of major ions, pH, and TDS.

*Major Ions, pH, and TDS.* The average baseline major ion chemistry for the surface water assessment nodes (refer to Figure 4.1-3 in the BA [Stantec 2024]) is summarized in Table 33. The EFSFSR and Sugar Creek sampling locations each exhibit a calcium-magnesiumbicarbonate water type, meaning that calcium and magnesium are the dominant cations in solution, and bicarbonate is the dominant anion. The samples from Meadow Creek had on average a higher relative proportion of calcium and are therefore classified as calciumbicarbonate water.

Average TDS concentrations also were consistent in the Meadow Creek and EFSFSR sampling locations. The average TDS ranged from 56 to 57 mg/L in the Meadow Creek samples and appears to increase downstream in the EFSFSR from about 53 mg/L in the farthest upstream reach (YP-SR-10) to 67 mg/L in the downstream reaches. It appears that despite the higher TDS load in Sugar Creek (116 mg/L), the creek does not appreciably contribute to TDS concentrations in the EFSFSR, based on the similar average TDS concentrations obtained for the EFSFSR sampling points located just upstream (YP-SR-4) and downstream (YP-SR-2) of the Sugar Creek confluence.

Baseline samples from Fiddle Creek exhibited a slightly different water quality signature compared to the EFSFSR and Meadow Creek. Although Fiddle Creek is classified as a calciumbicarbonate water, the creek has a lower proportion of magnesium and a higher proportion of sodium compared to the other monitoring locations. It also has a lower proportion of sulfate and higher proportion of bicarbonate. Some of these differences may be due to the relatively low average TDS concentration observed in Fiddle Creek during the baseline monitoring period (36 mg/L). The low sulfate and TDS concentrations also could point to a lack of mineralized deposits and historical mining-related impacts in the Fiddle Creek drainage, and different lithologies in the catchment area, specifically calcareous rock formations.

West End Creek stands out as having the most notably different major ion signature among the surface water assessment nodes (BA Figure 4.1-3) (Stantec 2024). During the baseline period, West End Creek surface water exhibited a calcium-magnesium-bicarbonate-sulfate water type. With the exception of chloride and sodium, the West End Creek samples also had the highest major ion constituent concentrations among the surface water assessment nodes considered, with baseline sulfate and TDS concentrations averaging 57 and 209 mg/L, respectively. West End Creek sample point YP-T-6 is located downstream of both the upper and lower historical West End waste rock dumps; it is therefore possible that the water chemistry at this location has been influenced by the waste material, especially where the creek flows directly through historical development rock piles. Mapped metamorphic bedrock in the West End valley (including marble, quartzite, and schist) in contrast to granitic batholith rocks in the EFSFSR drainage also may affect the stream chemistry, as these rock types locally tend to produce higher TDS and alkalinity (SRK 2021a).

<b>Node</b>	<b>Stream</b>	# of Samples	pH	<b>Hardness</b> as CaCO <sub>3</sub>	<b>Bicarbonate</b> as $CaCO3$			Calcium Chloride Magnesium Potassium		Sodium	<b>Sulfate</b>	<b>TDS</b>	Water <b>Type</b>
$YP-T-27$	Meadow Creek	45	7.3	37.4	38.4	11.5	1.25	2.13	0.87	2.44	5.97	57	Calcium- bicarbonate
$YP-T-22$	Meadow Creek	45	7.4	37.5	39.5	11.3	1.00	2.18	0.84	2.42	5.16	56	Calcium- bicarbonate
<b>YP-SR-10</b>	<b>EFSFSR</b>	45	7.4	35.3	38.7	10.3	0.63	2.25	0.78	2.12	4.15	53	Calcium- magnesium bicarbonate
YP-SR-8	<b>EFSFSR</b>	45	7.5	39.1	42.2	11.4	0.73	2.55	0.83	2.36	6.77	60	Calcium- magnesium bicarbonate
YP-SR-6	<b>EFSFSR</b>	45	7.4	39.0	40.3	11.4	0.68	2.54	0.83	2.34	6.44	58	Calcium- magnesium bicarbonate
YP-SR-4	<b>EFSFSR</b>	45	7.5	43.8	42.5	12.7	0.63	2.89	0.88	2.30	8.86	65	Calcium- magnesium bicarbonate
$YP-SR-2$	<b>EFSFSR</b>	45	7.6	48.4	48.1	14.4	0.52	3.01	0.85	2.31	9.31	67	Calcium- magnesium bicarbonate
$YP-T-6$	West End Creek	45	8.4	179	120	43.1	< 0.20	17.6	1.94	1.10	56.7	209	Calcium- magnesium bicarbonate -sulfate
$YP-T-1$	Sugar Creek	46	7.7	54.2	56.1	16.5	< 0.20	3.09	0.76	2.24	9.00	116	Calcium- magnesium bicarbonate

**Table 33. Average Major Ion Chemistry for Surface Water Assessment/Prediction Nodes (mg/L)**

**Source:** Data obtained from Midas Gold 2019b.

**Note:** Units are milligrams per liter except for pH, which is in standard units.

Values in the table represent the average of sample results collected between 2012 and 2018.

Average concentrations for calcium, magnesium, potassium, and sodium represent the dissolved fraction.

**Key:**  $\text{CaCO}_3 = \text{Calcium Carbonate}$ ;  $\text{EFSFSR} = \text{East Fork South Fork Salmon River}$ ; mg/L = milligrams per liter;

TDS = total dissolved solids.

Field-measured pH values for the surface water assessment nodes were generally in the range of 7 to 8 standard units. The highest average pH (8.4) was observed at West End Creek sample location YP-T-6. Elevated baseline pH measurements at this location are likely another indicator of the geochemical influence exerted by legacy waste rock material, natural mineralization, and the predominance of carbonate bedrock in the West End Creek drainage. Overall, the neutral to alkaline pH values observed in streams near the mine site show that the geochemistry of the natural mineralized deposits and the legacy mine materials is not conducive to acidic drainage.

*Contaminants of Concern.* The Surface Water Quality Baseline Study (HDR 2017) showed that most metals analyzed in mine site streams occur at concentrations that are below the strictest potentially applicable surface water quality standard. We focus our review on four contaminants of concern: copper, antimony, arsenic, and mercury. These four contaminants were selected because existing and potential future concentrations may reduce individual fitness and the ability of the habitat to support successful spawning, rearing, and migration.

Naturally occurring mineralization and historical mining activity have resulted in surface water quality impairments for arsenic, antimony, and mercury (Baldwin and Etheridge 2019). As such, recent surface water baseline studies conducted by both Perpetua and USGS have attempted to characterize antimony, arsenic, and mercury concentrations in the Headwaters EFSFSR and Sugar Creek subwatersheds.

Table 34 provides the baseline conditions for the four contaminants of concern compared to the applicable screening levels. NMFS uses the phrase "screening level" to reduce potential confusion since not all levels are equivalent to existing aquatic life criteria. At the time of this opinion, only copper had aquatic life criteria that had undergone ESA consultation and were effective for CWA purposes. Antimony does not have aquatic life criteria. The screening levels for mercury and arsenic are equivalent to the reasonable and prudent alternatives identified by NMFS (2014). Additionally, the mercury screening levels are very similar to the recently proposed aquatic life water column criterion for mercury in Idaho (EPA 2024)

Water quality in the mine site area, particularly in the EFSFSR and in Sugar Creek have chemical constituents, particularly arsenic, antimony, copper, and mercury, that exceed screening levels (USFS 2023a). Copper exceeds the screening level in Sugar Creek but is well below the screening level in the other mine site area streams. Antimony and arsenic exceed screening levels in all streams that can support ESA-listed fish in the project area. Mercury exceeds the screening level in the EFSFSR and in Sugar Creek. Table 34 provides a summary of the measured concentrations of these chemical constituents. Figure 4.1-3 in the BA (Stantec 2024) provides the monitoring locations corresponding with the nodes shown in Table 34. These chemical constituents have the potential to affect the growth and survival of Chinook salmon and steelhead at their current concentrations.

	<b>Contaminant of Concern</b>		Copper <sup>1</sup>		Antimony <sup>2</sup>		Arsenic <sup>3</sup>			Mercurv <sup>4</sup>			
	<b>Screening Levels</b>	$2.4 \mu$ g/L		5.6 $\mu$ g/L		$10 \mu g/L$		$2 \text{ ng/L}$					
<b>Node</b>	<b>Stream</b>	Avgl	Min   Max		Avg	Min	Max	Avg	Min	Max	Avg	Min	Max
$YP-T-27$	Meadow Cr.	0.3	0.1	0.7	6.1	2.04	16.9	35	11.8	60.7	2.5	$<$ 1	11.8
$YP-T-22$	Meadow Cr.	0.3	0.1	1	8.1	2.4	35.8	34	13.6	56.8	15.6	1.3	404
$YP-SR-10$	<b>EFSFSR</b>	0.2	< 0.1	0.5	12	3.9	47.1	25	8.6	41.4	6.1	2.0	31.5
$YP-SR-8$	<b>EFSFSR</b>	0.3	0.1	2.6	17	5.7	61.8	28	12.3	48.7	6	1.6	20.1
YP-SR-6	<b>EFSFSR</b>	0.2	0.1	0.5	19	6.4	46.9	31	12.6	41.4	5.6	1.9	24.7
YP-SR-4	<b>EFSFSR</b>	0.3	0.1	0.6	31	10.4	62	63	20.8	105	5.9	< 0.5	32.7
YP-SR-2	<b>EFSFSR</b>	0.2	0.1	0.6	22	6.8	38.2	45	14.7	71.1	41.3	3.1	395
$YP-T-1$	Sugar Cr.	$8.5^{5}$	0.1	3425	34	3.4	8.6	13	6.5	22.4	159	9.6	2,380

**Table 34. Average Measured Constituent Concentrations at Monitoring Locations in the Mine Area.**

Source: Midas Gold 2019b; SRK 2021a; and USFS 2023a (Table 6-9a);

Key: Cr. = Creek; EFSFSR = East Fork South Fork Salmon River; µg/L = micrograms per liter; ng/L = nanograms per liter,  $\Delta vg = Average$ ; Min = Minimum; Max = Maximum.

Notes: Screening levels pertain to protection of fish species. Arsenic and mercury criteria are based on total concentrations while copper and antimony are based on dissolved concentrations.

Copper screening level is based on the chronic criterion, which was derived using the Biotic Ligand Model per guidance contained in IDEQ (2017b). A conservative chronic copper standard was estimated by applying the lowest of the 10th percentile chronic criteria based on regional classifications for the Salmon River Basin, Idaho Batholith, and third order streams. Per the SGP Water Quality Management Plan (Brown and Caldwell 2020a), preliminary calculations using the Biotic Ligand Model and site-specific data have produced similar values (i.e., 2.6 ug/L chronic and 4.2 ug/L acute criteria for the mine WTP outfall and 1.5 ug/L chronic and 2.5 ug/L acute criteria for the working housing treatment plant outfall) to the standard derived using these regional classifications (i.e., conservative guidance of 2.5 ug/L chronic and 4.0 ug/L acute criteria for third order streams and 2.4 ug/L chronic and 3.9 ug/L acute criteria for the Salmon River Basin).

Antimony does not have a specified NMFS or USFWS criteria; the screening level is based on EPA's human health chronic criterion for consumption of water and organisms is 5.6 µg/L.

Arsenic: NMFS (2014) directed EPA to promulgate or approve new aquatic life criterion. In the interim, NMFS directed EPA to ensure the 10 µg/L human health criterion applied in all National Pollutant Discharge Elimination System permits. NMFS utilized this interim threshold as the screening level for our analysis.

Mercury: NMFS (2014) directed EPA to promulgate or approve a new criterion. In the interim, implement the fish tissue criterion that IDEQ adopted in 2005 (IDEQ 2005). Where fish tissue is not readily available, then NMFS specified application of a 2 ng/L threshold (as total mercury) in the interim. NMFS utilized this interim water column threshold as the screening level for our analysis.

Of the 38 dissolved copper values reported for YP-T-1, only one value was higher than 2.6 µg/L; therefore, it is likely that this single anomalous value of 342 µg/L was the result of a sampling, analytical, or data management error as all other observations were less than  $2.6 \mu g/L$ .

Ambient water quality criteria for the protection of aquatic life have not been established for antimony, so a screening threshold of 5.6 µg/L was selected. This screening threshold is the water quality criterion for protection of human health for the consumption of water and organisms. Average antimony concentrations currently exceed the screening threshold at every assessment node (Table 34). Monitoring by Baldwin and Etheridge (2019) found that antimony in mine site streams primarily occurs in the dissolved phase (primarily as Sb(V)) (Dovick et al. 2015). Lower antimony concentrations were recorded during high flow periods, suggesting a groundwater source.

Dissolved copper concentrations were almost always below the Biotic Ligand Model-based copper criteria, which was selected as the screening threshold, at all sites. The BA (Stantec 2024) reported that of the 38 samples for Sugar Creek at YP-T-1, only one value was higher than 2.6 micrograms per liter ( $\mu$ g/L). This value was reported as 342  $\mu$ g/L, and is considered to be an anomaly resulting from a sampling, analytical, or data management error.

The screening threshold for arsenic is  $10 \mu g/L$ , which was selected based upon the analysis in the Idaho Toxics opinion (NMFS 2014). Arsenic concentrations currently exceed the screening threshold at all assessment nodes (Table 34). Monitoring by Baldwin and Etheridge (2019) found that arsenic in mine site streams primarily occurs in the dissolved phase and was negatively correlated with streamflow, suggesting groundwater is the primary source of instream arsenic concentrations.

The screening threshold for mercury is 2 ng/L, which was selected based upon the analysis in the Idaho Toxics opinion (NMFS 2014). Mercury in the environment originates from both natural and anthropogenic (human-caused) sources. The most significant source of mercury in Idaho is air deposition; however, in the action area, mercury concentrations in Sugar Creek and the EFSFSR downstream of its confluence with Sugar Creek are significantly influenced by the Cinnabar Mine Site (Baldwin and Etheridge 2019; Eckley et al. 2021; Holloway et al. 2017). Mercury concentrations currently exceed the 2 ng/L screening threshold in the EFSFSR below Meadow Creek and Sugar Creek (Table 34). Mercury concentrations are also elevated in the upper reaches of West End Creek.

*Sediment and Tissue Concentrations.* Baseline sediment, macroinvertebrate, and fish tissue metals concentrations are described in the Aquatic Resources 2016 Baseline Study report MWH (2017) and summarized in tables 3.5-23 through 26 in the BA (Stantec 2024). Collectively, these data indicate the amount of metal contaminants entering the aquatic food web in the project area under baseline conditions. Results specific to mercury and arsenic concentrations in sediments and macroinvertebrate tissues are summarized in Table 35). Sediment concentrations, under current conditions, are elevated above selected screening levels. Arsenic concentrations in macroinvertebrate tissues are elevated in the EFSFSR downstream of Meadow Creek, and in Meadow Creek downstream of the SODA. Concentrations of arsenic in macroinvertebrates collected at sites upstream of historic mining influence in Sugar Creek, Meadow Creek, and the EFSFSR are below screening levels (Table 35). MWH (2017) provided a comparison table of macroinvertebrate tissue concentrations documented in 1995 through 1997 and 2016 (Table 36). Macroinvertebrate tissue concentrations for the lower Meadow Creek sites are substantially improved following restoration activities that occurred in the mid 1990s and early 2000s.

				<b>Sediment</b> (mg/kg dry weight)			<b>Macroinvertebrate Tissue</b>	(mg/kg dry weight)	
<b>Stream</b>	<b>MWH</b> <b>Site ID</b>	Analyte	Antimony	Arsenic	Mercury	Antimony	Arsenic	Mercury	Mercury <sup>1</sup>
		<b>Screening</b> levels (mg/kg)	$\overline{2}$	9.8	0.18		$>14$		>0.2
<b>EFSFSR</b> (below Rabbit Cr)	013		6.9	44.6	4.45	0.58	9.39	0.25	0.04
<b>EFSFSR</b> (below Meadow Cr.	012		743	200	11.6	13.3	38.2	0.23	0.04
<b>EFSFSR</b> (below Garnet Cr.)	011		222	246	0.305	6.33	57.8	0.19	0.03
<b>EFSFSR</b> (YPP alluvial fan)	074		138	184	0.962	6.67	29.1	0.23	0.02
<b>EFSFSR</b> (YPP outlet)	073		164	611	0.46				
<b>EFSFSR</b> (downstream of YPP)	030		241	1640	0.64	21.3	572	0.19	0.13
<b>EFSFSR</b> (above Sugar Cr.)	007		152	1090	1.4	27.2	331	0.18	0.03
<b>EFSFSR</b> (below Sugar Cr.)	009		48	219	62.6	8.06	100	0.43	0.03
Meadow (background)	016		2.9	11.1	0.037	0.24	6.49	0.212	0.02
Meadow (above EFMC)	075		186	307	0.447	$\mathbf{1}$	55.5	0.18	0.03
Meadow (below EFMC)	014		19.8	52.8	0.153	1.84	71.9	0.20	0.03
Sugar (near West End Cr.)	069		19.3	81.2	70.2	1.11	24	2.56	0.46
Sugar (below West End Cr.)	029		19	80	61.8	0.39	9.35	2.06	0.34

**Table 35. Metals Concentrations in Sediments and Macroinvertebrate Tissues Compared to Screening Levels (data collected in 2016).**

**Key:** EFSFSR = East Fork South Fork Salmon River; YPP = Yellow Pine Pit; mg/kg = milligrams per kilogram; PQL = Practical Quantitation Limit; YPP = Yellow Pine pit

**Note:** Shaded cells are higher than the applicable screening level.

<sup>1</sup>Mercury data in mg/kg wet weight

<sup>2</sup>Screening levels for sediment were obtained from EPA website on November 15, 2017, Freshwater Sediment Screening

Benchmarks developed for Region 3, at [https://www.epa.gov/risk/freshwater-sediment-screening-benchmarks.](https://www.epa.gov/risk/freshwater-sediment-screening-benchmarks)\*Several samples had higher mercury PQL due to required dilutions.

	$\cdots$	<b>Antimony</b>			<b>Arsenic</b>				<b>Mercury</b>			
<b>Stream</b>	<b>Site</b>	1995	1996	1997	2016	1995	1996	1997	2016	1995	1997	2016
	009	4.58	11.2	2.48	8.06	43.8	1117	17.9	100	0.8	0.24	0.43
<b>EFSFSR</b>	007	21	43.7	16.9	27.2	320	333	79.4	331	۰	< 0.25	0.18
	011	14.4	27	12	6.33	49.7	74.9	23	57.8	0.4	< 0.25	0.19
	012	57.4	14.9	12.8	13.3	130	75.9	44.5	38.2	$\overline{\phantom{a}}$	< 0.25	0.23
Meadow	014	36.8	76.8	13.9	1.84	209	147	63.5	71.9	0.59	< 0.25	0.20
Creek	075	11	23.3	12	0.99	105	291	45	55.5	0.44	< 0.33	0.18
Sugar Creek	029	1.59	$\overline{\phantom{a}}$	< 0.5	0.39	47.1	٠	10.9	9.35	1.09	0.6	2.06
	069	0.7	9.24	< 0.5	1.11	5.97	1.54	5.52	24	1.05	1.64	2.56

**Table 36. Comparison of Macroinvertebrate Tissue Metal Concentrations (mg/kg dry weight).**

**Source:** URS Corporation 2000 for 1995-1997 data

**Note:** Concentrations are milligrams per kilogram dry weight.

**Key:** '- '= Not sampled or not analyzed

Results specific to fish tissue concentrations are summarized in Table 37. Arsenic and antimony concentrations were highest in fish collected from the EFSFSR. Mercury was highest in fish collected from Sugar Creek.

Available information indicates that methylation is occurring within the vicinity of the SGP. Current fish tissue concentrations in resident fish in Meadow Creek and the EFSFSR are below levels that are thought to be associated with adverse effects (Table 37) (MWH 2017). Total mercury concentrations in fish in Sugar Creek are at levels that could elicit adverse effects (Kraus et al. 2022; Eckley et al. 2021; MHW 2017). Bull trout whole body tissue concentrations were collected by the USGS (Rutherford et al. 2020) from Sugar, Cinnabar, and Cane Creeks. Whole body total mercury concentrations in bull trout collected from habitats upstream of mining influence were generally below 0.1 mg/kg wet weight (ww); whereas whole body samples from habitats influenced by mercury contamination from the Cinnabar Mine ranged from 0.1 to 0.314 mg/kg ww (Figure 27). The data did not show any apparent relationship between fish size and mercury tissue burdens.

				Analyte (mg/kg wet $weight)^1$	<b>Mercury</b>	<b>Antimony</b>	<b>Arsenic</b>
<b>Stream</b>	<b>Site ID</b>	<b>Sample ID</b>	<b>Species</b>	<b>EPA 2015</b>	0.5	9.0	2.0
				<b>NMFS 2014</b>	$0.2 - 0.3$		$2.0 - 5.0$
	MWH-027	1 <sub>G</sub>	<b>WCT</b>		0.04	0.05	0.09
<b>EFMC</b>	<b>MWH-027</b>	2G	<b>WCT</b>		0.04	0.03	0.14
	MWH-027	3G	<b>WCT</b>		0.02	0.04	0.21
	EFSFSR01	$5G$	<b>WCT</b>		0.02	0.06	0.27
	EFSFSR02	6G	<b>WCT</b>		0.03	0.06	0.49
	EFSFSR03	1H	<b>WCT</b>		0.02	0.06	0.36
	EFSFSR04	2H	<b>WCT</b>		0.02	0.04	0.14
	EFSFSR05	4G	<b>WCT</b>		0.02	0.05	0.09
	Glory Hole	1A	<b>WCT</b>		0.02	0.12	0.07
	Glory Hole	2A	<b>WCT</b>		0.06	0.24	0.31
<b>EFSFSR</b>	Glory Hole	3A	<b>WCT</b>		0.03	0.15	0.15
	<b>MWH-011</b>	$1\mathrm{F}$	<b>WCT</b>		0.03	0.16	0.11
	<b>MWH-011</b>	2F	<b>WCT</b>		0.03	0.07	0.10
	<b>MWH-011</b>	3F	<b>WCT</b>		0.03	0.11	$0.18\,$
	MWH-026	4A	<b>WCT</b>		0.02	0.34	0.09
	MWH-026	5A	<b>WCT</b>		$0.05\,$	0.09	$0.04\,$
	MWH-026	4B	<b>WCT</b>		0.07	0.06	0.13
	MWH-026	$5\mathrm{B}$	<b>WCT</b>		0.05	0.06	0.05
	<b>MWH-014</b>	3E	<b>WCT</b>		0.05	0.05	0.19
	<b>MWH-014</b>	4E	<b>WCT</b>		0.03	0.05	0.23
Meadow	<b>MWH-014</b>	5E	<b>WCT</b>		0.04	0.04	0.12
Creek	<b>MWH-016</b>	1 <sup>C</sup>	<b>WCT</b>		0.02	0.07	0.01
	MWH-016	2C	<b>WCT</b>		0.02	0.06	0.03
	<b>MWH-016</b>	3C	<b>WCT</b>		0.02	0.05	0.02
	<b>MWH-018</b>	1D	Sculpin		0.20	0.06	0.05
Sugar Creek	<b>MWH-018</b>	1E	<b>WCT</b>		0.07	0.05	0.06
	<b>MWH-018</b>	$2E$	Sculpin		0.09	0.04	0.05

**Table 37. Fish Tissue Metals Concentrations (samples collected in 2015).**

**Note:** Effects thresholds obtained from literature derived values in EPA 2015 and NMFS 2014. Effects thresholds are mg/kg wet weight.

Shaded cells are higher than the applicable threshold. Bold text cells are the maximum concentrations for each metal.

<sup>1</sup>Laboratory values were in mg/kg dry weight but were converted to mg/kg wet weight for comparison with the effect's thresholds, which are typically in mg/kg wet weight. Values were converted using the equation: wet weigh  $=$ dry weight x [1-(percent moisture/100)]. Percent moisture used in the equation was 77.54, which is the value reported for muscle in EPA 2016. Concentrations below the MDL were not converted to wet weight and the MDLs are as noted below:

a Below MDL of 0.7 b Below MDL of 0.046 c Below MDL of 0.11 d Below MDL of 0.039 e Below MDL of 0.15 f Below MDL of 0.035 g Below MDL of 0.055

Key: EFMC = East Fork Meadow Creek EFSFSR = East Fork of the South Fork of the Salmon River; mg/kg = milligrams per kilogram;  $WCT = Westslope$  cutthroat trout



### **represent "background" concentrations, and the remaining sites are influenced by mercury contamination at the Cinnabar Mine site.**

# *2.9.2.4.5. Change in Peak/Base Flows*

The SGP is located in mountainous terrain with typically narrow valleys and steep slopes. Elevations range from 6,000 to 6,600 ft. amsl along valley floors, rising to elevations exceeding 8,500 ft. amsl in the surrounding mountains (HydroGeo 2012a). Climate in the analysis area is characterized by wide annual, seasonal, and diurnal variations in temperature and humidity. The area typically has cold wet winters and hot dry summers, with most precipitation falling in October through May, and long periods of little or no precipitation common from mid-June through mid-September. During winter, storms typically move through the region resulting in seasonal snowfall accumulations of six feet or more. Cloudy and unsettled weather is common during the winter with measurable precipitation occurring on about one-third of the days (Brown and Caldwell 2017; Stantec and Trinity Consultants 2017).

The main EFSFSR valley floor is around 6,400 ft. in elevation and the tributary valleys—which are at higher elevations like Meadow, Fiddle, Hennessy, and Sugar Creeks—all show a strong and pronounced asymmetry with steeper south-facing slopes (Midas Gold 2017). South-facing slopes are more open to sunlight and warm winds and are thus generally warmer and dryer because of the higher levels of evapotranspiration compared to steep north-facing slopes.

A long-term climatological record is not available for the SGP. Therefore, Parameter-elevation Regressions on Independent Slope Model (PRISM) data compared with the National Weather Service and SNOTEL (snow telemetry) Secesh Summit site is used to develop average precipitation and temperature estimates (Table 38). The Secesh Summit site is located 35 miles northwest of the SGP, at a comparable elevation (Brown and Caldwell 2017).

<b>Month</b>	Average Precipitation (inches)	Average Temperature (°F)	<b>Minimum</b> Temperature (°F)	<b>Maximum</b> Temperature (°F)
January	4.11	20.10	10.67	29.52
February	3.32	21.75	9.84	33.66
March	3.53	27.68	15.33	40.03
April	2.98	32.89	20.50	45.27
May	2.58	40.69	27.73	53.65
June	2.14	48.73	33.85	63.61
July	0.95	58.05	41.31	74.79
August	0.91	56.47	39.18	73.76
September	1.81	48.70	32.76	64.63
October	2.10	39.18	25.97	52.39
November	3.71	26.34	17.02	35.63
December	3.99	18.82	9.28	28.36
Annual	32.19	36.61	23.61	49.60

**Table 38. Estimated Average Monthly Precipitation and Temperature for the Analysis.**

**Source:** 800-meter PRISM data, Brown and Caldwell 2017.

Because the action area includes the SFSR downstream from the EFSFSR, and flows in that reach are influenced by water use throughout the SFSR drainage, the analysis in this opinion considers water usage throughout the SFSR drainage. There is no large-scale agriculture in the SFSR drainage. The largest irrigated acreages include pastures and/or hay production for pack and saddle stock, and turf maintenance at the Johnson Creek Airport. Other water uses include domestic use at summer homes and support of residences and businesses in the Yellow Pine community; support of summer homes and businesses in the Warm Lake area; and operation of small (off grid) hydropower operations. Small amounts of water are also periodically diverted to facilitate USFS dust abatement and for fire-fighting operations, and are occasionally diverted to facilitate minerals exploration. Overall, water use in the SFSR drainage appears to be very light, based on the number and size (i.e., allowable diversion rates) of water rights and the lack of water rights on the vast majority of tributaries, and flow in most stream reaches is essentially unimpaired by water diversion/use. Based on water use, the action area is likely FA for peak/baseflow changes. However, base flows are relatively low, compared to the rest of the hydrograph, suggesting that base flows may be naturally more limiting of salmonid production than in other portions of the Salmon River drainage (e.g., the Upper Salmon).

There are three instream flow water rights with quantification reaches in the action area. These are 77-14196 on Sugar Creek, 77-14190 on the EFSFSR, and 77-14174 on the SFSR. The

purposes of these instream flows are to preserve fish and wildlife, scenic values, and recreational values. The volume of these instream flows is equivalent to the unimpaired monthly 40 percent exceedance flows. As such, the establishment these instream flow water rights recognize the importance of greater than median flows for protection of aquatic resources. However, these water rights are be subordinate to all domestic, commercial, municipal, and industrial water rights, and therefore will not protect instream flows from the proposed action.

*Streams.* The SGP is in the Headwaters EFSFSR and Sugar Creek subwatersheds. The primary surface water features at the SGP include the EFSFSR and its tributaries, as well as intermittent drainages, ephemeral drainages, seeps, springs, wetlands, and ponds. These features include 10 named surface water channels: the EFSFSR, EFMC, Rabbit, Meadow, Garnet, Fiddle, Midnight, Hennessy, West End, and Sugar Creeks. Most of these streams occur in the Headwaters of the EFSFSR subwatershed except for Sugar and West End Creeks, which are in the Sugar Creek subwatershed. Brief descriptions of each stream are provided below, and specific drainage and channel characteristics are summarized in Table 39.

The EFSFSR is a perennial stream that flows from southeast to northwest through the SGP and has a drainage basin of 25 square miles upstream of its confluence with Sugar Creek. It is the principal stream draining the SGP and receives flow either directly or indirectly from all other drainages listed in Table 40. At ordinary high water, the EFSFSR is approximately 2 to 3 ft. deep and 25 to 30 ft. wide (Brown and Caldwell 2017).

<b>Drainage</b>	Approximate <b>Drainage Area</b> (square miles)	<b>Channel</b> Length (miles)	<b>Elevation</b> <b>Change</b> (feet)	Average Gradient (%)
<b>EFSFSR</b> (upstream of Sugar Creek)	25.0	7.04	2,129	5.7
Meadow Creek	7.7	4.78	1,570	6.2
<b>EFMC</b>	2.4	2.66	1,491	10.6
Rabbit Creek	0.6	1.19	1,506	24.0
Garnet Creek	0.5	1.24	1,558	23.8
Fiddle Creek	2.0	2.47	1,444	11.1
Midnight Creek	0.9	1.83	2,205	22.8
<b>Hennessy Creek</b>	0.7	1.16	1,499	24.5
West End Creek	0.6	1.55	2,234	27.3
Sugar Creek	17.4	7.14	2,356	6.2

**Table 39. Summary of Stream Characteristics in the SGP Area.**

**Source:** Brown and Caldwell 2017; HydroGeo 2012b

Historical mining activities have affected the course of the EFSFSR in the central portion of the SGP where it flows through the YPP Lake. The river enters the pit on the south side and exits from the north. The flow velocity of the EFSFSR slows as it passes through the YYP Lake, causing the river to drop much of its sediment load. The original YPP was excavated to a depth of 125 ft. below the current pit lake level, but sediment deposited through time has reduced the lake depth to only 35 ft. The lake has a surface area of approximately 4.75 acres and is estimated to contain approximately 92 acre-feet of water (Brown and Caldwell 2017). An artificial drop into the pit creates a steep whitewater cascade on the EFSFSR where it enters the pit and blocks upstream fish passage above the pit lake.

Meadow Creek originates southwest of the SGP, flows east into the EFSFSR, and drains an area of approximately 7.7 square miles. The Meadow Creek headwaters occur in an alpine lake, and the drainage contains multiple wetland complexes. At ordinary high water, Meadow Creek is approximately 2 to 4 ft. deep and 20 to 25 ft. wide at the bottom of the drainage (Brown and Caldwell 2017).

EFMC is a tributary to Meadow Creek that drains an area of 2.4 square miles in the southern end of the SGP. The creek previously supplied water to a reservoir that provided hydroelectric power and process water to the historical mill and smelter. EFMC is locally referred to as Blowout Creek because the dam forming the reservoir breached in 1965, causing large-scale scouring of the steep channel downstream, and deposition of an alluvial fan. From its headwaters, EFMC meanders through a former wetland area that dried up due to stream incision and declining groundwater levels related to the dam failure.

Rabbit and Garnet Creeks are small tributaries of the EFSFSR that drain 0.6 and 0.5 square miles, respectively. Rabbit Creek is in a steep drainage that has steep side slopes, with numerous seeps and springs occurring throughout its headwaters. Garnet Creek is formed from seeps and springs located in the eastern portion of the SGP. The current shop, camp facilities, and the historical Garnet Pit are in the Garnet Creek drainage. Historical waterworks from the 1940s and 1950s, as well as a 1990s diversion, are present below the former open pit.

Fiddle Creek occurs in a well-defined glacial cirque, drains an area of two square miles, and flows into the EFSFSR from the west. The drainage area for Fiddle Creek includes forested and open scree slopes. The middle reach of Fiddle Creek also contains a former reservoir and dam, and a former townsite occurs in the lower reach above and below the County Road. In addition, the creek itself was diverted from its natural outfall site to the north under the County Road through a culvert in the 1980s.

Midnight Creek is a small tributary that drains an area of 0.9 square miles and flows into the EFSFSR from the east, just above the YPP Lake. Several miles of current and historical exploration and haul roads exist in the Midnight Creek drainage.

Hennessy Creek is a small tributary that drains an area of 0.7 square miles and flows into the EFSFSR from the west. The upper end of the drainage is heavily forested, and the lower portion of the drainage has been modified by current access roads and historical mine workings. Hennessy Creek also has a historical water diversion just above the county road that included a large pipe system. The creek flows in the direction of, and then adjacent to, Stibnite Road (CR 50-412) in a channel around the Bradley Northwest mine dump complex, through a diversion installed in 2022 under the Stibnite ASAOC to avoid historical mine development rock piles, and through two culverts before entering the EFSFSR.

West End Creek flows into Sugar Creek from the south and has a drainage area of 0.6 square miles. The drainage basin of West End Creek was modified extensively and diverted into a now failed French drain system during construction of the large waste rock dump in the middle reach. The current creek flow disappears and reemerges among historical waste rock piles. Several miles of current and historical exploration roads are present in the West End Creek drainage.

Sugar Creek drains an area of approximately 17.4 square miles and flows into the EFSFSR downstream of the YPP. A portion of the upper Sugar Creek valley has been impacted by past mercury mining activities at the former Cinnabar Mine, located in the upper Cinnabar Creek drainage which is a tributary to Sugar Creek. These activities included underground mine development and operations, development rock disposal, ore processing, deposition of tailings in the valley, construction and use of buildings and housing (several of which still exist), and road construction.

Nine USGS streamflow gages (Figure 27) in and near the analysis area provide data to characterize the existing environment. Table 40 provides streamflow statistics for these gaging stations, and Figure 27 presents average monthly discharge hydrographs for the six gaging stations in the project area. The hydrographs illustrate the snowmelt-dominated streamflow pattern observed in the area with flows beginning to rise in March and April and peaking in May or June, before receding to base flow conditions in late summer/fall and remaining low through the winter.

Baseflow and groundwater recharge estimates were derived using data from two of the USGS gaging stations in the analysis area (Brown and Caldwell 2017). Groundwater recharge over the Sugar Creek and EFSFSR drainage areas was calculated to 8.1 inches per year over the alluvial valley bottom areas and 6.2 inches per year in the bedrock dominated mountainous areas. These values represent about 20 percent of the estimated annual precipitation for the SGP analysis area, which is equal to 32.19 inches (Perpetua 2021d).

USGS data also were used to derive peak flow statistics for the seven major drainages in the analysis area. Results from the peak flow analysis were summarized in the baseline study (HydroGeo 2012b) and Table 41. Peak flows were calculated for the bottom of each drainage using the USGS StreamStats program [\(https://streamstats.usgs.gov/ss/\)](https://streamstats.usgs.gov/ss/).

In addition to the USGS data, streamflow data were collected in conjunction with surface water quality baseline sampling on a monthly or quarterly basis at 32 non-USGS monitoring stations (Figure 27). The monitoring points were selected at upstream and downstream locations to bracket historical and potential future mining activities in the analysis area (Brown and Caldwell 2017). Table 42 provides streamflow statistics derived from baseline measurements collected between 2012 and early 2016. The mean flows calculated from this dataset for the EFSFSR ranged from 4.47 cfs at the farthest upstream monitoring location YP-SR-14, to 31.31 cfs at the most downstream location YP-SR-2. Note that the baseline monitoring sites are at different locations than the USGS gaging stations, thus providing additional site-specific data proximal to historical and proposed facilities.



**Figure 28. USGS Gaging Stations, Stibnite Gold Project.**

Gage <b>Number</b>	<b>Gage Name</b>	Drainage Area $(mi^2)$	Min (cfs)	Max (cfs)	Mean (cfs)	<b>Period of Record</b> (# years monitored)
13310850	Meadow Creek near Stibnite, Idaho	5.6	1.37	129	11.0	09/2011-02/2022 $(10 \text{ years})$
13310800	EFSFSR above Meadow Creek near Stibnite, Idaho	9.0	2.20	159	11.8	09/2011-02/2022 $(10 \text{ years})$
13311000	EFSFSR at Stibnite, Idaho	193	3.50	413	31.5	1928-1943 1982-1997 2010-2022 $(41 \text{ years})$
13311450	Sugar Creek near Stibnite, Idaho	18.0	4.00	252	22.9	09/2011-02/2022 $(10 \text{ years})$
13311250	EFSFSR above Sugar Creek near Stibnite, Idaho	25.0	4.39	366	36.9	09/2011-02/2022 $(10 \text{ years})$
13311500	EFSFSR near Stibnite. Idaho <sup>1</sup>	43.0	10	783	50.4	06/1928-09/1941 $(13 \text{ years})$
13312000	EFSFSR near Yellow Pine. Idaho <sup>1</sup>	107.0	28	1.660	142.4	08/1928-07/1943 $(13 \text{ years})$
13313000	Johnson Creek at Yellow Pine, Idaho	2180	28	5,440	342.5	09/1928-02/2022 $(93 \text{ years})$
13310700	SFSR near Krassel Ranger Station, Idaho	330.0	35	6,200	5366	10/1966-02/2022 $(55 \text{ years})$

**Table 40. USGS Gaging Station Drainage Area and Flow Statistics.**

Source: Brown and Caldwell 2017 – Table 7-9; Flow data from 2017-2022 updated from waterdata.usgs.gov.<br><sup>1</sup> Inactive

<b>Drainage</b>	$1.5$ -year event	2-year event	$2.33$ -year event	5-year event	10-year event	25-year event	50-year event	100-year event
	<b>PK1.5</b> (cfs)	PK2 (cfs)	<b>PK2.33</b> (cfs)	<b>PK5</b> (cfs)	<b>PK10</b> (cfs)	<b>PK25</b> (cfs)	<b>PK50</b> (cfs)	<b>PK100</b> (cfs)
Meadow Creek (13310850)	83 $(76-91)$	98 $(90-107)$	105 $(97-114)$	132 $(122-144)$	152 $(140-168)$	175 $(159-200)$	191 $(170-223)$	205 $(179-247)$
<b>EFSFSR</b> above Meadow Creek (13310800)	83 $(75-90)$	97 $(89-105)$	104 $(96-112)$	130 $(120-141)$	149 $(138-165)$	171 $(156-195)$	186 $(167-218)$	200 $(176-241)$
<b>EFSFSR</b> below Meadow Creek (13311000)	174 $(154-195)$	215 $(193-240)$	235 $(211-261)$	316 $(285 - 353)$	379 $(341-432)$	454 $(401-539)$	507 $(438-623)$	557 $(469-710)$
<b>EFSFSR</b> above Sugar Creek (13311250)	229 $(205 - 254)$	279 $(252-307)$	301 $(273-331)$	395 $(359-437)$	466 $(423-525)$	550 $(491-643)$	608 $(532 - 733)$	662 $(566-826)$
<b>EFSFSR</b> below Sugar Creek (13311500)	372 $(327-418)$	465 $(415-520)$	508 $(454 - 567)$	693 $(622 - 777)$	837 $(749-959)$	1010 $(888-1207)$	1133 $(973-1403)$	1249 $(1044 - 1606)$
Sugar Creek (13311450)	143 $(124-162)$	181 $(160-204)$	199 $(177-224)$	278 $(247-314)$	340 $(301-393)$	415 $(361-502)$	469 $(398-589)$	520 $(429 - 680)$
Johnson Creek (13313000)	2,497	2,962 $(2,268-2,727)(2,713-3,230)(2,911-3,563)(3,737-4,491)(4,356-5,375)(5,058-6,592)(5,521-7,574)(5,936-8,617)$	3,175	4.079	4.789	5,652	6,273	6,877

**Table 41. Peak Stream Flow Statistics for Drainages in the Analysis Area.**

**Source:** Rio ASE 2021, Appendix C.

cfs = cubic feet per second; peak flow volume statistic reported followed by its 95 percent confidence interval in parentheses.

<b>Monitoring Site</b>	Stream	Min (cfs)	Max (cfs)	Mean (cfs)
$YP-SR-2$		8.97	74.56	31.31
YP-SR-4		7.67	37.84	16.92
$YP-SR-6$		8.00	50.76	20.38
$YP-SR-8$	<b>EFSFSR</b>	5.88	61.08	19.33
$YP-SR-10$		6.23	106.21	23.97
$YP-SR-11$		3.32	40.67	10.41
YP-SR-13		2.05	54.92	11.56
$YP-SR-14$		0.48	22.25	4.47
$YP-T-1$	Sugar Creek	5.71	78.06	21.24
$YP-T-6$		0.16	1.68	0.51
$YP-T-37$	West End Creek	0.003	0.12	0.03
YP-T-49		0.37	1.37	0.71
YP-T-7		5.25	34.12	12.51
$YP-T-SA$	Sugar Creek	4.61	77.36	19.27
$YP-T-10$		0.15	2.62	0.67
$YP-T-42$	Midnight Creek	0.12	3.59	0.99
$YP-T-11$	Fiddle Creek	0.22	20.57	3.30
$YP-T-12$		0.15	17.87	3.59
$YP-T-15$	Scout Creek	0.04	0.62	0.15
$YP-T-21$	Rabbit Creek	0.22	3.47	0.95
$YP-T-22$		3.91	86.61	17.94
$YP-T-27$	Meadow Creek	2.78	76.45	14.86
$YP-T-33$		1.96	41.13	9.22
$YP-T-43$		1.97	49.00	13.48
YP-T-29	<b>EFMC</b>	0.78	24.45	4.69
$YP-T-35$	Garnet Creek	0.01	1.16	0.19
$YP-T-40$	Salt Creek	0.80	13.38	2.80
$YP-T-41$		0.15	7.37	1.25
$YP-T-48$	Hennessy Creek	0.09	5.09	1.00
$YP-T-44$	Fern Creek	0.06	2.65	0.54
YP-T-45	North Fork Meadow Creek	0.24	19.01	3.92
$YP-T-46$	South Fork Meadow Creek	0.28	9.67	3.04

**Table 42. Baseline Monitoring Surface Water Flow Statistics.**

Source: Brown and Caldwell 2017.

USFS 2023b, Sections 6.2.4 and 6.2.5 provide a detailed description of the stream flows associated with the Stibnite portion of the action area, a summary of which is provided here.

USGS data were used to derive peak flow statistics for the ten major drainages in the analysis area (BA Figures 4.1-4 and 4.1-5) (Stantec 2024). Results from the peak flow analysis were summarized in the baseline study (HydroGeo 2012a) and are presented in the SGP Water Quantity Specialist Report (USFS 2023a). Peak flows were calculated for the bottom of each drainage using the USGS StreamStats program. Predicted peak flows for a 1.5-year event ranged from 1.84 cfs for West End Creek to 237 cfs for the upper EFSFSR, and for a 500-year event they ranged from 13.4 cfs to 931 cfs, respectively.

Base stream flow data were collected in conjunction with surface water quality sampling on a monthly or quarterly basis at 32 non-USGS monitoring stations (USFS 2023b; Brown and Caldwell 2017). The monitoring points were selected at upstream and downstream locations to bracket historical and potential future mining activities in the analysis area (Brown and Caldwell 2017). The mean flows calculated from this dataset for the EFSFSR ranged from 4.47 cfs at the farthest upstream monitoring location to 31.31 cfs at the most downstream location. Table 43 shows average monthly stream flows during the August to March low flow period at five gaging stations and location in lower Meadow Creek in the SGP mine site streams for the years 1929 to 2017.





Note: the low-flow period is August to March.

Key: cfs = cubic feet per second; Ck = Creek; EFSFSR = East Fork South Fork Salmon River; USGS = U.S. Geological Survey.

Climate change conditions resulting in increasing air temperatures will potentially transition snow to rain resulting in diminished snowpack and earlier season streamflow along with changes in groundwater recharge to aquifers that discharge to streams. Mean annual streamflow

projections suggest a slight increase, but summer low flows are expected to decline (Halofsky et al. 2018).

A review of IDWR water right records indicates that there are no downstream consumptive-use water rights on the EFSFSR until after the river merges with Johnson Creek (HDR 2017). The IWRB maintains minimum streamflow rights on various rivers and creeks in the state, including a location near the end of the EFSFSR below the confluence with Johnson Creek, which is covered under water right 77-14190. The purpose of these minimum flows is to preserve fish and wildlife, scenic, and recreational values and to protect and enhance water quality. The minimum flow protected by water right 77-14190 varies throughout the calendar year (Table 44), with a base flow minimum of 173 cfs between October 1 and October 31 as measured on the EFSFSR at the confluence of the EFSFSR with the SFSR. Water right 77-14190 is subordinate to future domestic, commercial, municipal, and industrial uses and future non-domestic, commercial, municipal, and industrial development up to 8.2 cfs.

<b>Usage Period</b>	<b>Discharge Rate (cfs)</b>
$8/1$ to $8/31$	223
$9/1$ to $9/30$	179
$10/1$ to $10/31$	173
$11/1$ to $11/30$	214
12/1 to 12/31	222
$1/1$ to $1/31$	254
$2/1$ to $2/28$	232
$3/1$ to $3/31$	291
$4/1$ to $4/30$	625
$5/1$ to $5/31$	1,829
$6/1$ to $6/30$	2,269
$7/1$ to $7/31$	590
<b>Total Diversion</b>	2,269

**Table 44. State of Idaho, IDWR Water Right No. 77-14190 Minimum Stream Flow EFSFSR at the SFSR.**

Source: HDR 2017

cfs = cubic feet per second.

The IWRB also holds a minimum streamflow water right downstream (approximately 26.4 miles from the SGP and approximately nine miles from the EFSFSR confluence) on the SFSR (77- 14174). Water right 77-14174 is also subordinate to all future domestic, commercial, municipal, and industrial uses and future non- domestic, commercial, municipal, and industrial development up to 20.6 cfs.

Next, the IWRB holds a minimum streamflow water right on Sugar Creek above its confluence with the EFSFSR (77-14193). Water right 77-14193 is subordinate to all future domestic,

commercial, municipal, and industrial uses and future non- domestic, commercial, municipal, and industrial development up to 0.3 cfs.

*Riparian Conservation Areas.* The width of the RCA is dependent on the stream type: (1) Perennial streams have an RCA width of 300 ft. on each side of the stream channel as measured from the bankfull width; (2) intermittent and ephemeral streams have an RCA width of 150 ft. on each side of the stream's bankfull edge; and (3) special aquatic features such as wetlands have RCA widths between 50 ft. to 300 ft. Under baseline conditions, RCAs were observed to be 62 percent intact for the upper EFSFSR (Kuzis 1997), while the lower EFSFSR tributaries have relatively intact drainages and are likely to have more than 80 percent intact RCAs. Riparian vegetation in the BNF is considered below the desired conditions due to severe and widespread historical fires.

Road density within the headwaters EFSFSR subwatershed is approximately 1.8 miles per square mile (mi./mi.<sup>2</sup>), with 15.2 miles of road in RCAs. In the BNF, road densities range between less than 0.7 mi./mi.<sup>2</sup> (for 47 percent of the subwatersheds) to 1.7 mi./mi.<sup>2</sup> (for 12 percent of the total subwatersheds).

The disturbance within the watershed for an acceptable condition is less than 15 percent equivalent clearcut area (ECA); however, within the upper EFSFSR, the ECA is roughly 25 percent due to extensive wildfire in 2007 (Armstrong and Nelson 2011), with disturbance concentrated in RCAs. However, according to Nelson et al. (2004), there is no discernable proof that the high ECA has any observable effect on salmonid habitat. The ECA in the BNF is greater than 15 percent for the entire watershed.

# **Lemhi River**

### *2.9.2.5.1. Watershed Condition Indicators*

*Water Temperature.* The Lemhi River water temperature is often too high to be considered suitable for Chinook salmon spawning (Idaho Transportation Department 2017). Water temperatures in the Lemhi River are affected by the loss of riparian vegetation, flows reduced by irrigation diversions, and channelization. Total Maximum Daily Loads (TMDLs) for shade have been identified for multiple segments of the Lemhi River and its tributaries. The USGS gage (13304050) in Big Creek, just upstream from the town of Leodore, started recording water temperature in June 2022. Temperatures recorded during this period dropped to near zero between early November and mid-April. However, summer temperatures rapidly increased, with daily maximum temperatures ranging between around 13°C and nearly 20°C.

*Sediment and Turbidity.* The headwater streams of the Lemhi River are considered sediment supply zones, affected by weathering and erosion of the bordering slopes. Sediment has accumulated in the alluvial fan, creating terraces along the valley margins. As a result, the Lemhi River exhibits a pronounced deposition zone. Sediment input from land use practices in the Lemhi River basin continue to affect the mainstem Lemhi River. Several segments of the Lemhi subbasin have TMDLs approved for sediment targets in the Lemhi tributaries, though the Lemhi River TMDL is for fecal coliform bacteria and not sediment (Idaho Department of Health and Welfare 1999).

*Fish Passage.* There are no structural fish passage barriers in the Lemhi portion of the action area. Upper Lemhi River rehabilitation activities have included irrigation diversion consolidation screening and improvements for fish passage barrier removals for habitat access and tributary flow reconnection. Inadequate flows have resulted in passage barriers to salmonids.

*Chemical Contaminants.* The Lemhi River is categorized as a Category 4 water, which is defined as those impaired for one or more standards for one or more beneficial uses. The Lemhi River is primarily impaired by bacteria (*Escherichia coli*) for primary recreation based on the 1999 Lemhi River Watershed TMDL (Idaho Department of Health and Welfare 1999). Pathogens are likely a result of agricultural runoff where livestock occur. Other potential contaminants include roadway runoff, agricultural runoff containing pesticides and fertilizers, and household or commercial cleaning or other waste products.

*Peak/Base Stream Flow.* The upper Lemhi River has a complex hydrology, with interactions of snowmelt surface flows, groundwater gains and losses, and an extensive network of irrigation diversions and returns. Peak flows occur from snowmelt runoff, typically in May and June (Rio ASE and Biomark 2021). Groundwater discharge and recharge is an important factor affecting the year-round water budget, caused by an extensive alluvial aquifer. Irrigation diversions combined have legal rights totaling 205 cfs, with most diversions made between late April and early October (Rio ASE and Biomark 2021).

Water management improvements have been made for several decades, including improving flow conditions in tributary streams. Despite these improvements, the upper Lemhi River still requires a more normative hydrologic regime to promote habitat formation and improve access, particularly in light of future climate change effects. Since June 2022, flows have been recorded at USGS gage stations at Big Timber Creek near Leodore (USGS gage 13304050) and in the Lemhi River near McFarland (USGS gage 1330470), just downstream from the restoration site. Flows in Big Timber Creek had short-term peak flows just over 200 cfs in June 2022, and peak flows around 130 cfs in June 2023. Flows rapidly drop from the peak flows in early July. Flows begin to increase in March.

Flows in the Lemhi River at the USGS gage at McFarland show that, between August 10, 2022, and July 24, 2023, low flows occurred in August and September (between 45 and 60 cfs). Peak flows were as high as 298 cfs in April and 249 cfs in June. The average flow for the 11 months of available data is 122 cfs.

#### **2.10. Effects of the Action**

Under the ESA, "effects of the action" are all consequences to listed species or critical habitat that are caused by the proposed action, including the consequences of other activities that are caused by the proposed action but that are not part of the action. A consequence is caused by the proposed action if it would not occur but for the proposed action and it is reasonably certain to occur. Effects of the action may occur later in time and may include consequences occurring outside the immediate area involved in the action (50 CFR 402.17).

The BA provides a detailed analysis of the effects of the proposed action on SR Basin steelhead and SR spring/summer Chinook salmon. The BA uses the Southwest Idaho Matrix WCIs to evaluate elements of the proposed action that have the potential to affect the ESA-listed fish or PBFs of their critical habitat. The potential effects of the proposed action on ESA-listed salmon and steelhead and their critical habitat can be broadly categorized into the following bullets:

- Effects to water quality, including chemical contamination and accidental spill or discharge of toxic substances, during and after mine operation;
- Changes to surface water quantity resulting from water withdrawal from action area streams;
- Effects to fish passage and habitat conditions during and after mine operation;
- Effects to riparian vegetation in the project area; and,
- Fish handling and sampling of ESA-listed fish associated with ongoing mining and restoration efforts.

The potential for mine-related adverse effects on ESA-listed salmon, steelhead, and their critical habitat, varies depending upon a variety of factors. The potential for effects varies based on sitespecific features such as: (1) topography; (2) location of the mining; (3) the mining method; (4) toxicity of chemicals and minerals used or generated during mining; (5) mine management and engineering (including water management); (6) haul road location and design; (7) control and disposal of mine wastes; (8) reclamation methods; (9) monitoring and adaptive management effectiveness; (10) proximity to critical habitat; and (11) life stages of fish that may be affected. Effects will also vary through time, and Table 45 describes the anticipated timeline for major activities affecting action area stream habitats.

Effects from the BRGI portion of the proposed action are expected to be either minor or unlikely to occur, as described in NMFS' concurrence letter on March 14, 2022 (NMFS No: WCRO-2022-00428). Because these activities will proceed as originally described by the BNF in their BA (BNF 2022), and they will occur during the pre-construction phase of the SGP, effects of the action remain unchanged from those previously analyzed in our concurrence letter. This analysis summarizes effects and conclusions as described in our concurrence letter, referring the reader to our concurrence letter for more detailed analysis. Potential effects associated with the BRGI portion of the SGP are incorporated into this analysis and considered into our overall conclusions regarding the SGP.

<b>Period and</b> <b>Mine Years</b>	<b>Activity</b>
	Pre-Production/Construction (-3 to -1)
$-3$	<b>Conduct Burntlog Geophysical Investigation</b>
	<b>Construct Burntlog Route</b>
	Transmission Line Construction.
	Existing Garnet Creek diversion extended around plant site; restored downstream from plant site (design reach GC2).
$-3$ to $-1$	Begin construction of EFSFSR fish tunnel around YPP (up to approximately 2 years to build).
	Divert Meadow Creek and tributaries around TSF and TSF buttress area including low-flow pipes to moderate temperature.
	Lemhi River Restoration
	Fiddle Creek piped beneath growth media stockpile.
	Midnight Creek diverted into EFSFSR upstream from the tunnel, and Hennessy Creek diverted into Fiddle Creek.
	EFSFSR tunnel and associated fishway completed; EFSFSR diverted into tunnel and YPP Lake dewatering begins.
$-1$	Upper Midnight Creek placed in pipe under the West End haul road.
	West End Creek diverted around West End Pit (design reach WE2).
	Enhancement in EFSFSR (excluding YPP) and the lower portion of Meadow Creek (design reaches MC6, EF2A, EF2B, and EF2C).
	Sediment control and rock drain constructed on EFMC (design reach BC2).
	Mine Operations (1 to 15)
$\mathbf{1}$	Upper EFMC meadow, groundwater table, and associated wetlands restored.
3	Divert Meadow Creek into a constructed channel around Hangar Flats Pit footprint and downstream approximately 1,000 feet (design reaches MC4B, MC5, and MC6).
	Restore the lower section of EFMC (394 ft. downstream from the rock drain) to its new confluence with Meadow Creek (design reach BC3).
5	YPP backfill begins.
$6 - 7$	Hangar Flats pit backfilled (design reach HF1).
8	Midnight pit backfilled.
	YPP backfill completed.
10	YPP backfill surface preparation for stream liner and placement of floodplain material and growth media.
	Construct West End Pit Lake overflow channel.
	YPP stream restoration including EFSFSR, Hennessy Creek, and Midnight Creek (design reaches EF3, MNC2, and HC1&2).
11	Flow restored to EFSFSR and Hennessy Creek over the YPP backfill.
	EFSFSR diversion tunnel inactive with option to divert extreme high flows through tunnel to protect riparian vegetation development.
	Stibnite lake fills and spills.
12	Pipe removed from upper Midnight Creek haul roads and stream segment restored (design reach MNC1).
13	Flow restored to lower Midnight Creek including restored stream over YPP backfill.

**Table 45. Annual Timeline of Major Changes to Physical Stream Habitats.**



**Note:** Figures 6-2 through 6-4 in Brown and Caldwell (2021b) (Stibnite Gold Water Management Plan) depict operations period water management changes.

Key:  $EFSFSR = East$  Fork South Fork Salmon River;  $TSF = Tailing$  storage facility

The BA also describes the Lemhi Restoration part of the Project as compensatory mitigation for wetland impacts of the SGP, with the intent of improving habitat conditions in the Lemhi River basin. This restoration effort is scheduled to occur in MY-2 (Table 45). The project will enhance habitat conditions across approximately 7,000 ft. of the Lemhi River channel, increasing habitat quality and complexity, reducing channel W:D, enhancing floodplain connectivity, improving instream structure and velocity, increasing pool quantity and complexity, facilitating surface/groundwater interchange for temperature moderation, providing instream cover for fish, and establishing a riparian corridor for shade, cover, and bank stability. The potential effects of this portion of the proposed action on ESA-listed salmon and steelhead and their critical habitat can be broadly categorized into effects from fish salvage, and temporary habitat-related effects to water quality (i.e., turbidity, chemical contamination) and habitat (i.e., sedimentation of spawning gravels). In the long-term, reconnecting the Lemhi River to its floodplain is expected to improve habitat complexity and will improve habitat conditions across the reach.

Effects described below for the Stibnite mine portion of the project will affect the SFSR SR Spring/summer Chinook salmon MPG, specifically the SFSR and EFSFSR populations. Although the effects of the Lemhi Restoration portion of the project will affect the same ESU, it will affect a different MPG (i.e., Upper Salmon) and different population (i.e., Lemhi River). For SR Basin steelhead, the mine will affect the Salmon River MPG, specifically the SFSR population; while effects of the Lemhi Restoration portion of the project will affect the same MPG, but the Lemhi River population.

This analysis includes the effects of the SGP and the consequences of mine activities on non-Federal lands; therefore, in addition to the activities being permitted by the USFS and the USACE, the effects of activities on salmon and steelhead occurring on private and/or patented lands were also considered.

# **2.10.1. Effects to Critical Habitat – SFSR**

The action area contains DCH for SR spring/summer Chinook salmon and SR Basin steelhead. Critical habitat within the action area has an associated combination of PBFs essential for supporting freshwater rearing, migration, and spawning for steelhead and Chinook salmon. Individual PBFs present in the action area (Table 28), and their condition were described in Section 2.9.2. Effects from the BRGI portion of the proposed action are expected to be either minor or unlikely to occur. Adverse effects to DCH are likely to occur as a result of the USFS approving and the USACE permitting the SGP. Adverse effects will be primarily related to potential changes in water quality (contaminants or temperature) and quantity, impacts to riparian vegetation, effects to natural cover/space, reduced forage (due to chemical contaminants and water quantity effects), potential increased sediment deposition, fish passage, and floodplain connectivity. Modification of these PBFs may affect freshwater spawning, rearing or migration in the action area. Proper function of these PBFs is necessary to support successful adult and juvenile migration, adult holding, spawning, incubation, rearing, and the growth and development of juvenile fish. The following sections discuss the action and any effects on critical habitat as organized by PBF and WCI effect pathway.

# **Water Quality**

The water quality PBF is essential to steelhead and Chinook salmon spawning, rearing, and migration. Potential effects to the water quality PBF were described as potential effects to the Water Quality Pathway in the Matrix, specifically for the Chemical Contamination, Temperature, and Intragravel Quality WCIs. Effects to the water quality PBF will be described in detail below for chemical contamination, water temperature, and suspended sediment.

Water quality could be affected through chemical contamination, which could occur: (1) should a spill occur onsite or in the transportation corridor; (2) when construction equipment is working within or adjacent to the stream channel; (3) as a result of weed treatment; (4) as a result of effluent discharge; or (5) through groundwater contamination.

#### *2.10.1.1.1. Accidental Spill of Toxic Substances During Transport*

Toxic spills have the potential to enter streams in the action area either through a spill occurring during transport, or through an on-site spill and subsequent discharge into nearby waterbodies. An accidental spill of fuels or toxic chemicals being transported to or on-site at the SGP could affect ESA-listed salmon and steelhead if the spill reached any of the action area streams. The potential for, magnitude of, and severity of those effects would be dependent upon a number of variables, such as: (1) proximity to streams; (2) whether the spill reached a stream; (3) accident severity; (4) amount of material spilled; (5) volume and attributes of receiving waters at the time of the spill; (6) type of chemical; (7) form of chemical (dry or liquid); (8) transportation container; (9) weather; (10) spill response time; (11) effectiveness of spill containment; and (12) salmonid life stage(s) present and exposed.

Annually, the SGP will transport substantial quantities of fuels and other materials to the mine site (Table 19). The SGP will also ship large quantities of antimony concentrate. Many of these substances are toxic to salmonids or their food base if released to surface waters. Table 19 describes the quantity and types of chemicals that will be annually transported to/from the mine. Spills of toxic materials could occur along access roads as materials are trucked to and from the SGP during construction of access roads and mine facilities, as well as during operations and closure activities. If a spill were to occur at a stream crossing or near a stream, surface water could be impacted. Although not all waterbodies crossed via culverts are fish-bearing, spills into any waterway could travel downstream to fish-bearing waters.

As described in the BA, mine transport begins on Warm Lake Road (CR 10-579) where the risk of spills will be lower, as it is paved, maintained by Valley County, and has overall gentler grades. At the intersection of Warm Lake Road and Johnson Creek Road (CR 10-413) the two mine access routes begin, with the Johnson Creek Route north along Johnson Creek Road (CR 10-413) and the Burntlog Route east onto Burntlog Road (FR 447). The location of the spill risk will change as the SGP progresses under the proposed action. Johnson Creek and the portion of the EFSFSR between Yellow Pine and the Operations Area Boundary will be at risk of spills of hazardous materials during the first one to two years of the SGP when the Johnson Creek Route is used as the access route during the Burntlog Route construction. For the remainder of the mine life, the waterbodies along the Burntlog Route will be at risk for spills of hazardous materials.

Close proximity of access roads to surface water resources increases the potential for spilled material on the roadways to enter water, thus increasing the potential consequences of a spill. The Johnson Creek Route crosses 43 different streams and includes 27 miles of road that are within 0.5 mile of surface water resources, including several miles that parallel the fish-bearing EFSFSR and Johnson Creek waterways. The Burntlog Route crosses 37 streams, but only includes 9 miles of road that are within 0.5 mile of surface water resources. Though the Burntlog Route includes a greater number of actual stream crossings, the Johnson Creek Route includes significantly greater proximity to water resources occupied by spawning, migrating, and rearing ESA-listed salmonids. Because the majority of the Johnson Creek Route is immediately adjacent to Johnson Creek and the EFSFSR, the potential consequences from trucking spills will thus be greater along the Johnson Creek Route that will be utilized for transport while the Burntlog Route is being constructed.

As suggested in the BA, the most probable release scenario associated with truck transport on the access routes to the SGP will be relatively small amounts (less than 5 gallons) of fuel spilled from vehicles. Under this scenario, immediate cleanup actions to contain, recover, and remove the contaminated soils is anticipated. Because trucks and pilot cars will be equipped with spill response kits, fuel spilled to soils/roadbed should be readily contained and recovered, although fuel which enters waterways via roadside drainages may be difficult or impossible to fully recover and there will be potential for spread and downstream effects beyond the immediate spill area. The risk from spills of this magnitude should be minimal, as spill response materials on the vehicles are pre-positioned along the access routes, and SGP response vehicles will contain materials to not only clean up the site but also to contain and recover floating oil.

A more serious release of liquid petroleum or hazardous material from a bulk truckload could potentially occur assuming the puncture of the bulk tanker in an accident. Under this scenario, spilled material will be released to the immediate roadbed area, and potentially impact physical resources and ecological receptors (e.g., vegetation) and nearby surface water depending on the topography and location.

Table 19 identifies the materials, supplies, and reagents that will be delivered to the SGP. As bulk liquids, diesel, lubricants, gasoline, antifreeze, propane, solvents, Aerophine 3418A, AP 3477, methyl isobutyl carbonyl, sulfuric acid, magnesium chloride, and carbon dioxide present the highest risk of release should an accident occur. Of all the substances to be transported, fuel may pose the highest risk to fish and fish habitat with delivery of 5.8 million gallons of diesel and 0.5 million gallons of gasoline expected annually via tanker truck. This is because large quantities of fuel are transported in each load, numerous trips are made each year, and the substance is a liquid that rapidly flows down gradient toward nearby streams in the event of a spill.

Many of the streams with segments in proximity to access roads support Chinook salmon and/or steelhead. The intensity of the impact of a hazardous materials spill on fish and fish habitat could be high; as a large diesel spill could result in the mortality of 100 percent of the Chinook salmon and steelhead juveniles, adults, alevins, and eggs for a considerable distance (several miles) downstream from the accident (NMFS 1995). In terms of toxicity to water-column organisms, diesel is one of the most acutely toxic oil types. Fish, invertebrates, and aquatic vegetation that come in direct contact with a diesel spill may die (NOAA 2023). The severity of the impact will depend on the timing, size, and location of the spill. Small spills in deep open waters are expected to rapidly dilute; however, fish mortality has been reported for small spills in confined, shallow water (NOAA 2023).

As displayed in Table 19, diesel will be the substance hauled to the SGP most frequently (estimated 5.8 million gallons/year, 580 deliveries per year, 22% of all deliveries), while gasoline (500,000 gallons/year) and propane (2.02 million gallons/year) represent an additional 13% of the expected deliveries to the SGP.  $MgCl<sub>2</sub>$  will also be delivered in large quantities, estimated at 250,000 gallons/year, 56 deliveries per year, and as liquids, each will be hauled in large volume aluminum tanks that are readily punctured, and hauled in relatively close proximity to action area streams. These factors result in fuel spills during transport being the most likely type of accidental chemical contamination from the proposed action and the most likely

substance that ESA-listed salmon, steelhead, and their habitat may be exposed to from accidental spills within the action area. The potential effects from an accidental release of any of these liquids is described in more detail below.

*Diesel.* Of the substances being transported, diesel has been identified as posing the highest risk of a spill affecting ESA-listed salmonids and critical habitat. This is because diesel is delivered in large quantities (10,000 gallons), will be hauled to the mine site about 580 times per year, containers are aluminum and easily ruptured, and the substance is a liquid that can rapidly flow down gradient into nearby streams.

The 96-hour  $LC_{50}$ <sup>8</sup> for diesel and rainbow trout has a range of 18 to 25 mg/L (Conoco 2022). The wide range of toxicity for diesel reflects variations in the petroleum compounds which occur for each source of crude, where it was refined, and the time of year the petroleum was produced. Petroleum products which are refined for use during the summer have different toxicity than products refined during the winter because the additives used to maintain the desired viscosity are applied in different proportions.

Diesel and gasoline contain petroleum aromatic hydrocarbons (PAHs), which have been linked to severe developmental abnormalities at incredibly low concentrations (µg/liter) (NWFSC 2022). Severe PAH toxicity is characterized by complete heart failure, with ensuing extra-cardiac defects (secondary to loss of circulation) and mortality at or soon after hatching. More moderate forms of PAH toxicity, such as might be expected for untreated/unfiltered roadway runoff, include acute and latent alterations in subtle aspects of cardiac structure, reduced cardiorespiratory performance and latent mortality in surviving larvae and juveniles. Total PAH levels in the range of 5-20 µg/L resulted in cohorts of pink salmon (*O. gorbuscha*) that survived the exposure and appeared outwardly normal, but nevertheless displayed reduced growth and reduced survival to reproductivity maturity. Follow-up studies at the Northwest Fisheries Science Center have linked this poor survival to reduced individual fitness manifested by reduced swimming performance and subtle changes in cardiac structure. In essence, embryonic exposure to petroleum mixtures leads to juvenile fish that show signs of pathological hypertrophy of the heart (Incardona et al. 2015, 2021; Gardner et al. 2019).

In smaller streams, like those along the Burntlog Route, a relatively small spill, if not contained, could cause toxicity. In larger streams, like Johnson Creek or the EFSFSR, a spill of as little as 10 gallons could be toxic. Diesel spills have the potential to affect listed salmon, steelhead, and their habitat. Diesel spills have not been uncommon in Idaho, and have included but not necessarily limited to: (1) Little Salmon River (1993, 900 gallons diesel + 900 gallons gasoline), 133 dead rainbow trout and brook trout; (2) Lochsa River (2003, 6,300 gallons diesel), no documented fish kill; and the (3) Middle Fork Clearwater River (2002; 10,000 gallons diesel), no documented fish kill.

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<sup>8</sup> LCx (Lethal Concentration): The estimated concentration where a specified percentage of the test organisms die. The LCx is estimated using a statistical distribution or regression model of the dose-response relationship based on toxicity testing. The most common test statistic used in acute studies is the  $LC_{50}$  (the lethal concentration where 50% of the organisms die).

Two additional spills in the Salmon River basin warrant further discussion, representing local examples of diesel spills standing out as having negatively affected critical habitat and aquatic life. On August 19, 1983, the IDFG reported a diesel fuel spill of 2,800 gallons into the Little Salmon River. The IDFG calculated that approximately 30,000 fish total (all species) died in the kill and also reported aquatic insect mortality. Another spill occurred on September 6, 1989, occurring on Johnson Creek and resulting in 400 gallons of diesel reaching Johnson Creek. The precise effects of this spill and the emergency response to it were not well documented. However, given the toxicity of PAHs previously described, PAHs retained in the sediments may have led to development abnormalities. Also, depressed populations of aquatic insects were reported for 3.5 miles downstream, and diesel odor was evident in stream substrates a year after the spill. The IDFG notes do indicate seeing Chinook salmon and steelhead showing obvious signs of stress. Given the SGP will on average have approximately 15,000 gallons of diesel delivered per day, spilling even a portion of one delivery truck's diesel load into surface water source could have substantial adverse effects to water quality.

A large diesel spill of 10,000 gallons in Johnson Creek or the EFSFSR would impact water quality to the point that it will kill Chinook salmon and steelhead juveniles, adults, alevins, and eggs for miles downstream of the accident, depending on the time of year. Spills along the Burntlog Route would be more likely to affect rearing Chinook or steelhead, but could also affect staging or spawning adults or their redds should diesel transmit downstream to spawning areas.

Any diesel spilled into action area streams would tend to travel downstream in a slug and dissipate slowly. Diesel from a spill could mix with spawning gravels and sand and be retained in the stream substrate for a year or more, and thereby negatively affect salmon and steelhead eggs, alevins, and juveniles for several years. Large amounts of petroleum products can suffocate aquatic organisms by coating their gills. Diesel fuel, like most petroleum products, contains toxic organic compounds that adversely affect water appearance and odor.

*Gasoline.* CITGO (2023) reported that various grades of gasoline exhibited a range of lethal toxicity (LC<sub>100</sub>) from 40 mg/L to 100 mg/L in ambient stream water for rainbow trout. Chevron identifies a 96-hour  $LC_{50}$  for rainbow trout and unleaded gasoline at 2.7 mg/L (Chevron 2023).

Transported as a bulk liquid in 5,000-gallon containers, gasoline poses both a high risk of accident and a high risk of delivery to streams should an accident occur. Approximately 500,000 gallons will be delivered in 100 trips per year to the SGP. As with diesel, containers will be aluminum and easily ruptured, and, and as a liquid, it will be able to rapidly flow down gradient into nearby streams.

The bulk of the available literature on gasoline relates to the environmental impact of monoaromatic (benzene, toluene, ethylbenzene, xylenes: BTEX) and diaromatic (naphthalene, methylnaphthalenes) constituents. In general, non-oxygenated gasoline exhibits some short-term toxicity to freshwater and marine organisms, especially under closed vessel or flow-through exposure conditions in the laboratory. The components which are the most prominent in the water-soluble fraction and cause aquatic toxicity, are also highly volatile and can be readily biodegraded by microorganisms. This material is expected to be readily biodegradable following a spill (USDA Forest Service 2007; Chevron 2023).
The effects of even a 30-gallon gasoline spill into occupied streams would be extreme and would likely result in mortality of all life stages of ESA-listed salmon or steelhead present immediately downstream from the spill. As with diesel, the magnitude and extent of effect would vary dependent upon the amount of water in the receiving waterbody and the amount of material spilled. As with diesel, effects would be most likely to occur to fish of all life stages exposed in Johnson Creek and the EFSFSR during the first couple of years while the Burntlog Route is being constructed, and to rearing fish exposed in tributaries of Johnson Creek and the upper EFSFSR once the Burntlog Route is in use.

*Propane.* The safety data sheet for liquid propane (Global 2024) identifies a 96-hour LC<sub>50</sub> for rainbow trout at 1.38 mg/L, for fathead minnow (*Pimephales promelas*) at 3.20 mg/L, and water fleas (*Daphnia magna*) at 0.09 mg/L. Bioaccumulation is not expected based on the volatile nature of propane.

Transported as a bulk liquid in 6,000- and 11,000-gallon containers, propane poses a high risk of accident as approximately 2.02 million gallons will be delivered in 226 trips per year to the SGP. As with diesel, containers could be easily ruptured in the event of an accident. However, unlike diesel or gasoline, the effects of a propane spill would be less likely to contaminate local steams, as spilled propane liquid turns to vapor, dissolving in the air and not polluting soil and water resources (Foster Fuels 2024).

*Magnesium Chloride.* Magnesium is a naturally occurring mineral, found in both terrestrial soils and aquatic sediment, posing very little threat to the environment (Vincoli 1997). As described by El-Mowafi and Maage (1998), magnesium is an essential element for fish, where deficiency has been found to result in retarded growth, anorexia, sluggishness, high mortality, reduced ash content, and reduced concentrations of magnesium and calcium in the whole body and vertebrae. Magnesium is nearly insoluble (0.01%) and highly persistent in water, with a half-life >200 days. Anthropogenic sources of magnesium include discharges and spills from industrial and municipal waste treatment plants (Vincoli 1997).

Magnesium and its salts have slight acute toxicity to aquatic life, and little information exists regarding its short-term toxicity to plants, birds, or terrestrial animals. The EPA does not currently consider magnesium a hazardous substance (Vincoli 1997). Mount et al. (1997) reported 24-, 48-, and 96-hour LC<sub>50</sub>s of 3,520, 2,840, and 2,120 mg/L MgCl<sub>2</sub> for fathead minnows, and 24-, and 48-hour LC<sub>50</sub>s of 1,560 and 3,330 mg/L MgCl<sub>2</sub> for *D. magna*. For rainbow trout eggs exposed to MgCl<sub>2</sub>, the Pesticide Action Network's PAN Pesticides Database (2008) reported a mean 28-day LC<sub>50</sub> of 1,355 mg/L, ranging from a low of 119.9 mg/L to a maximum of 1,507 mg/L. After 14 days, Shearer and Åsgård (1992) noted elevated whole body concentrations of magnesium in fish exposed to 150 mg/L and 1,000 mg/L, further noting that fish exposed to the higher concentration were significantly smaller than fish exposed to lesser concentrations. Mortalities began to occur in rainbow trout after 2 days of exposure to the 1,000 µg/L concentration, increasing to 48% total mortality at the end of the experiment (14 days). In the absence of dietary magnesium, the authors concluded that rainbow trout are able to meet their magnesium requirement at waterborne concentrations of 46 mg/L.

MgCl<sub>2</sub> will be transported to the site in 4,500-gallon truckloads, delivered up to 56 times per year. Spilling even a portion of one delivery truck's load into surface water source could have adverse effects to water quality. Spills of large quantities of  $MgCl<sub>2</sub>$  into action area streams could impact water quality to the point that it is likely to kill aquatic invertebrates, Chinook salmon and steelhead juveniles, adults, alevins, and eggs downstream of the accident, depending on the time of year. Spills along the Burntlog Route would be more likely to affect rearing Chinook or steelhead, but could also affect staging or spawning adults or their redds should the spill transmit downstream to spawning areas.

*Solid Materials.* A release of large quantities of solid hazardous materials such as cyanide or antimony concentrate could also occur during transport. Breaches of the shipping containers for these materials in the case of an accident could release the solid materials to the ground where it will reside until response actions are taken to mechanically clean it up, along with any contaminated soil. Migration of these solid materials from the immediate release site will be less likely than for liquid materials, but could be possible in wet weather conditions.

Cyanide toxicity to fish in the project area is associated with catastrophic spill risk and acute toxicity including mass mortality. Free cyanide (CN-) kills fish by disrupting oxygen uptake at the cellular level leading to suffocation, hepatocyte necrosis, or other tissue damage (Davis et al. 2017; David and Kartheek 2015). Mass mortality of salmonids would occur at CNconcentrations approaching 45 µg/L, the median lethal concentration for rainbow trout (Barber et al. 2003). Antimony is far less toxic than cyanide, with acute toxicity not being observed in 24 hour exposures to 11,400 µg/L (Brook et al. 1986).

A large spill could potentially cause the mortality of a substantial number of adult Chinook salmon and steelhead depending on various factors (NMFS 1995). A spill in the fall months could result in the mortality of all the 1-year-old juveniles and zero age eggs/alevins, thus eliminating 2 years of Chinook salmon progeny.

A release of large quantities of solid hazardous materials such as cyanide or antimony concentrate during transport will be unlikely. Transportation of cyanide and antimony concentrate will represent less than one percent of the site truck trips. The overall incident rate involving hazardous materials is very low. Transportation of cyanide and antimony concentrate will represent less than one percent of the site truck trips. Shipment of these materials will occur in packaging and containments designed to prevent release to the environment even in the event of a traffic incident. Quick spill response and recovery measures will help to limit impacts.

*Overall Transportation Spill Risk.* It is expected the risk associated with a transportation spill large enough to negatively affect fish or aquatic habitat will generally be low but possible. An exception may be when materials are transported during inclement weather conditions, this could increase the risk to moderate. Spills during the winter will be easier to contain because spilled material will not penetrate frozen ground as readily as unfrozen ground, and snow could absorb the spilled material, in addition the visual contrast between snow and fuel could aid in cleanup. However, areas that are harder to access (e.g., remote or in a canyon) may increase the time it takes to access and cleanup a spill, creating the potential for fish or fish habitat to be in contact with a hazardous material longer and could impact more fish or fish habitat.

The available information supports the BA's determination that a large spill is unlikely to occur as a result of the proposed action. The BA described the likelihood of a spill occurring as being relatively small, stating that accident rates with release of hazardous materials range between 1 every 522 million miles to 1 every 714 million miles. There will be estimated 240,000 truck trips to the site, with not all trips associated with the transport of hazardous materials. Applying the round-trip distance from Cascade (approximately 150 miles) to that number of trips will be a travel distance of 36 million miles. This considered, the BA suggests zero to one accident involving a release of hazardous materials over the life of the SGP. This considered, the probability that a spill will occur and it will be delivered to action area streams is low.

A SPCC Plan will be developed prior to construction to establish procedures for responding to accidental spills and releases of petroleum products. In addition, a Hazardous Materials Handling and Emergency Response Plan will be developed prior to construction to address procedures for responding to accidental spills or releases of hazardous materials to minimize health risks and environmental effects. Although not specifically identified in the BA as documents that will be reviewed by the IARB, NMFS expects that both the SPCC Plan and the Hazardous Materials Handling and Emergency Response Plan will be provided for review by the IARB as finalized to ensure they both adequate and likely to be effective.

NMFS expects that the risk of a hazardous material spill should be effectively minimized by proven, proposed transport EDFs. For example, pilot vehicles will accompany all transports of fuel or hazardous materials between the SGLF and the Operations Area Boundary, and will carry spill response tools and materials, communications equipment, and all drivers trained in spill responses. Thus, response to a small-to-moderate spill of fuel or hazardous material during transit over the SGP access roads will essentially be immediate. Spill response and recovery measures such as containment, deployment of absorbent materials, removal of impacted roadbed material and vegetation, and deployment of water-based spill recovery materials and equipment will also be quickly applied to help to limit impacts.

## *2.10.1.1.2. Onsite Spills*

Aboveground storage tanks at the SGP will be used for fuels and other fluids, including gasoline, diesel fuel, MgCl<sub>2</sub>, lubricants, coolants, hydraulic fluids, and propane. Approximately 200,000 gallons of diesel fuel, 10,000 gallons of gasoline, 20,000 gallons of MgCl2, and 30,000 gallons of propane will be stored at the SGP in addition to a variety of materials, supplies, and reagents (Table 19). The storage tank facility for gasoline, diesel fuel, and propane will be located near the maintenance workshop with additional propane storage at the ore processing facility area, the underground portal area, and the worker housing facility. The aboveground storage tanks will be installed on containments sized to contain 110 percent of the capacity of the tank. Refueling will occur on concrete-paved areas designed to contain refueling spills (i.e., berms around their perimeters). There will be no below ground fuel storage or piping used for refueling. Storage management will be outlined in the SPCC Plan required by the CWA, which NMFS expects will be provided for review by the IARB for adequacy and effectiveness.

Should a spill occur onsite, spawning, rearing, and migratory habitat could all be affected in the EFSFSR. However, the combination of the proposed monitoring, planning, and EDFs for storage and handling of fuels and hazardous materials will effectively minimize the risk of accidental

releases during the storage, management, and use of hazardous materials. While the likelihood of a spill is low given the measures described above, if there is a spill, water quality impacts could be catastrophic to individuals exposed to harmful concentrations of hazardous materials depending on the type of material releases, the location of the spill, and the presence of Chinook salmon, steelhead, and critical habitat for each species in the affected area.

## *2.10.1.1.3. TSF Buttress, Process Water Pipeline, and Containment Ponds*

Design criteria for the TSF were established based on the facility size and risk using applicable dam safety and water quality regulations and industry best practices for the TSF embankment on a stand-alone basis; the addition of the buttress substantially increases the safety factor for the design to about double the minimum requirements. The upstream face of the TSF embankment and the Meadow Creek valley where the TSF impoundment will be located will be fully lined to minimize leakage (Stantec 2024). The final designs for the TSF will be brought to the IARB for review and verification. Compliance with IDWR's Dam Safety Division safety regulations (IDAPA 37.03.05 and IDAPA 37.03.06) (Table 2) requires an acceptable monitoring plan be in place to assure the TSF buttress and various containment ponds functions within the acceptable safety factors. Containment ponds are designed to contain the contents of process water pipelines and runoff from the pond and lined pipeline corridor from a 100-year, 24-hour storm event plus snowmelt. Given IARB review, and compliance with dam safety regulations, catastrophic failure of the TSF or the containment ponds is unlikely, and the effects of a catastrophic failure on DCH in the action area are not further considered.

## *2.10.1.1.4. Herbicide Application*

Noxious weed and invasive species plant control will be conducted according to methods described in the PNF's 2020 Programmatic Activities Biological BA (PNF 2020) and NMFS opinion (NMFS 2020) on that programmatic Weed treatments typically occur between April and November, depending on elevation. During this period, water quality could be affected at a time when all life stages of Chinook salmon and steelhead could potentially be exposed to herbicides, overlapping incubating eggs, rearing juveniles, or migrating/holding adults. The potential effects of weed treatment when applied as proposed have been described in detail in Section 2.5.4.1 of NMFS' opinion (NMFS 2020), incorporated here by reference and summarized below.

Application of herbicides may lead to contamination of surface water, which could then harm individual fish. Herbicides proposed for use could potentially drift to waterways or leach into soils, contaminate groundwater, and eventually show up in streams where listed fish spawn and rear. This risk depends on a number of variables, including but not limited to the rate of application, concurrent precipitation, herbicide degradation, solubility, and distance to water. An accidental spill of herbicides directly to waterways could also result in water quality conditions toxic to salmonid fish species and result in harm/death of ESA-listed fish.

Herbicides (and adjuvants) applied as proposed cannot be kept entirely out of the water. However, herbicides applied by the applicant are not expected to reach streams in concentrations that kill fish; although concentrations may be of sufficient magnitude to elicit short-term sublethal effects. The PNF's weed treatment program has been in place for over two decades, and implementation of weed control project design features (PDFs) developed over that long

timeframe are expected to effectively reduce the risk of chemical contamination associated with weed treatment activities.

Although direct lethal effects are not expected, the types of chemicals used are ones that can be capable of causing harmful sublethal effects. It is possible that water quality will be impacted so that individual or smaller groups of fish could be exposed to and experience sublethal effects. Herbicide applications will not occur over large contiguous areas, and most of the action area will not be subjected to spraying in any given year. Considering the low level of effects that may occur coupled with the very small impact area, we do not expect widespread or long lasting effects to the water quality PBF from herbicide application.

## *2.10.1.1.5. Surface and Groundwater Quality Analysis.*

The SGP will address some of the historical nonpoint sources of contamination in the action area and will create new point and nonpoint sources of contamination during construction, operations, and closure. Once in surface waters, the fate and transport of contaminants is dictated by biogeochemical processes (Alpers et al. 2000a, 2000b; Bricker 1999; Chadwick et al. 2004; Johnson et al. 2005). Contaminants may remain suspended in the water column, settle onto stream substrates, diffuse into interstitial pore spaces, or be taken up by benthic organisms, plankton, fish, or other species. Kraus et al. (2022) documented mercury movement from aquatic ecosystems to terrestrial ecosystems when aquatic invertebrates emerged as adults and were preyed upon by riparian spiders. Ultimately, the risk of toxicity from contaminant exposures is greatest within the mine site and is generally expected to decrease with downstream distance. Mercury is an exception to this general rule of thumb, because methylation is expected to continue to occur downstream.

A SWWC model was developed to evaluate potential impacts of the SGP on surface and groundwater quality. The model and its sensitivity analysis are fully described in the SWWC report (SRK 2021a), SWWC sensitivity analysis (SRK 2021b), and Water Quality Specialist Report (Section 7 in USFS 2023a). A summary of the modeling performed, including notable assumptions underlying the modeling is provided in Appendix E of this opinion.

As previously described (Section 2.9.2.1), baseline water quality conditions within the upper EFSFSR watershed have been, and continue to be, impacted by past mining activities. The proposed action involves removal and reprocessing of Bradley tailings (described in Section 1.7.6) and removal of SODA and Hecla Heap leach materials in the Meadow Creek valley, and closure of the Bradley tunnel in the Sugar Creek watershed. While these existing sources of contamination will be eliminated, the proposed action also involves handling and storing new sources of mineralized materials that are expected to leach major ions (e.g., calcium, magnesium, potassium, sulfate, etc.), TDS, and/or metals. These new nonpoint sources of contamination are expected to impact surface and groundwater quality during mine operations and closure. Potential nonpoint sources of chemical contaminants include: (1) ore stockpiles; (2) TSF buttress and embankment; (4) backfill in the YPP, HFP, and Midnight Pit; (5) pit walls newly exposed to oxygen and water; (5) groundwater outflow from the WEP lake; and (6) the TSF.

In addition to the nonpoint sources of contaminants, surface water quality will be impacted by wastewater treatment plant discharges. As stated in Section 1.7.10 of this opinion, the types of wastewater treated include sanitary wastewater discharged in the EFSFSR upstream of its confluence with Meadow Creek and mine contact water<sup>9</sup> discharged into either the EFSFSR below Meadow Creek or into Meadow Creek. Discharge into Meadow Creek will occur during operation if needed to augment streamflow. Discharge of treated water into Meadow Creek will occur during early closure. The SWWC model accounted for discharge from the mine contact water treatment plant to Meadow Creek and the EFSFSR during different time periods (Appendix E). The assumed effluent quality used in the SWWC model is summarized in Table 15. This assumed effluent quality means that mixing zones will not be authorized for the contaminants in the discharge and that the discharges will be required – in the relevant IPDES discharge permits – to meet applicable water quality criteria at the point of discharge, i.e. end of pipe.

The SWWC model made water quality predictions for assessment nodes in Meadow Creek, West End Creek, Sugar Creek, and the EFSFSR. The assessment nodes are downgradient of mine facilities (Figure 28). Predictions were made for existing conditions, operations period, earlyclosure period, and late-closure period. The operations period extends from MY -2 through MY 12. During operations, mine contact water (including pit dewatering water) that is collected and is in excess of water needed for ore processing will be treated and then discharged to surface water. The operations period includes the last two years of construction (i.e., MYs -2 and -1) because mine contact water will be generated and collected. The early-closure period includes MYs 13 through 40. This period begins when open pit mining is complete and it includes two years of continued processing of the stored ore stockpiles (MYs 13 and 14) as well as the earlyclosure years where consolidation water from the TSF continues to be collected, treated, and subsequently discharged. The late-closure period occurs after the TSF has completed its consolidation and all mine facilities have been fully reclaimed. During this period, water is no longer collected and treated. The tabulated predictions for each assessment node are presented in Appendix E (refer to Tables E-2 through E-10). Figures 28 through 32 graphically display the average and maximum water quality predictions for arsenic, antimony, copper, and mercury.

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<sup>9</sup> Examples of mine contact water include: snowmelt or stormwater runoff from development rock storage facilities, ore stockpiles, or other disturbed mining areas; toe seepage from the TSF buttress or ore stockpiles; pit dewatering; and underdrain seepage.



**Figure 29. Surface Water Quality Prediction Nodes (SRK 2021a).**



**Figure 30. Modeled average (A) and maximum (B) arsenic concentrations for existing conditions, operations, early closure, and late closure periods at assessment nodes within the project area.** 



**Figure 31. Modeled average (A) and maximum (B) antimony concentrations for existing conditions, operations, early closure, and late closure periods at assessment nodes within the project area.**



**Figure 32. Modeled average (A) and maximum (B) copper concentrations for existing conditions, operations, early closure, and late closure periods at assessment nodes within the project area.**



**Figure 33. Modeled average (A) and maximum (B) mercury concentrations for existing conditions, operations, early closure, and late closure periods at assessment nodes within the project area. Note the y-axis scale break.**

Considering the toxicity of contaminants is of paramount importance when assessing the ability of the water quality PBF to support spawning, rearing, and migration. To evaluate the risk of adverse effects from contaminant exposures, NMFS calculated a risk quotient by dividing the predicted instream concentrations by a selected toxicity threshold. The risk of adverse effects from contaminant exposure was then initially characterized as low (risk quotient  $< 0.5$ ), moderate (risk quotient  $0.5 - 0.99$ ) or high (risk quotient  $\geq 1.0$ ). Tables 46 and 47 summarize the results. Our review focused on those constituents with a moderate or high risk of causing adverse effects either directly (through waterborne exposures) or indirectly (through dietborne exposures). Table 48 identifies these contaminants for each stream and mine phase. Our assessment of mixture toxicity considered all contaminants and was not limited to just those receiving a more detailed review.

Cyanide is also considered a contaminant of concern given its toxicity and presence in the TSF. Even after accounting for liner leakages, cyanide was not included in the SWWC model because it is not expected to persist at detectable concentrations because geochemical conditions that favor rapid breakdown of the cyanide molecule exist at the site (G. Fennemore, personal communication, July 2, 2024). Cyanide will be present in the wastewater treatment plant effluent; however, cyanide is expected to rapidly breakdown in the receiving stream given sunlight exposure and circumneutral pH. The assumed permit limit of  $3.9 \mu g/L$  or less of WAD cyanide is below concentrations that would limit the ability of the water quality PBF to support spawning, incubating, rearing, and migrating life stages of fish. The greatest risk relative to cyanide is associated with accidental spills or any upset condition that releases process solution. Risks associated with these unlikely events were described previously.

Predicted concentrations for remaining metals not included in Table 47 (refer to Tables E-2 through E-10 in Appendix E) were well below their respective lowest effect evaluation thresholds; therefore, sublethal or lethal effects from exposures to those contaminants are considered extremely unlikely. As such, a detailed analysis of these constituents has not been conducted, because the predicted water quality concentrations of these metals are expected to continue to support the spawning, rearing, and migratory life stages of anadromous fish. Similarly, impacts on stream pH levels are expected to be minimal through all periods. Predicted pH levels range from 6.1 to 7.6, with an average of 7.0 to 7.3. These pH levels are expected to continue to support the spawning, rearing, and migratory life stages of anadromous fish within the action area.



### **Table 46. Predicted average and maximum monthly concentrations and calculated risk quotients for antimony and arsenic for specific stream reaches in the project area.**

**Key:** Pred. Conc. = Predicted Concentration;  $\mu g/L$  = microgram per liter

**Note:** Low risk quotients have no shading, Moderate risk quotients are shaded yellow, and high-risk quotients are shaded red.

<sup>1</sup>Highest concentrations among sample locations YP-T-27 and YP-T-22

<sup>2</sup>Highest concentrations among sample locations YP-SR-10 and YP-SR-8

3Highest concentrations among sample locations YP-SR-6 and YP-SR-4



**Table 47. Predicted average and maximum concentrations and calculated risk quotients for copper and mercury for specific stream reaches in the project area.** 

**Key:** Pred. Conc. = Predicted Concentration;  $\mu g/L$  = microgram per liter; ng/L = nanogram per liter

**Note:** Low risk quotients have no shading, Moderate risk quotients are shaded yellow, and high risk quotients are shaded red.

<sup>1</sup>Highest concentrations among sample locations YP-T-27 and YP-T-22

<sup>2</sup>Highest concentrations among sample locations YP-SR-10 and YP-SR-8

3Highest concentrations among sample locations YP-SR-6 and YP-SR-4

<b>Stream</b>	<b>Operation Period</b>	<b>Early-Closure Period</b>	<b>Late-Closure Period</b>			
Meadow Creek	Arsenic Copper Mercury Antimony	Arsenic Mercury Antimony	Arsenic Copper, Mercury Antimony			
EFSFSR (above YPP)	Arsenic Mercury Antimony <b>Total Dissolved Solids (TDS)</b>	Arsenic Mercury Antimony <b>TDS</b>	Arsenic Mercury Antimony			
EFSFSR (from YPP to above Sugar Creek)	Arsenic Copper Mercury Antimony <b>TDS</b>	Arsenic Copper Mercury Antimony <b>TDS</b>	Arsenic Mercury Antimony			
<b>Sugar Creek</b>	Arsenic Mercury Antimony	Arsenic Mercury Antimony	Arsenic Mercury Antimony			
Arsenic EFSFSR (below Sugar Cr) Mercury Antimony		Arsenic Mercury Antimony <b>TDS</b>	Arsenic Mercury Antimony			

**Table 48. Contaminants of particular concern identified with the site-wide water chemistry model for the SGP.**

Information regarding the toxicity of arsenic, antimony, copper, mercury, and TDS is provided in Appendix F, and briefly summarized below. We discuss the effects in terms of toxicity to salmonids because this provides direct insight into the degree to which the water quality PBF can support spawning, incubating, rearing, and migrating life stages of fish.

*Antimony.* Predicted antimony concentrations are summarized in Table 46. Antimony is not expected to bioaccumulate in salmonids, nor is it expected to cause acute mortality at predicted concentrations. Adverse sublethal effects that may arise from chronic exposure to antimony range from changes in growth, developmental abnormalities, and normal physiological function (e.g., locomotor behavior, swimming, development, and respiratory activities) (Xia et al. 2021; Brooke et al. 1986). The preponderance of studies documented toxicities to fish at concentrations orders of magnitude higher than those predicted to occur during any phase of the project. However, one study documented low levels of mortality at concentrations similar to those predicted for the proposed action. In that study (Birge et al. 1978), the  $LC_{50}$  and  $LC_1$  values for rainbow trout exposed to antimony from fertilization through 4 days post-hatch were reported to be 580  $\mu$ g/L and 28.6  $\mu$ g/L, respectively. The 95 percent confidence intervals around the LC<sub>1</sub> were quite large, ranging from a lower limit of  $\sim$ 5  $\mu$ g/L to an upper limit of  $\sim$ 72  $\mu$ g/L. Table 49 summarizes how the predicted antimony concentrations could impact the ability of the water quality PBF to support salmon and steelhead over the life of the project in various stream reaches.



### **Table 49. Biological response that will occur as a result of exposure to predicted antimony concentrations.**

<sup>1</sup>Snake River spring/summer Chinook salmon designated critical habitat.

<sup>2</sup>Snake River Basin steelhead designated critical habitat.

*Arsenic.* Predicted arsenic concentrations are summarized in Table 46. Arsenic (As) speciation is an important consideration when evaluating arsenic toxicity, with arsenite  $(As^{III})$  being more toxic than arsenate  $(As<sup>V</sup>)$ . Acute toxicity of arsenic occurs at very high concentrations (i.e., much greater than 1,000  $\mu$ g/L). Chronic toxicity may occur from longer term exposures to much lower waterborne concentrations; coho salmon experienced delayed mortality following a 6-month exposure to 33 µg/L (Nichols et al. 1984). Rainbow trout embryos were similarly sensitive, with reported LC<sub>1</sub> and LC<sub>10</sub> values of 42 and 134  $\mu$ g/L, respectively, following a 28-day exposure to arsenic (Birge et al. 1980). While these concentrations are low, dietborne exposure to arsenic is believed to be a greater threat to salmonid health than waterborne exposure. Dietborne exposures have been linked to liver and gall bladder damage, growth reduction, and reduced reproduction in fishes. These sublethal effects are thought to occur when arsenic concentrations in prey are approximately 20 mg/kg dry weight (dw) or greater, which can accumulate when water levels are about 10 µg/L. More recent literature further complicates the toxicity assessment by demonstrating arsenic speciation may be an important factor in toxicity. Erickson et al. (2019) found that dimethylarsinate and monomethylarsonate were far less toxic than inorganic arsenic. The authors present a compelling case for obtaining a better understanding of arsenic speciation in water and macroinvertebrates in order to better assess the risk of growth reductions and other sublethal effects in juvenile fish. One limitation of their study was utilizing older fish that may be more tolerant of arsenic-laden food relative to younger fry that have just begun exogenous feeding.

Dovick et al. (2016) found that  $As<sup>V</sup>$  was the dominant form of arsenic in surface water collected in the project area in 2010; although  $\text{As}^{\text{III}}$  accounted for up to 30 to 37 percent of the total arsenic in the lower reaches of Meadow Creek. Even if the waterborne concentration is comprised mostly of the less toxic form of arsenic, biota at the bottom of the food chain may contain a higher proportion of the more toxic form, which Erickson et al. (2019) attributed to the "reducing environment in the organism."

Average arsenic concentrations are predicted to decrease ten-fold compared to baseline conditions during operations due to removal of historic contamination in the valley. Yet, predicted maximum concentrations will exceed the 10 µg/L threshold identified in NMFS (2014) occasionally (i.e., for some years at the start of operations and during early closure); however, average concentrations are expected to be below the 10 µg/L. Table 50 summarizes how the predicted arsenic concentrations could impact the ability of the water quality PBF to support salmon and steelhead over the life of the project in various stream reaches.

<b>Stream</b>	Operations $(MY-2-12)$	Early Closure (MY13 to 40)	Late Closure (MY41+)				
Meadow Creek <sup>1</sup>							
	None expected from waterborne exposure.	None expected from waterborne exposure.	None expected from waterborne or dietary exposures.				
<b>Biological Response</b>	Juvenile fish may experience sublethal effects as a result of dietary exposures in years with prolonged maximum predicted arsenic concentrations.	Juvenile fish may experience sublethal effects as a result of dietary exposures in years with prolonged maximum predicted arsenic concentrations.					
<b>EFSFSR</b> above YPP <sup>1</sup>							
<b>Biological Response</b>	Low risk of mortality if redds or juvenile fish are chronically exposed to maximum predicted concentrations	Juvenile fish may experience sublethal effects as a result of dietary exposures.	Juvenile fish may experience sublethal effects as a result of dietary exposures.				
	Juvenile fish may experience sublethal effects as a result of dietary exposures.						
<b>EFSFSR</b> below YPP <sup>1,2</sup>							
<b>Biological Response</b>	Risk of mortality if redds are chronically exposed to maximum predicted concentrations.	Risk of mortality if redds are chronically exposed to maximum predicted concentrations.	Risk of mortality if redds are chronically exposed to maximum predicted concentrations.				
	Juvenile fish may experience sublethal effects as a result of dietary exposures.	Juvenile fish may experience sublethal effects as a result of dietary exposures.	Juvenile fish may experience sublethal effects as a result of dietary exposures.				
Sugar Creek <sup>1,2</sup>							
<b>Biological Response</b>	Juvenile fish may experience sublethal effects as a result of dietary exposures.	Juvenile fish may experience sublethal effects as a result of dietary exposures.	Juvenile fish may experience sublethal effects as a result of dietary exposures.				

**Table 50. Biological response that will occur as a result of exposure to predicted arsenic concentrations.**

<sup>1</sup>Snake River spring/summer Chinook salmon designated critical habitat.

<sup>2</sup>Snake River Basin steelhead designated critical habitat.

*Copper.* Predicted copper concentrations are summarized in Table 47. Copper is highly toxic to aquatic life. Toxicity to fish includes, but is not limited to, direct mortality, reduced growth, and reduced olfactory function. The estimated 96-hour  $LC_{50}$  for Chinook salmon is 7.4  $\mu$ g/L in test water at a pH of 7.7 and a hardness of 35 mg/L. Sandahl et al. (2007) reported reduced olfaction function after short-term (i.e., 3 hours) exposures to copper concentrations as low as increases in copper concentrations by as little as 0.18 µg/L over background (where background was identified as 3 µg/L). Hansen et al. (1999a) documented avoidance responses in juvenile Chinook salmon and rainbow trout at concentrations as low as 0.75 ug/L in soft water. Morris et al. (2019) documented a 20 percent reduction in alarm cue response after being exposed to copper concentrations of 2.7 and 2.5  $\mu$ g/L for 24 or 96 hours, respectively. Reported EC<sub>10</sub> values for reduced growth in Chinook salmon (Chapman 1982) and rainbow trout (Marr et al. 1996) chronically exposed (i.e., 120 days) to elevated levels of copper were 1.9 µg /L and 2.8 µg /L, respectively. Mebane and Arthaud (2010) employed regression models describing the relation between juvenile Chinook salmon growth and copper concentrations. Based on their models, the corresponding length reductions associated with exposure to 2.1 µg/L copper ranged from zero to 4.5 percent.

It is evident that sublethal effects, such as reduced growth, avoidance, and reduced olfactory function can occur at very low copper concentrations, with the later response occurring after short exposure durations. Copper concentrations are generally expected to be less than 1  $\mu$ g/L and are not expected to result in reduced growth; however, maximum predicted concentrations are at levels that could impart some low level of effects (e.g., reduced olfactory function, avoidance) to rearing juvenile salmonids or migrating adults. Inhibition of olfaction has been shown to reduce juvenile salmonids' ability to evade predation, so it is possible that juveniles could die as a result of copper concentrations at the site. Table 51 summarizes how the predicted copper concentrations could impact the ability of the water quality PBF to support salmon and steelhead over the life of the project in various stream reaches.

<b>Stream</b>	Operations $(MY-2-12)$	Early Closure (MY13 to 40)	Late Closure (MY41+)				
Meadow Creek <sup>1</sup>							
<b>Biological Response</b>	There may be some sublethal effects such as reduced olfaction or avoidance at maximum predicted concentrations	There may be some sublethal effects such as reduced olfaction at maximum predicted concentrations	None expected				
<b>EFSFSR</b> above YPP <sup>1</sup>							
<b>Biological Response</b>	There may be some sublethal effects such as reduced olfaction or avoidance at maximum predicted concentrations	There may be some sublethal effects such as reduced olfaction or avoidance at maximum predicted concentrations	There may be some sublethal effects such as reduced olfaction or avoidance at maximum predicted concentrations				
<b>EFSFSR</b> below YPP <sup>1,2</sup>							
<b>Biological Response</b>	There may be some sublethal effects such as reduced olfaction or avoidance at maximum predicted concentrations	There may be some sublethal effects such as reduced olfaction or avoidance at maximum predicted concentrations	There may be some sublethal effects such as reduced olfaction or avoidance at maximum predicted concentrations				
Sugar Creek <sup><math>1,2</math></sup>							
<b>Biological Response</b> $\bullet$ $\bullet$ $\bullet$ ית ו-10	There may be some sublethal effects such as reduced olfaction at maximum predicted concentrations.	There may be some sublethal effects such as reduced olfaction at maximum predicted concentrations.	There may be some sublethal effects such as reduced olfaction at maximum predicted concentrations.				

**Table 51. Biological response that will occur as a result of exposure to predicted copper concentrations.** 

<sup>1</sup>Snake River spring/summer Chinook salmon designated critical habitat.

<sup>2</sup>Snake River Basin steelhead designated critical habitat.

*Mercury*. Predicted mercury concentrations are summarized in Table 47. Mercury is a potent neurotoxin that causes neurological damage, which in turn can lead to sublethal effects that can impair growth and reproduction. Predicted mercury concentrations are not expected to cause sublethal or lethal effects via waterborne exposures; however, predicted mercury concentrations may lead to sublethal effects as a result of dietary exposures. Substantial work has been done to relate water column concentrations to elevated fish tissue concentrations, and deserves attention here (effects through dietborne exposures are also addressed in Section 2.10.1.3 for effects to the prey PBF).

The proposed action will contribute additional mercury to the stream network through groundwater, overland flow, wastewater treatment plant discharge, and atmospheric deposition. The greatest increases will occur primarily during the operating and early post-closure periods. The degree to which these increases will translate to increased bioaccumulation is influenced by a myriad of site-specific factors as described in Appendix F. Those factors include the amount of bioavailable inorganic mercury, presence of suitable methylation conditions (e.g., pH, organic carbon, bacterial communities and activity, dissolved oxygen, wetland abundance), growth rate efficiencies, and trophic position in the food web.

NMFS (2014) established a linear relationship between total mercury concentrations in the water column and total mercury concentrations in fish tissue using paired data reported by Essig (2010). Applying this linear regression equation to average and maximum predicted total mercury concentrations yields potential fish tissue concentrations ranging from 0.172 to 0.528 mg/kg ww and 0.231 to 0.759 mg/kg ww, respectively. Available water column and fish tissue information suggests that these predictions may overestimate future realized tissue burdens. While we were not able to access readily available paired water column and fish tissue information for the action area, the range of water column concentrations collected as part baseline monitoring from 2012 through 2016 in Meadow Creek, the EFSFSR, and Sugar Creek are shown in Table 52. Also shown are estimated fish tissue concentrations based on the average and maximum water column values.





 $Ng/L =$  nanograms per liter; mg/kg ww = milligrams per kilogram wet weight; min = minimum; max = maximum; EFSFSR = East Fork South Fork Salmon River

1MWH 2017

Documented Westslope cutthroat trout (*O. clarkia lewisi*) whole body fish tissue concentrations in Meadow Creek, EFSFSR, and Sugar Creek range from 0.02 to 0.05 mg/kg ww, 0.02 to 0.07 mg/kg ww, and 0.07 to 0.09 mg/kg ww, respectively. Fish tissue samples were collected in 2015 (MWH 2017). These fish tissue burdens are well below levels that have been linked to sublethal effects. Documented whole body mercury concentrations in bull trout collected from the lowermost Sugar Creek site in 2016 ranged from 0.101 to 0.314 mg/kg ww (Rutherford et al. 2020). Documented westslope cutthroat trout fish tissue concentrations are generally about an order of magnitude less than what is predicted from the linear regression. The difference between predicted and actual fish tissue concentrations is likely due to the aquatic systems' net methylation potential and food web composition.

The proposed action is expected to cause incremental increases in total mercury in the water column of streams inhabited by salmon and steelhead. These increases could lead to greater bioaccumulation of mercury in fish tissues. Additionally, an incremental increase in organic carbon content in the EFSFSR due to sanitary wastewater effluent could increase in methylation potential in the EFSFSR, contributing to increased bioaccumulation. For these reasons, the ability of the water quality PBF to support rearing juvenile salmon and steelhead will be further reduced as a result of the proposed action.

*Total Dissolved Solids***.** Increased concentrations of TDS are a potential effect associated with mining projects. Increased TDS concentrations typically result from sulfide oxidation and acidneutralization reactions that make constituents such as sulfate, calcium, and metals available for dissolution in surface water and groundwater. The SGP has limited sulfide mineralization present in its non-ore materials; therefore, the potential to leach major ions that contribute to TDS from these materials is limited. Ore material is capable of generating higher TDS concentrations, particularly as it is processed. The process solutions will be treated to reduce TDS concentrations prior to discharge (treatment target of 500 mg/L). The low TDS generating potential of the nonore materials and treatment of process water result in little change in predicted surface water TDS concentrations.

Predicted TDS concentrations in all streams during all phases of the mine are well below than concentrations that have been associated with adverse effects to salmon and steelhead. As such, predicted changes in TDS will not reduce the ability of the water quality PBF to support spawning, rearing, and migration.

*Contaminant Mixtures.* In addition to the effects described for the individual metals, mixtures of metals can also have interactive effects leading to toxicity at lower or higher levels than from a single metal alone (Finlayson and Verrue 1982). Predicting whether the complex mixture of contaminants will have synergistic, antagonistic, additive, or independent interactions is difficult. In some studies, mixtures of metals have shown both additive toxicity (with adverse effects observed at mixture concentrations of one-half to one-third the approximate toxicity threshold of metals in isolation (NMFS 2014). However, in other studies of metal mixtures, metal combinations have been less toxic than their single-metal toxicities (Finlayson and Verrue 1982; NMFS 2014). There is considerable uncertainty about the toxicity of the predicted future chemical mixtures in the EFSFSR, Meadow Creek, and Sugar Creek. It is not possible to rule out

the potential for any adverse effects to the water quality PBF due to chemical mixtures with any reasonable certainty.

Given the complexities of predicting the toxicity of metal mixtures, an integrated approach to water quality management is often taken by measuring and controlling toxics using a combination of numeric criteria, whole effluent toxicity (WET) testing, and biological monitoring of streams that receive point or nonpoint discharges. Only with ongoing chemical and biological monitoring can we continue to evaluate the potential for mixture toxicity, detect potential issues, and subsequently trigger adaptive management to ameliorate the situation

Contaminant toxicity is often exacerbated by increased temperatures because elevated temperatures accelerate metabolic processes and thus the penetration and harmful action of toxicants. Changes in stream temperatures within the Stibnite project area due to project implementation are not expected to be substantial enough to enhance toxicity. Climate change is expected to exacerbate the impacts of the proposed action, making it more difficult (though not impossible) to achieve the performance objectives for temperature. Adherence to the performance objectives relevant to stream temperatures will adequately minimize the risk of enhanced toxicity due to climate change.

*SWWC Model Exclusions*. The SWWC model excluded stormwater runoff from haul roads, access roads, maintenance facilities, and the residential camp. Constituent leaching from haul roads and access roads by meteoric and snowmelt runoff was evaluated using the site-wide water balance to estimate flows and humidity cell data to estimate runoff water chemistry. Leachate chemistry from road surface materials is predicted to have circumneutral pH with analyte concentrations below surface water quality criteria (SRK 2021a). Use of chemical additives for dust control on roadways is not expected to add constituents to surface water for the following reasons. Dust control products, such as MgCl<sub>2</sub>, CaCl<sub>2</sub>, lignin sulfonate, or other appropriate and environmentally-acceptable products, to further enhance dust control at the site will be used. The USFS will require that where haul roads pass within 25 feet (slope distance) of surface water, dust abatement will only be applied to a 10-foot swath down the centerline of the road. The rate and quantity of application will be regulated to ensure the chemical is absorbed before leaving the road surface. Therefore, effects of dust abatement chemicals in stormwater runoff from haul roads and access roads were not incorporated into the water chemistry modeling, but were incorporated into the analysis of sediments and hazardous materials.

Construction and operation of the Burntlog Maintenance Facility and worker housing facility will have the potential for increased runoff, erosion, sedimentation (as a result of vegetation removal and excavation of soil, rock, and sediment) and fuel and/or material discharge to nearby waterbodies during operations (if not properly stored or contained) and early closure. Implementation of EDFs proposed by Perpetua, regulatory and Forest Plan requirements required by the USFS, and permit stipulations from state and federal agencies (including BMPs, sanitary wastewater, and SPCC Plan) is expected to minimize runoff, erosion, sedimentation, and the potential for discharges. For these reasons, effects associated with stormwater runoff from the Burntlog Maintenance Facility were not quantified using the SWWC model. Although not modeled, stormwater discharges are reasonably certain to occur and will affect surface water quality. Stormwater that is not mine contact water will be generated from the

maintenance facilities, worker housing facility, haul roads and other roadways. Sources of pollution in this stormwater includes vehicle-related contaminants that accumulate on paved roads and parking lot surfaces due to increased vehicular use (McIntyre et al. 2015; Peter et al. 2018; Spromberg et al. 2016), as well as contaminants that accumulate on the building rooftops (WDOE 2014). This stormwater will be managed through application of EDFs aimed at capturing runoff and directing it to stormwater basins where sediment can collect and water can evaporate, infiltrate into the ground, or be discharged to nearby surface water. Even with application of EDFs, discharges into surface water are expected to contain an array of contaminants, most notably PAHs, tire wear particles, and some metals. The toxicity of PAHs was previously described, and best available information regarding the toxicity of tire wear particles is below.

Tire wear particles contain 6PPD [N-(1, 3-dimethylbutyl)-N'-phenyl-p-phenylenediamine], which can transform to 6PPD-quinone (6PPD-q) in the presence of ozone. 6PPD-q was identified by Tian et al. (2021) and has been linked to mortality of salmon and steelhead following acute exposures to extremely low concentrations. Available science suggests variation in species sensitivities with coho salmon being most sensitive followed by steelhead and then Chinook salmon (Brinkman et al. 2022; Chow et al. 2019; French et al. 2022; Lo et al. 2023; McIntyre et al. 2018; and McIntyre et al. 2021). Lo et al. (2023) reported LC<sub>50</sub> values of 41 ng/L and ~67  $\mu$ g/L for newly feeding ( $\sim$ 3 weeks post swim-up) coho and Chinook salmon, respectively, following a 24-hour exposure to 6PPD-q. Acute mortality from exposures to this compound can be prevented by infiltrating road runoff through soil media containing organic matter, which removes 6PPD-q and other contaminants (Fardel et al. 2020; Spromberg et al. 2016; McIntyre et al. 2015). No information exists regarding the sublethal toxicity of 6PPD-q; however, the overall field of 6PPD-q toxicity is rapidly evolving.

While stormwater runoff from non-mine contact water was not included in the SWWC model, it is reasonable to conclude that pollutants from stormwater generated from increased vehicular use on haul roads and other roadways, new parking areas, and new facilities will contain pollutants that can exert toxic effects. Feist et al. (2017) found that road density and traffic intensity was positively related to coho salmon mortality. At its peak, during construction, the proposed action will contribute 65 annual average daily trips to the project area. The increased traffic is expected to release additional vehicle-related contaminants onto the road surface, which will be captured during runoff events. The vast majority of runoff from Warm Lake Highway is directed to a vegetated right of way where it infiltrates into the ground. Some stormwater runoff is expected to enter streams, particularly in locations where the road crosses the stream channel. Untreated stormwater has the potential to reduce the water quality PBF in localized areas immediately following rain events. However, considering the traffic intensity associated with the proposed action, it is unlikely that water quality would be reduced to a degree that would reduce its ability to support spawning, incubating, rearing, and migrating life stages of salmon and steelhead.

The SWWC model also excluded sanitary wastewater treatment plant discharge into the EFSFSR upstream of Meadow Creek. The discharge volume from the wastewater treatment plant will vary between the mine construction, operation, and closure and reclamation periods, depending on the number of workers present at the SGP. Influent flows during construction and operations are estimated to be approximately 0.10 and 0.05 cfs, respectively (Brown and Caldwell 2020a).

We assume influent flows are representative of effluent flows. Sewage effluent systems will have waste containment and runoff control structures to prevent escape of untreated waste to the EFSFSR. The effluent is expected to contain phosphorus, nitrogen, caffeine, pharmaceuticals, personal care products, plasticizers, food additives, flame retardants, microparticles, and per and polyflouryl alkyl substances (PFAS) (Bothfeld 2021). The toxicity of these emerging contaminants of concern are poorly understood, but a number of authors have reported a variety of adverse sublethal effects (e.g., vitellogenesis, decreased whole body lipid content, reduced brain sodium-potassium ATPase activity, liver damage, and reductions in plasma glucose) in fish exposed to diluted wastewater treatment plant effluent (Ball et al. 2024; Meador et al. 2018; Popovic et al. 2023). These types of sublethal effects can impact individual fitness by altering behaviors and reducing survival (unable to evade predators, unable to successfully compete for food, etc.). The discharge will represent a very small proportion of the EFSFSR streamflow (~1 to 3 percent); however, dilution alone is not able to fully ameliorate the potential for adverse effects (Ball et al. 2024). It is possible that rearing juvenile salmonids in the EFSFSR in the vicinity of the discharge could experience adverse sublethal effects from continuous exposure to constituents in the effluent.

*Uncertainty.* When evaluating the impacts to the water quality PBF, we have assumed no untreated contact water at the ground surface (e.g., pond overflow or failures in contact water collection) will be discharged directly into surface waters and the SWWC model outputs represent the best available information regarding SGP impacts.

We recognize there is a substantial amount of uncertainty in whether the predicted instream concentrations represent actual conditions that will exist during operations and closure. Predicting future instream concentrations is incredibly difficult and there is uncertainty surrounding the validity of the assumptions employed. A variety of factors can impact the ability to accurately predict water quality changes including: inadequate geochemical characterization; inadequate hydrologic characterization (e.g., overestimating dilution, underestimating volume of contact water, underestimating storm size); and ineffective implementation of mitigation measures (e.g., liner leaks more than anticipated, underdrain system unable to capture all tailings leakage, ineffective segregation of rock material, etc.) (Kuipers et al. 2006). Some of the assumptions used in the various modeling efforts are conservative (e.g., no attenuation of solutes) while others may be less conservative (e.g., proportion of material effectively contacted by meteoric waters). Employing model outputs from one model as inputs to a different model results in additional uncertainty in predicting future water quality conditions during project implementation and following closure. For purposes of our opinion, we have assumed that the uncertainty associated with these factors have been appropriately minimized and the modeling represents the best scientific and commercial information available.

In addition to modeling uncertainty, the SWWC model does not account for contribution of contaminants from aerial deposition, sanitary wastewater treatment plant discharge, or stormwater runoff (beyond that which is collected and treated). As previously described, we have assumed that implementation of appropriate stormwater erosion control BMPs will minimize contributions of contaminants from this pathway. Air emissions from the SGP have the potential to contribute metals to the ground surface via wet and dry deposition which will subsequently have the potential to affect surface water chemistry. Actual local mercury deposition rates from

SGP emissions depend on the fractions of particulate versus gaseous mercury emissions. Particulate emissions generally deposit on the ground surface nearer to their source while gaseous emissions tend to deposit farther from the source or potentially become part of global atmospheric mercury burden. Most of the SGP contributions will be in the form of particulate matter, but a portion of the local aerial deposition of mercury may also occur in elemental form. Total mercury emissions from the SGP are predicted to be approximately 13.6 pounds of mercury per year. Mercury deposition rates from the air quality analysis of SGP emissions are predicted to be  $0.056$  g/km<sup>2</sup>/year compared to baseline deposition rates estimated to be between 12.7 and 13.9 g/km<sup>2</sup>/year (Air Sciences 2021). This suggests there would be a fractional increase (<0.5 percent) in the amount of mercury deposited in the area.

Ratios of stream mercury loads to atmospheric mercury deposition rates have been reported in watersheds affected by gold and silver mining (Domagalski et al. 2016). The effects of aerial mercury deposition on stream loads are variable based on watershed area, mineralization present, land development, rainfall, and soil adsorption characteristics. In smaller watersheds hosting precious metal mining, total mercury stream loads are higher relative to the mass associated with aerial deposition. This is because erodible sediments contribute relatively more to the stream load. Contributions from aerial deposition appear in stream loads over time as deposited mercury retained in soils is re-mobilized by local precipitation. Because the ratios reported in Domagalski et al. (2016) are variable and dependent on site-specific characteristics, they were not quantitatively applied for the analysis of the SGP watershed.

Ensuring a rigorous effectiveness monitoring and adaptive management program is essential in light of these uncertainties. Perpetua has incorporated an environmental monitoring and adaptive management program into the SGP (refer to Sections 1.10 and 1.11) and committed to finalizing that program following issuance of federal and state permitting and regulatory decisions (Brown and Caldwell 2021d). Furthermore, NMFS expects that the monitoring and mitigation plans under the EMMP (including the adaptive management) will be provided for review by the IARB to ensure the programs will be adequate and effective. Implementation of a thorough and robust monitoring program will enable Perpetua and the USFS to identify contaminant concentrations that are exceeding predictions relied upon in this analysis. The ongoing monitoring and adaptive management processes as described in the EMMP (Brown and Caldwell 2021d) and its component plans are key to ensuring effects do not rise to levels greater than predicted and analyzed. Ongoing monitoring, evaluation, and discussions with the IARB are expected to increase the probability of identifying issues early. In addition, the requirement to identify and implement appropriate responses to address issues not foreseen and adaptively manage the project and requiring adaptive management should reduce the magnitude and duration of unanticipated water quality impacts.

#### *2.10.1.1.6. Water Temperature.*

In order to support adult migration and spawning, incubation, and juvenile rearing, 7DADM water temperatures should be below 18, 13, and 16°C, respectively (EPA 2003). As previously mentioned, DCH for Chinook salmon occurs throughout the action area; whereas DCH for steelhead is more limited in the action area and excludes habitat upstream of the YPP. The Water Quality Specialist Report (Section 7 in USFS 2023a) provides details and references regarding the predicted surface water temperatures during the construction, operations, and

closure/post-closure periods. Water temperatures at the mine site will be impacted as a result of channel relocation and reconstruction, EFSFSR diversion tunnel construction, riparian vegetation removal, and point source discharges. Table 45 identifies the mine years where major changes to physical stream habitats are expected to occur and is incorporated by reference here.

Potential temperature impacts in watersheds along the transportation and transmission line corridors are expected to be minor because these activities involve little riparian disturbances. The transmission line ROW overlaps with 132.4 acres of RCAs (including non-anadromous RCAs) (Stantec 2024), of which 14.8 acres will be new disturbance associated with the new powerline ROW. The utility poles are not directly along creeks or within RCAs, so clearing of construction pads for footings will not reduce riparian vegetation. However, some vegetation removal may occur if trees exceeds the maximum height specified for the wire and border zones in the RCA. Because lower-growing vegetation will be maintained along the ROW, stream temperatures are not expected to be measurably altered. Construction of new access roads (including the Burntlog Route) will involve multiple stream crossings, but the new routes will generally not closely parallel or constrict streams. Most of the constructed and retrofitted stream crossings will occur high in tributary headwaters. Although removal of trees will be avoided where possible, there will be localized area where vegetation will be removed and stream shading will be reduced. Given the location in the watersheds where this will occur, and given the limited area of impact, any realized temperature increases will be minor.

The greatest temperature impacts are expected to occur within the mine site area, and are the focus of this evaluation. Under the SGP, there will be limited mining activity in the Sugar Creek watershed, with the majority of temperature-related effects being associated with diverting the West End Creek around the WEP. West End Creek is not fish bearing and contributes relatively minor flow volumes to Sugar Creek (i.e., West End Creek inflow [mean flow of 0.51 cfs] is approximately 2 percent of Sugar Creek flow [21.2 cfs]). Predicted increases in stream temperature will be between 0.1 and 0.3°C, with predicted maximum summer temperatures ranging from 15.5°C to 15.7°C compared to a baseline temperature condition of 15.4°C. Predicted maximum fall temperatures will range from 12.2°C to 12.3°C, compared to a baseline temperature condition of 12.2°C (Brown and Caldwell 2021a). Resultant temperature changes in Sugar Creek are not expected to reduce the ability of temperature conditions to support spawning, incubation, rearing, and migration.

In order to quantitatively evaluate the potential effects of the SGP, a Stream and Pit Lake Network Temperature (SPLNT) model was developed by Brown and Caldwell (2021a). The SPLNT model was used to first simulate existing conditions, expressed as a maximum weekly maximum temperature ( $MWMT^{10}$ ). The existing conditions model was developed and calibrated primarily using extensive site-specific meteorological, hydrologic, and stream data collected at the mine site (Brown and Caldwell 2021a, Appendix A). The model uses widely accepted numerical modeling approaches that consists of stream temperature and shading models (QUAL2K) and the General Lake Model for simulating pit lake temperatures. Results of the SPLNT model describing existing conditions (maximum weekly summer [July] and fall [September] temperatures) are shown in Table 53.

 $\overline{a}$ <sup>10</sup> The MWMT is the highest recorded 7DADM in a given season or period of interest.

#### **Table 53. SPLNT Modeled Baseline Maximum Weekly Summer and Fall Stream Temperatures for Specific Stream Reaches.**



Key: C = degrees Celsius; EFSFSR = East Fork South Fork Salmon River; EFMC= East Fork Meadow Creek; SPLNT = Stream and Pit Lake Network Temperature; YPP = Yellow Pine Pit

<sup>1</sup>Summer temperatures are represented by July daily maximum temperatures

<sup>2</sup>Fall temperatures are represented by September weekly maximum temperatures.

After the existing conditions SPLNT model was calibrated, it was used to generate future temperature predictions in Meadow Creek, West End Creek, Sugar Creek, and the EFSFSR resulting from SGP implementation. A post closure timeline also was simulated to represent how the site will function after the mine facilities and permitted discharges have been removed, dewatering and mining have been discontinued, and the channels and vegetation have been fully reclaimed.

The following years were selected for simulation in the SPLNT modeling: MYs 6, 12, 18, 22, 27, 32, 52, and 112. Mine year 6 was selected to represent the construction and operations periods because it is considered to represent the warmest conditions associated with the SGP prior to the YPP reclamation activities in MY 12. According to the USFS (Stantec 2024), characteristics that make MY 6 the warmest year for the construction and operations periods are:

- Largest project disturbance footprint and removal of riparian shading effects;
- Riparian plantings associated with early time restoration and reclamation activities will not have time to provide stream channel shading;
- Peak dewatering operations will result in the lowest groundwater elevations and lowest groundwater recharge to streams; and,
- Treated water discharges of excess dewatering water will be highest with the greatest volume contribution of warmer treated water to streams.

Simulation of MY 6 also includes the effects of the Meadow Creek diversion channel and low flow pipelines around the TSF area that will be installed at the start of construction and remain in use until closure of the TSF and restoration of the Meadow Creek channel on top of the reclaimed TSF. This is expected to result in stream cooling, which will be consistently present starting in MY -3 until completion of the TSF reclamation, channel restoration, and associated mitigation measures as necessary (anticipated in MY 27).

MY 12 represents the discontinuation of the use of the fish tunnel and the restoration of surface water flow across the restored stream channel across the YPP and through the Stibnite Lake feature but prior to the reestablishment of riparian vegetation along that restored channel segment. MY 18 represents the conditions at the end of mining operations. MY 22 represents the conditions following completion of most closure activities with the notable exception of the TSF closure and restoration of the Meadow Creek channel on the reclaimed TSF surface. Restoration of Meadow Creek on the reclaimed TSF will take longer to complete due to the need to dry the TSF surface tailings so equipment can access the TSF surface to install the reclamation cover. MY 27 represents the conditions following completion of the TSF closure and restoration of the Meadow Creek channel on its reclaimed surface but prior to the reestablishment of riparian vegetation. This includes discontinuation of the use of the Meadow Creek diversion around the TSF (i.e., via a constructed channel and low flow pipelines) and resumption of surface flow in the restored Meadow Creek channel. MYs 32, 52, and 112, represent interim and final conditions associated with the reestablishment of riparian vegetation associated with restored stream segments.

A summary of water temperatures predicted to occur as a result of SGP implementation are presented in Table 54. A key assumption of the SPLNT model is that revegetation success will be achieved as identified in the Reclamation Closure Plan and Stream Restoration Design (refer to Section 1.10.2.1).

<b>Stream</b> (Life Stages)			<b>Mine Year</b>							<b>Change from Baseline</b>				
	<b>Season</b>	<b>Baseline</b> $({}^{\circ}C)$	6	12	18	22	27	32	52	112	to 6	to $27$	to $52$	to $112$
			$({}^{\circ}{\rm C})$	$({}^{\circ}{\rm C})$	${}^{\circ}C$	$({}^{\circ}{\bf C})$	$({}^{\circ}{\rm C})$	$({}^{\circ}{\rm C})$	${}^{\circ}C$	$({}^{\circ}{\bf C})$	$({}^{\circ}{\rm C})$	$({}^{\circ}C)$	${}^{\circ}C$	$({}^{\circ}{\rm C})$
EFSFSR upstream from Meadow Cr.	Summer	13.4	13.3	13.3	13.3	13.3	13.3	13.3	13.3	13.3	$-0.01$	$-0.01$	$-0.01$	$-0.01$
(spawning, rearing, migration)	Fall	11.0	10.9	10.9	10.9	10.9	10.9	10.9	10.9	10.9	$-0.01$	$-0.01$	$-0.01$	$-0.01$
Meadow Cr. upstream from EFMC (spawning, rearing, migration) $3$	Summer <sup>1</sup>	14.0	12.4	12.3	12.4	12.4	20.8	<b>18.6</b>	17.1	15.1	$-1.6$	6.8	3.1	1.1
	Fall <sup>1</sup>	12.0	10.5	10.5	10.5	10.5	16	13.8	12.7	11.3	$-1.5$	4.0	0.7	$-0.7$
	Summer <sup>2</sup>	16.8	13.5	13.0	13.1	13.1	21.7	20.2	18.5	16.0	$-3.3$	4.9	1.7	$-0.8$
	Fall <sup>2</sup>	14.2	11.2	11.0	11.1	11.0	15.9	14.4	13.1	11.5	$-3.0$	1.7	$-1.1$	$-2.7$
Meadow Cr. downstream from EFMC	Summer	19.4	17.6	16.5	6.3	16.1	<b>18.5</b>	17.9	16.6	15.2	$-1.8$	$-1.4$	$-2.8$	$-4.2$
(spawning, rearing, migration)	Fall	15.9	15.5	13.6	$\overline{3.2}$	13.0	13.9	13.3	12.4	11.6	$-0.4$	$-2.0$	$-3.5$	$-4.3$
<b>EFMC</b>	Summer	14.6	15.8	15.4	15.3	15.2	14.9	14.8	14.4	14.2	$\overline{2}$	0.3	$-0.2$	$-0.4$
(rearing)	Fall	12.6	13.5	13.1	2.9	12.8	12.8	12.6	12.4	12.3	0.9	0.2	0.0	$-0.3$
<b>EFSFSR</b> between Meadow Cr. and YPP	Summer	17.3	16.3	15.6	5.8	15.9	16.3	15.9	15.2	14.7	$-1.0$	$-1.0$	$-2.1$	$-2.6$
(spawning, rearing, migration)	Fall	13.9	13.5	12.6	12.6	12.4	12.5	12.3	11.9	11.7	$-0.4$	$-1.4$	$-2.0$	$-2.2$
EFSFSR between YPP and Sugar Cr.	Summer	14.1	16.1	15.8	15.7	15.6	15.6	15.4	14.8	14.5	2.0	1.5	0.7	0.4
(migration, some rearing)	Fall	11.2	13.0	12.4	12.0	11.8	11.8	11.6	11.3	11.1	1.8	0.6	0.1	$-0.1$
EFSFSR downstream from Sugar Cr.	Summer	14.9	16.0	15.0	15.1	15.1	15.0	14.9	14.7	14.5	1.1	0.1	$-0.2$	$-0.4$
(spawning, rearing, migration)	Fall	11.9	12.5	11.6	.1.6	11.5	11.6	11.5	11.3	11.3	0.6	$-0.3$	$-0.6$	$-0.6$

**Table 54. Predicted maximum weekly maximum temperatures during summer (July) and fall (September) for modeled mine years.**

**Key:** °C = degrees Celsius; EFSFSR = East Fork South Fork Salmon River; YPP = Yellow Pine Pit

<sup>1</sup>Temperatures based on distance weighted average of all QUAL2K reaches.

<sup>2</sup>Temperatures based on distance weighted average of the QUAL2K reaches along the TSF and TSF buttress area. Anadromous salmonids will not inhabit these reaches. <sup>3</sup>Anadromous spawning, rearing, and migration will only occur downstream in reach MC4 (below the TSF).

*Construction.* The major activities at the mine site during construction that could impact stream temperatures include construction and activation of the EFSFSR flow diversion, diversion of Meadow Creek and its tributaries around the TSF and TSF buttress area; dewatering the YPP Lake (increased maximum temperatures); and stream enhancements. Modeling was not performed to examine stream temperatures in the years immediately following these activities. It is expected that some activities, such as routing the EFSFSR through the diversion tunnel, will reduce stream temperatures given the reduced exposure to solar radiation. Similarly, diversion of upper Meadow Creek around the TSF may result in some warming of surface flows during the early summer months, but installation of the low flow diversion pipe is expected to reduce stream temperatures during base flows. On the whole, stream temperatures are expected to be similar to baseline conditions and the ability of the habitats to support Chinook salmon and steelhead spawning, rearing, and migration will not be reduced as a result of construction activities.

*Operations and Closure/Post-Closure.* Stream temperatures for the various stream reaches will be impacted differently over the course of operations and closure/post-closure periods. Impacts to DCH are discussed by stream reach below.

*EFSFSR Upstream of Meadow Creek.* Baseline stream temperatures in the EFSFSR upstream from Meadow Creek tend to be cooler than the downstream reaches because: (1) it is a headwater reach; (2) it has minimal effects from historic mining; and (3) much of the of riparian vegetation is intact (though still recovering from historic fires). Water temperatures in this section of the EFSFSR under all phases of the SGP will be similar to those under baseline conditions because there will limited mine-related activities in riparian areas. As summarized in Table 54, stream temperatures in this reach of the EFMC are predicted to be lower than temperatures known to be supportive of migration, spawning, incubation, and rearing life stages.

The sanitary wastewater treatment facility will be a source of thermal loading to the stream. While the expected effluent temperature is unknown, the discharge makes up a very small portion of the receiving stream (1 to 3 percent), will be subject to an IPDES permit and CWA requirements, and is not expected to cause thermal shifts to a degree that would negatively impact fish. As such, stream temperatures are expected to continue to support migration, spawning, and rearing life stages of Chinook salmon. This stream could be used as cool water refugia.

*Meadow Creek Upstream from EFMC.* Meadow Creek upstream from the EFMC is predicted to have decreasing water temperatures relative to baseline conditions during mine operations (MYs 6 through 18) because baseflows will be piped around the TSF and will not be exposed to solar radiation. The predicted MWMT from MY6 through MY22 are predicted to be below thermal thresholds supportive of all salmonid life stages. Vegetation replanting on top of the TSF will take place in the five years following cover placement and stream restoration (i.e., MYs 23 through 27). When the pipeline is removed (MY 22), water temperatures will increase until around MY 27. Stream temperatures are predicted to peak in MY27 at 20.8°C. While these temperatures will not cause outright mortality, exposed fish could experience a myriad of sublethal effects. Absent localized cold water refugia that may be created by hyporheic flow additions to this stream reach, juvenile fish may move into adjacent, thermally-suitable habitat

(e.g., EFMC). Predicted summer MWMTs will remain at levels that could cause sublethal effects in rearing juveniles through at least MY52. Therefore, implementation of the proposed action will create a long-term negative effect to the temperature PBF for Chinook salmon. Predicted fall temperatures will generally be below 13°C, suggesting Chinook salmon spawning and incubation will not be negatively impacted most mine years. However, the predicted MWMTs for the fall period will exceed 13°C in this reach from MY27 through MY32 and potentially up to MY52. It is possible that some embryo mortality will occur during this time. Stream temperatures are expected to decrease in the years following MY 27 as replanted riparian vegetation becomes more established and provides more effective stream shading. This decrease is expected to continue through MY 112 as vegetation grows and shading increases. While habitat is expected to be continued to be used for Chinook salmon adult migration, spawning, incubation, and rearing, temperature conditions will not provide optimal conditions for adult migration in MYs 23 through 27, incubation of embryos during the early part of the spawning season in MYs 23 through 27, or juvenile rearing spawning/incubation in MYs 23 through 52.

*Meadow Creek Downstream from EFMC.* Water temperatures in the warmer summer and fall months in Meadow Creek downstream from EFMC substantially decrease relative to the baseline conditions during mine operations and closure/reclamation activities (MY 6 through MY 18), though there is an increase at MY 27 (post-closure), which then continues to decline until MY 112. These decreases during mine operations are a result of decreased solar radiation in upper Meadow Creek. The removal of the low-flow piping along the TSF in MY 22 will result in water temperatures increasing, though temperatures are predicted to remain lower than baseline conditions. Temperatures are predicted to subsequently decrease as vegetation grows and provides more shade. This section retains some connection to groundwater which will help maintain a lower temperature as well.

The predicted MWMT during MY 6 is 17.6°C, which could lead to decreased juvenile fitness. Predicted temperatures during MY 12 through MY 22 are slightly above 16°C, which may also lead to some small reductions in individual fitness. The predicted summer MWMT for MY 27 is 18.5°C, which is a temperature where reduced growth and increased disease risk may occur in rearing juvenile fish. Stream temperatures will remain elevated beyond MY 52, creating a longterm adverse impact to the temperature/water quality PBF in this reach. Similarly, predicted fall temperatures will exceed 13°C in MY 6 through MY 18. It is possible that some Chinook salmon embryo mortality will occur during this time.

Although the USFS indicated the MY 6-model run does not incorporate shade benefits (Stantec 2024), it is possible that MYs immediately following Meadow Creek channel relocation efforts in restoration reaches MC4, MC5, and MC6, may not be as cold as predicted. A similar channel reconstruction and riparian restoration effort took place in 2005 on Meadow Creek. Photos of this restoration reach are available for 2005, 2007, 2010, and 2015 (Arkle and Pilliod 2021; Zurstadt 2022). In 2007, grasses and some small shrubs were growing along the streambank. In 2010, the grasses and shrubs were denser along the streambanks, but were not large enough to provide much shade to the stream. By 2015, alder and willow had grown to sizes that afforded some stream shading. Arkle and Pilliod (2021) collected stream temperature data upstream and downstream of restored reaches on Meadow Creek. Maximum downstream temperatures were about 5°C greater than upstream temperatures in 2006 and 2007. In addition, no temperatures

greater than 16°C were recorded at the upstream monitoring location, but there were 78 and 171 occurrences of stream temperatures greater than 16°C at the downstream location in 2006 and 2007, respectively. While the monitoring locations were located about 1.4 miles apart and temperature data immediately upstream and downstream of the restored reach are not available, it is reasonable to infer that water traveling through the restored reach was exposed to greater solar radiation and as a result, stream temperatures warmed. It is possible for MY 4 and 5 (the years immediately following channel relocation), the 1.8°C reduction in summer MWMT will not be realized.

Predicted stream temperatures in lower Meadow Creek are expected to be less than 18°C in all modeled mine years, with the exception of MY 27, when stream temperatures are anticipated to be slightly elevated. These slightly elevated temperatures are likely to persist up to MY 32, which is when predicted summer MWMT will fall below 18<sup>o</sup>C. Adult migration is expected to continue in this reach during these five years; however, individuals may experience sublethal effects. Conditions for Chinook salmon spawning and incubation are expected to improve, but will continue to be elevated above 13<sup>o</sup>C from MY 6 through MY 18 and also in MY 27 through MY 32. During these time periods, some Chinook salmon embryo mortality may occur. Similarly, conditions for Chinook salmon juvenile rearing will improve, but will remain slightly elevated above 16°C through MY 52. Temperatures are not predicted to reach lethal levels; however, the predicted warm temperatures are expected to cause some individual fish to experience sublethal effects. Considering the elevated stream temperatures that are likely to persist in this reach, the ability of the temperature PBF to support Chinook salmon spawning, incubation, adult migration, and juvenile rearing will be reduced in this reach.

*East Fork Meadow Creek.* EFMC experiences around a 1°C increase in summer and fall maximum water temperatures during mine operations and into the post-closure phase (starting at MY 6) through sometime before MY 52 (post-closure phase), at which point the temperatures decline compared to the baseline conditions. Predicted summer MWMTs for all model years are less than 16°C, indicating juveniles rearing in the reach will not experience adverse effects from elevated temperatures. The temperature increases in EFMC during the operations phase are due to the removal of vegetation as part of the channel restoration work. Restoration activities on the EFMC are scheduled to begin in MY 1, with the construction of the rock drain starting in MY 3. Once the restoration activities (including the establishment of sufficient riparian vegetation to provide stream shading) are fully completed, water temperatures are expected to begin to decrease. Water temperatures are predicted to drop below baseline conditions by MY 52. By MY 112, the reduction in water temperature between the upper meadow and the lower EFMC is around 0.3 and 0.4°C for both the summer and fall maximums. Stream temperatures in lower EFMC will fully support juvenile rearing in all years, which is the only Chinook salmon life stage anticipated in this reach.

*EFSFSR Between Meadow Creek and the YPP.* The EFSFSR between Meadow Creek and YPP experiences decreases in summer MWMTs relative to baseline conditions during the operations period. There is a minor increase in temperatures after MY 22 once the Meadow Creek low-flow piping along the TSF is removed; however, temperatures are predicted to remain lower than baseline conditions. Temperatures continue to decrease once the riparian vegetation in upstream reaches grows to a size that will provide stream shade. Stream enhancements made to the reach

above the YPP are predicted to lower temperatures relative to baseline conditions. This lowering is expected to be greater than the increase associated with warmer water entering the EFSFSR from Meadow Creek in the closure period. Fall maximum water temperature is predicted to decrease throughout the operations, closure, and post-closure periods. While the SGP is expected to improve stream temperatures in this reach of the EFSFSR relative to baseline conditions, temperatures will remain above levels needed to fully support Chinook salmon juvenile rearing, particularly in MYs 6 and 27; and maximum fall temperatures will be slightly above levels known to fully support Chinook salmon spawning and incubation at least for the first few weeks of the season in MY 6. Stream temperatures will fully support adult Chinook salmon migration and holding.

Predicted maximum summer stream temperatures in this reach are below 16.5°C, therefore, the ability of the temperature PBF to support adult migration will not be reduced. The ability of the temperature PBF to support juvenile Chinook salmon rearing will be slightly reduced in MY 6 and MY27 because predicted stream temperatures are 16.3°C and juveniles rearing in these temperatures for extended periods of time may experience sublethal effects. Predicted maximum fall stream temperatures will be below  $13^{\circ}$ C, with the exception of MY 6 (predicted temperature of 13.5°C). This slightly elevated fall stream temperature will slightly reduce the ability of the temperature PBF to support spawning and incubation during the early portion of the spawning period (stream temperatures can decline relatively quickly in the fall, as shown in Figure 26).

*EFSFSR Between YPP and Sugar Creek.* Summer and fall maximum temperatures in the EFSFSR between YPP and Sugar Creek are predicted to increase during MY 6. This increase is caused primarily by the draining of the YPP lake followed by active mining and mine dewatering that removes cooling influences of upstream shading and groundwater discharge to surface water. By MY 12, water temperatures are expected to drop, as the YPP is backfilled, the EFSFSR stream channel is restored, and Stibnite Lake is created. The Stibnite Lake is expected to reduce maximum water temperatures in the EFSFSR downstream of the lake outlet, similar to the existing YPP lake. As riparian vegetation establishes along the EFSFSR riparian corridor and begins to provide stream shade, water temperatures will continue to drop. By MY 112, predicted summer maximum water temperatures in the EFSFSR between YPP and Sugar Creek are about 0.4 $\degree$ C higher than baseline conditions, but fall maximum temperatures end up 0.1 $\degree$ C below baseline conditions. Stream temperatures in this reach of the EFSFSR are predicted to be less than 16°C during the summer for all modeled years except MY 6 and temperatures will be equal to or less than 13°C during the fall for all modeled years. Based on these model results, the temperature PBF will be able to support migration, spawning, and incubation for Chinook salmon. During MY 6, the maximum predicted summer temperature is 16.1°C. This small increase in temperature above 16°C will slightly decrease the ability of the water temperature PBF to support salmon and juvenile rearing.

*EFSFSR Below Sugar Creek.* The EFSFSR, roughly 0.62 miles (1 km) downstream from Sugar Creek, is influenced by both the changes in temperature in the EFSFSR upstream (described above), as well as inflow from Sugar Creek. The substantial increase in temperatures in the early operations phase is caused primarily by the YPP dewatering and mining. By MY 12, temperatures are expected to drop below baseline conditions due to the restoration of the EFSFSR and creation of the Stibnite Lake. After this point in time, temperature changes will be

minimal (due to the influence of Sugar Creek) and are expected to remain below baseline conditions. The predicted summer stream temperatures for all modeled years will be equal to or less than 16.0°C. Considering this, the temperature PBF will continue to support juvenile salmon and steelhead rearing and adult migration (Chinook are the only species where adults are likely to be present in this reach during the summer months). Predicted fall tempeaturs will remain below 13°C, therefore the temperature PBF will be able to continue to support spawning and incubation.

*Uncertainty*. There are a number of inherent sources of uncertainty when interpreting results from the SPLNT temperature model. The most important of described below.

*Riparian Vegetation.* The SPLNT model assumed that vegetation will not grow to their full potential height, but rather, vegetation would only reach the midpoint of their growth potential range (Brown and Caldwell 2021e). Incorporation of this assumption into the model accounts for some uncertainty regarding the effectiveness, timing, and sustainability of the shading effects of riparian plantings. Even so, performance standards such as percent cover, and measured stream temperatures will be used to assess restoration and reclamation success and to determine the need for adaptive management.

*Stibnite Lake.* The assumed influence of the lined Stibnite Lake feature on stream temperatures and whether the simulated surface water temperature reductions will be realized. Lining the lake feature atop the lined and covered backfill in the YPP will modify the volume of diffuse subsurface groundwater inflow. While the lined Stibnite Lake feature will receive minor inflow from the cover material, it will not exactly mimic the existing groundwater inflow from native bedrock into the YPP Lake. Depending on the hydraulic properties of the cover material compared to the native bedrock, the volume of groundwater inflow to the lake could differ from existing inflow rates. To address this uncertainty in a conservative manner, the current temperature model did not incorporate any potential cooling effects from subsurface inflow into the Stibnite Lake feature.

*Groundwater Levels.* The reduction and recovery of groundwater levels and groundwater discharge to surface water varies spatially and temporally. This difficult-to-predict spatial and temporal variability contributes uncertainty in the stream temperature predictions.

*Wastewater Treatment Plant Discharge to Meadow Creek.* The SPLNT model does not include thermal contributions from the wastewater treatment plant discharge to Meadow Creek. During colder months (October through April), the temperature of treated water is estimated to be  $7.3^{\circ}$ C (Brown and Caldwell 2021b). During the warmer months, retention times for contact water in ponds will be up to 34 days, resulting in warmer water treatment plant influent. The wastewater treatment plant discharge will have the greatest influence during the operational period, specifically MYs 4 through 6. During this period, the effluent will comprise seven and 55 percent of the flow in Meadow Creek. The USFS predicts the discharge will increase stream temperature in Meadow Creek by 1 to 3°C below the outfall during the summer and fall. However, warmer water treatment plant discharge temperatures will be offset by the cooling effect of the piped diversion of Meadow Creek around the TSF with the net effect of water treatment on temperature of Meadow Creek expected to be less than  $0.25^{\circ}$ C (Brown and

Caldwell 2021b). This additional source of thermal loading, if limited to be less than  $0.25\text{ °C}$ , is not expected to preclude the habitat from supporting adult migration; however, it will further reduce the ability of Meadow Creek to support spawning/incubation (likely restricted to the first few weeks of spawning) and juvenile rearing.

*Climate Change Considerations.* The SPLNT model does not account for future climate change effects on air temperature, meteoric precipitation, weather events, wildfire, and plant growth. This means that modeled future water temperatures assumed that without the SGP, stream temperatures will be similar to the historic water temperature data (Brown and Caldwell 2018). In reality, water temperatures will likely be higher if climate change had been incorporated into the model. Climate change will be expected to increase water temperatures from baseline estimates to the end of the mine operations by as much as 0.1°C to 2.0°C based on forecasts for 2030-2059 (Isaak et al. 2016), depending on the stream reach. Into the future, baseline estimates for water temperatures could increase by as much as an additional degree (2070-2099).

Given ongoing climate change, model results likely underpredict future temperatures. The extent to which the model may have underpredicted stream temperatures is difficult to quantify because not only does climate change modeling have inherent uncertainties, there are uncertainties associated with climate change impacts on stream restoration and revegetation success. The effects of different air temperature conditions on stream temperatures were evaluated through a sensitivity analysis (Brown and Caldwell 2018) and an uncertainty analysis (USFS 2023a). Brown and Caldwell (2018) evaluated the sensitivity of the model to altered air temperatures by increasing and decreasing model inputs by 5°C. Across this 10°C variation, simulated water temperatures varied by up to approximately 1°C (Brown and Caldwell 2018). Stream water temperature modeling uncertainty relates largely to spatially and temporally variable implementation success of closure activities. This uncertainty, combined with broader climate conditions, could result in higher than predicted stream temperatures (USFS 2023a).

Ultimately, climate change with or without implementation of the SGP will make it more difficult to maintain temperatures that are protective of all anadromous salmonid life stages within the action area. Climate change is expected to exacerbate the effects of the proposed action on stream temperatures directly as well as indirectly through impacts to air temperature, streamflow, and vegetative recovery.

The above sources of uncertainty relate largely to spatially and temporally variable implementation success and sustainability of closure activities, further complicating the simulation of future stream temperatures. Qualitatively, insufficient and/or ineffective closure activities and/or adverse changes in broader climate conditions may result in higher than predicted stream temperatures. Because of this uncertainty, the USFS is requiring additional monitoring and mitigation measures for the SGP including increased riparian planting, mechanical shading, and snow harvesting during the closure/post-closure period.

# *2.10.1.1.7. Turbidity.*

Water quality could be affected through project generated turbidity, which could occur due the following ground disturbing activities: (1) BRGI site investigations; (2) road building, use, and maintenance; (3) facility and infrastructure construction; (4) mining and mining exploration (i.e.,
operations); or (5) through road decommissioning and site/stream restoration. Ground disturbing activities will generate sediment and deliver it to action area streams to some degree from BRGI investigations (MY -3), from the onset of construction (MY -3, to -1), through closure and reclamation (MY 16+).

*Pre-Construction.* Travel on unpaved roads, test pit investigations, and drilling associated with the BRGI portion of the project could result in fine sediment reaching streams. Because travel on unpaved roads will be via maintained USFS roads, will mostly be via light vehicles, and heavy vehicles will be routed to minimize adverse effects on roads adjacent to streams, the fine sediment effects of travel on unpaved roads should be minimal. Because proven erosion control practices will be implemented while test pits are active, because test pits will be appropriately reclaimed, because test pits will in relatively flat areas, and because most test pits are on the upslope side of FR 447, the sediment-related effects will be effectively minimized, and the effect on the water quality PBF from the BRGI portion of the project will be minor.

During construction, turbidity is expected from all ground-disturbing activities associated with developing the site (e.g., channel restoration, road work, powerline and facility construction, etc.). During operation and closure, turbidity can be expected from mine road use/maintenance, surface exploration, road decommissioning, channel restoration, etc.

The most critical aspects of a sediment effects analysis are timing, duration, intensity, and frequency of exposure (Bash et al. 2001). Depending on the level of these parameters, turbidity can cause lethal, sublethal, and behavioral effects in juvenile and adult salmonids (Newcombe and Jensen 1996). For salmonids, turbidity has been linked to a number of behavioral and physiological responses (i.e., gill flaring, coughing, avoidance, and increase in blood sugar levels) which indicate some level of stress (Bisson and Bilby 1982; Berg and Northcote 1985; Servizi and Martens 1987). The magnitude of these stress responses is generally higher when turbidity is increased and particle size decreased (Bisson and Bilby 1982; Servizi and Martens 1987; Gregory and Northcote 1993). Although turbidity may cause stress, it has been shown that moderate levels of turbidity (35 to 150 NTUs) accelerate foraging rates among juvenile Chinook salmon, likely because of reduced vulnerability to predators (camouflaging effect). Turbidity and fine sediments can reduce prey detection, alter trophic levels, reduce substrate oxygen, smother redds, and damage gills, among other deleterious effects (Bjornn and Resier 1991; Spence et al. 1996).

For our assessment, we considered sediment delivery in the temporary (0-3 years), short-term (3- 15 years), and long-term (greater than 15 years) timeframes. Sediment generated from construction-related activities (i.e., in the temporary timeframe) is addressed qualitatively. Sediment modeling conducted by the Tetra Tech (2024) informed our assessment of the impact of the project on short- and long-term sediment delivery to streams from roads within the action area relative to baseline conditions. While the GRAIP-Lite model used by the Tetra Tech represents some of the best available tools for evaluating potential project impacts, there are some limitations. Model outputs are influenced by the assumptions made about: (1) storm damage risk reduction (SDRR) treatment type, extent, and effectiveness; (2) erodibility of road segments based on maintenance level and other factors; and (3) probability of sediment delivery to nearby streams. Considering their limitations, NMFS views these modeling results as one line

of evidence for potential effects of the action. NMFS also relied heavily on the assumption that EDFs and BMPs will be properly implemented in our overall assessment of the degree and extent to which, adverse effects may occur.

*Construction.* Construction of a road network can greatly accelerate erosion rates in a watershed (Haupt 1959; Swanson and Dyrness 1975; Swanston and Swanson 1976; Beschta 1978; Gardner 1979; Cederholm and Reid 1987). Sediment generated through road construction and reconstruction can reach streams through surface erosion and mass movements of destabilized soil, the effects of which can be both dramatic and long-lasting (Meehan 1991). Unpaved road surfaces continually erode fine sediments, adding significant amounts of sediment to streams (Reid and Dunne 1984; Swanston 1991). Roads and related ditch networks are often connected to streams, providing a direct conduit for sediment. On steep slopes, road construction or improper maintenance can greatly increase landslide rates relative to undisturbed forest (Swanson and Dyrness 1975; Swanston and Swanson 1976; Furniss et al. 1991), delivering large amounts of sediment to streams.

Increases in sediment supply beyond the transport capability of the stream can cause channel instability, aggradation (sometimes to the extent that perennial streams become intermittent) (Cederholm and Reid 1987), widening, loss of pools, and a reduction in gravel quality (Sullivan et al. 1987; Furniss et al.1991; Swanston 1991). For salmonids, these changes can mean reduced spawning and rearing success when spawning areas are covered, eggs and fry suffocate or are trapped in redds, food abundance is reduced, and over-wintering habitat is reduced (Cederholm and Reid 1987; Hicks et al. 1991).

Roads built in riparian areas often eliminate part of the riparian vegetation (Furniss et al. 1991), reducing LWD recruitment and shade. Riparian roads also constrain the natural migration of the stream channel where channel migration zones are present. Roads can intercept, divert, and concentrate surface and subsurface water flows, thereby increasing the watershed's drainage network (Hauge et al. 1979; Furniss et al. 1991; Wemple et al. 1996). This can change peak and base stream flows and increase landslide rates. Stream crossings can restrict channel geometry and prevent or interfere with migration of adult and juvenile anadromous fish (Furniss et al. 1991). Culverts also can be a source of sedimentation, especially if they fail or become plugged with debris (Furniss et al. 1991; Murphy 1995).

Surface erosion from forest roads affects the fine sediment budget in streams and may impose a chronic condition of sediment inputs that directly affect the stream substrate and the health of aquatic life (Luce et al. 2001). Road work results in increased runoff rates and fine sediment delivery, particularly during rain events at tributary crossings or where roads are in close proximity to streams. Road surfaces are compacted, and impermeable surfaces lack vegetation. As a result, most water falling on the road surface in the form of rain or delivered there from upslope does not infiltrate, but rather, quickly runs off the road surface. Water flowing on the road surface picks up fine sediment and delivers it to nearby streams, resulting in increased turbidity and fine sediment deposition onto substrate. Widening of the roads exposes new soil and prevents vegetation from becoming established on a larger portion of the watershed. Generally, a vegetative strip between the road and the stream allows runoff to infiltrate or be filtered prior to entering streams. However, in many locations within the action area, vegetation between the road and action area streams is narrow to non-existent, and runoff is quickly delivered directly into waterways (particularly along the EFSFSR and Johnson Creek).

Planned activities such as placement of cross drains, ditching, grading and graveling may result in disturbances that typically create short-term increases in sediment delivery that taper off after disturbed areas become compacted or after several runoff events occur. To facilitate new road construction and road widening during the construction phase, existing vegetation will be cleared and trees will be felled. Land clearing and tree cutting will create a temporary disturbance to the ground and will expose more soil to potential erosion by eliminating the soil binding capabilities of the roots. This will result in sediment delivery to streams, with the greatest impacts likely to take place along the Burntlog route in the short term  $(0 \text{ to } 3 \text{ years}^{11})$  as disturbed cut and fill slopes become revegetated.

Although exposed cut and fill slopes are likely to increase sediment delivery in the temporary and short-term timeframes, erosion control and road maintenance EDFs should limit the amount of sediment delivered to action areas in the long term. Maintaining roads in good condition decreases chronic delivery of fine sediment to streams, as roads in close proximity to streams can convey large amounts of fine sediments (Furniss et al. 1991). As proposed, surfacing these road segments and their ditches with gravel aggregate, insloping the road prism to drain to the ditchline, and treating the gravel surface with dust abatement chemicals, should result in a localized reduction on overall sediment delivery to the action area streams over the long term. Although not likely to be of a magnitude sufficient to benefit conditions at the stream reach or watershed scales, these localized improvements could locally reduce factors which limit reproductive success and juvenile survival, abundance, and productivity.

In the temporary and short-term timeframes, construction equipment and roadwork are expected to result in increased sediment delivery to action area streams. Sediment-related effects are most likely to be realized to Chinook salmon habitat along the EFSFSR, and Johnson, Meadow, Sugar, and lower Burntlog Creeks; and steelhead habitat along the EFSFSR, and Johnson, Sugar, Burntlog, Trapper, and Riordan Creeks.

The level of turbidity is likely to exceed ambient levels and potentially affect critical habitat downstream of project activities. Ground-disturbing activities will likely deliver a series of temporary sediment pulses or plumes (expected duration of minutes to hours). However, quantifying turbidity levels and their effect on fish species and their habitat is complicated by several factors. First, turbidity from an activity will typically decrease as distance from the activity increases. The time needed to attenuate these levels depends on the quantity of material in suspension (e.g., mass or volume), particle size, the amount and velocity of ambient water (dilution factor), and the physical/chemical properties of the sediments. Second, the impact of turbidity on fish is not only related to the turbidity levels but also to the particle size of the suspended sediments. Whether these increases in sediment production delivers to nearby streams depends on a variety of factors including the proximity of the road to a stream (Nelson et al. 2012), construction methods employed, and effectiveness of erosion control BMPs. Instream work will also lead to increased sediment delivery during and immediately following

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 $11$  In this analysis, we use definitions for duration of effects as defined in the Southwest Idaho Forest Plan and used in its Matrix (i.e., temporary =  $<$ 3 years, short term = 3 to 15 years, and long term > 15 years.

construction. Fording of streams can also cause temporary spike in turbidity and subsequent sediment deposition.

The project will initiate with three years of construction (MY -3, to -1). During the initial 2 years of project construction, mine-related traffic will access the mine site via the Johnson Creek route, much of which is located in close proximity to streams (i.e., within 100 feet of Johnson Creek and the EFSFSR). A total of 65 vehicle trips per day will occur along Johnson Creek and the EFSFSR during the construction phase, consisting of 20 light vehicles and 45 heavy vehicles. There will be no road alignment modification or widening of the road prism along this route in this phase, as the current road is able to accommodate the equipment and materials needed to be transported during the construction period. However, minor surface improvements (e.g., ditch and culvert repair, adding gravel, winter snow removal, resurfacing (i.e., gravel addition) if required, and summer dust suppression, will occur on the route to reduce sediment runoff and dust generation during this phase.

Under baseline conditions, turbidity is reportedly low (less than 5 NTU) with occasional spikes of up to 70 NTU during snowmelt or rainfall events (USFS 2023b). The greatest potential for Project-related increases in sediment delivery will come during storm events causing overland flow across exposed soil, excavated areas, and roads. Implementation of EDFs and BMPs are expected to minimize the amount of sediment delivered to nearby streams. However, temporary spikes in turbidity are still expected to occur during or following instream or ground-disturbing activities. Turbidity increases will occur immediately adjacent to and downstream of activities and will dissipate as suspended materials settle to the channel bottom.

The magnitude, duration, and extent of turbidity pulses is dependent upon the type and extent of work being performed along with the EDFs implemented. Based on observations from full-size vehicle fording, spikes in turbidity are expected to dissipate quickly and have relatively small magnitudes. Dewatering will also occur at all culvert removals or installation locations, and all reaches will be rewatered slowly to minimize turbidity. Instream work should be completed within a few hours (i.e., maintenance or removal of a culvert) to a few weeks (i.e., the more extensive work required for stream restoration, bridges, and larger culvert replacements).

Turbidity plumes associated with instream work (e.g., excavating culverts, re-watering, etc.) are anticipated to travel up to 1,000 feet downstream prior to dissipating to levels that are no longer harmful to aquatic species. This assumption is based on turbidity monitoring reports from past projects in the region which involved reconstruction of stream channels, including culvert, bridge, and diversion replacement projects (Eisenbarth 2013; Connor 2014). In many cases turbidity plumes upon rewatering will last less than 2 hours, but the plumes may last for up to 24 hours (Connor 2014; Jakober 2002; Casselli et al. 2000; Eisenbarth 2013). Similar turbidity monitoring results have been reported for rewatering reconstructed side channels (CH2MHill 2012). However, based on review of turbidity monitoring reports for habitat restoration projects completed between 2015 and 2017, turbidity plumes in excess of 50 NTUs were usually of short duration of 15 to 30 minutes, but sometimes lasted up to just under 6 hours. Giving the proposed conservation measures to address turbidity, turbidity plumes are rarely expected to reach levels where they mix across entire stream channels, more typically tending to hug one streambank.

In an effort to better understand how much sediment will be produced, delivered, and accumulated in receiving streams over time, Tetra Tech conducted GRAIP Lite modeling (Tetra Tech 2024) for roads in the action area. GRAIP Lite modeling (Tetra Tech 2024) simulated three scenarios, existing conditions, construction conditions, and operational conditions. Results estimated a 32 percent decrease in sediment production during the construction period, with a 30 percent decrease to sediment delivery, and 31 percent decrease in sediment accumulation as compared to baseline conditions (Table 55). Table 56 breaks down sediment delivery by structure type and watershed.

BMPs will be employed for near-stream or instream work, such as removal of legacy materials and stream restoration, to minimize the potential for coarser sediment generation or mass wasting that will affect sediment transport and deposition. Under baseline conditions, sediment entering the EFSFSR primarily comes from Sugar Creek, Meadow Creek, and EFMC (USFS 2023b). Applicable sediment control BMPs will be used to minimize sediment runoff and erosion along roads and excavated areas. On the mine site and along the Burntlog Route, EDFs will protect streambank vegetation, require culvert maintenance, and require low impact snow removal techniques. Efforts to resurface the road, improve drainage, and maintain project-related roads are expected to reduce effects from access roads over their existing condition.

Baseline conditions are generally FA for turbidity in action area streams (Table 30). Proposed sediment control measures and road maintenance are expected to effectively limit the amount of sediment delivered from road-related activities. NMFS believes that the road upgrades proposed and the conservation measures described should be adequate to keep turbidity and erosion to levels to a minimum. Although sediment delivery will occur in reaches bordering roadwork, GRAIP-Lite modeling suggests the amount is not likely to appreciably diminish the conservation value of DCH at the 5<sup>th</sup> field HUC level.





### **Table 56. Annual Sediment Delivery (kg/year) to Drainage Crossings under Baseline and Construction Conditions (Stantec 2024).**



New and existing roads also pose an increased risk of landslide; however, the GRAIP-Lite model used to estimate sediment delivery does not account for landslides (Dixon 2019). Many of the subwatersheds in the action area have a high inherent risk of landslides, which are exacerbated by the presence of roads (Dixon 2019). Although studies on the effect of road obliteration on landslide risk are lacking, it is reasonable to assume that road decommissioning with full obliteration proposed for new road segments at the end of operation will reduce the long-term risk of road-related landslides that could contribute sediment to streams because those roads will no longer intercept and route water.

As previously mentioned, turbidity will also be created during the construction and retrofitting of stream crossings, the bulk of which will take place during construction of the Burntlog Route. Erosion and sediment control for this in-water work will be consistent with BMPs being used for other aspects of the SGP. To minimize turbidity effects to the water quality PBF, cofferdams will be installed, streams bypassed, and in-channel work will be completed in the dry. Water will be slowly reintroduced to minimize turbidity levels, and turbidity monitoring will be conducted. Should turbidity exceedances occur, work will stop to address the turbidity issues.

In addition to the transportation network, additional infrastructure will also need to be constructed during the site preparation phase (MY -3 to -1). This will include the construction of surface facilities, water management features, power transmission lines and substations, OSV routes, and channel diversions, all of which have the potential to cause ground disturbance and sediment delivery to action area streams.

Ground disturbance will occur during the construction of the worker housing facility, main gate, administrative office, and Burntlog maintenance facility. As facilities are being developed during pre-construction, water management BMPs will be constructed to reduce erosion and sediment delivery to streams, including sedimentation ponds; run-on water diversion ditches, trenches, and/or berms; runoff water collection ditches; silt fences; water bars; culverts; energy dissipation structures; terraces, etc. Although construction of these facilities will generate sediment, the predisturbed nature of the majority of these construction sites, the relatively flat nature of the construction sites, and the application of proven erosion control practices, are expected to combine to effectively minimize the amount of sediment delivered to action area streams during the pre-construction phase of the project.

A new, temporary 16-foot-wide groomed OSV trail will be created adjacent to Johnson Creek Road between the proposed Cabin Creek Groomed OSV Route (Section 1.6.4 and Figure 7). There will also be a 2-acre parking area constructed west of FR 467, and a new 1.9-mile groomed access trail from the USFS Warm Lake Project Camp on Paradise Valley Road (FR 488). Another 16-foot-wide groomed OSV trail will be created and maintained north of Warm Lake Road to connect the southern end of the Cabin Creek Road OSV trail to the Warm Lake Road (FR 579), providing access to North Shoreline Drive (FR 489) from the Cabin Creek Road OSV trail. This 0.3-mile route will be used throughout construction and operations and will require the removal of some vegetation and trees.

The new segment of OSV trail near Johnson Creek is on the upstream side of the Johnson Creek Road and not expected to deliver measurable quantities of sediment to Johnson Creek before it

becomes stabilized. The remaining OSV trails will be constructed largely along existing roads. Although some vegetation and tree removal will need to occur to ensure safe snowplowing, these OSV trails are not expected to produce meaningful sediment delivery and turbidity beyond that already generated by the existing road network.

Work associated with power transmission lines will also clear vegetation and disturb soils, potentially contributing to sediment delivery to action area streams. During the construction phase of the project, access roads for powerlines, communicator towers, and repeater sites will be constructed, and vegetation clearing will be necessary at tower locations and along the powerline ROW. The powerline will cross 37 different streams, 26 of which are related to the upgrade of existing IPC infrastructure where the existing transmission line ROW crosses various streams. The existing transmission line currently crosses Cabin Creek, Trout Creek, and Riordan Creek. The new transmission line will cross three creeks with only one being perennial (Riordan Creek).

Riordan Creek supports *O. mykiss*, and although the BA considered these fish resident rainbow trout versus steelhead, steelhead are present in Johnson Creek, and NMFS is not aware of any passage barrier in Riordan Creek that would preclude steelhead access to this stream. Therefore, in this analysis, NMFS considers *O. mykiss* in Riordan Creek steelhead upstream to approximately Riordan Lake.

During transmission line upgrades and new transmission line construction, the potential exists for increased runoff, erosion, and sedimentation as a result of vegetation removal within the ROW, and the localized excavation of soil, rock, and sediment for structure work and/or ROW access roads. Although often requiring a wider clearing zone, vegetation clearing along the ROW is not expected to generate large amounts of sediment as work will be limited to trimming of trees that pose a fire risk to the power line. During all vegetation clearing activities, IPC will ensure there is no disturbance of the soil surface that will create an added risk of erosion, the promotion of the establishment or expansion of invasive species (including noxious weeds), damage to cultural resources, sensitive species, or ESA-listed species. Vegetation clearing will retain vegetation root structure within soils, therefore reducing erosion concerns.

Access routes will be required from the existing access road to reach powerline structure locations without current access (Table 11). Roads will be opened/cleared for use by trucks transporting materials, excavators, drill rigs, bucket trucks, pickup trucks, and crew-haul vehicles. These overland service routes will require a 14-foot-wide ROW to accommodate construction and maintenance equipment. Access roads will generally be left as close to an undeveloped nature (i.e., two-track road) as possible without creating environmental degradation (e.g., erosion or rutting from poor water drainage). Specific actions, such as installing water bars and dips to control erosion and stormwater, will be implemented to reduce construction impacts and will follow standard designs. Routine inspection and maintenance of service and access roads, such as blading the road to maintain the surface condition and drainage, removing minor physical barriers (i.e., rocks and debris), replacing culverts or rock crossings, and rehabilitating after major disturbances requiring heavy equipment (such as slumping). Heavy equipment will travel and maneuver on existing service and access roads. Given proposed road maintenance and application of erosion control EDFs will be similar to those used for other SGP access roads,

NMFS therefore expects that powerline access roads will deliver some quantity of sediment to action area streams, but they are not likely to become a chronic source of sediment delivery, and are not expected to meaningfully impact the water quality PBF.

The new powerline ROW overlaps with 14.8 acres of RCAs. Overall, the transmission line ROW overlaps with 132.4 acres of RCAs (including non-anadromous RCAs) (Stantec 2024). Riparian vegetation removal as a result of powerline construction has the potential to decrease bank stability, which may result in bank erosion and an increase in sediment to the waterways. For the new portion of the powerline, vegetation will need to be cleared for structure footings. However, the utility poles are not directly along creeks or within RCAs, so clearing of construction pads for footings is not likely lead to sediment deliver and turbidity impacts to the water quality PBF. Instead, vegetation removal in RCAs, if it occurs, will target taller vegetation that exceeds the maximum height specified for the wire and border zones. Because lower-growing vegetation will be maintained along the ROW, bank stability is not expected to be compromised to a degree that will result in high levels of sediment delivery.

Overall, for the powerline transmission line portion of the action, turbidity plumes are expected to be limited to storm events following construction activities due to the distance project activities occur from streams. In addition, implementation of various erosion control BMPs is expected to limit not only the amount of sediment produced, but also subsequent delivery to nearby streams. As such, turbidity pulses are expected to be low-intensity, localized, infrequent, and short-lived. Therefore, the conservation value of the water quality PBF in the action area is not expected to be appreciably diminished from transmission line activities.

As described in the BA (Stantec 2024), surface water quality also could be impacted during construction by fugitive dust from vehicles and heavy equipment that settles into adjacent waterbodies. Perpetua plans to address these potential impacts through dust abatement measures on the roads and at the SGP. In dry months, Perpetua will spray water and dust abatement chemicals on mine haul roads as necessary to mitigate dust emissions in compliance with state and USFS requirements.

In addition to runoff from new ground disturbance, diversions of stream channels during the construction period have the potential to introduce turbidity as water is diverted into new channels. A 0.9-mile-long tunnel will be built in MY -1 to direct the EFSFSR around the west side of YPP to allow mining in the pit and fish passage during construction and operations (Figure 13). Fiddle, Hennessy, and Midnight Creeks will also be diverted in MY -1. Meadow and Garnet Creeks will be diverted in MY -2, while segments of Garnet Creek and EFMC will also be reconstructed during this same period. Diversion tunnels will be completed in the dry, and EDFs require pre-rinsing of diversion channels and introducing flows slowly will effectively limit the amount of turbidity caused by these actions (i.e., limited to short, low magnitude pluses, extending less than 1,000 ft. downstream).

New surface and underground exploration activities will be conducted beginning in the construction period and continuing through operations. Although underground exploration is unlikely to deliver sediment to action area streams, surface exploration may. Surface exploration is estimated to occur on approximately 65 acres of disturbance within the operations boundary

(Figure 15), including approximately 25 acres of temporary roads and 40 acres of drill pads. Other than at 11 locations, the exact locations of the drilling pads have not yet been determined, with general areas expected to be targeted displayed in Figure 16. The locations will be located in, but not necessarily limited to, areas around the WEP, YPP, Fiddle Creek GMS, and HFP. As outlined in the proposed action, foreseeable drill areas are offset from flowing streams with no exploration areas along Sugar Creek.

Perpetua proposes using the same or similar drilling methods and environmental protection measures that have been employed for exploration drilling in the past (i.e., exploration under the Golden Meadows Exploration Plan). However, the BA did not provide any specificity regarding this approach. This considered, NMFS includes a list of project design features and BMPs in Appendix B that we deem most pertinent to protection of ESA-listed fish species and critical habitat in the ESA Section 7 consultation completed on that project (NMFS No.: WCR-2015- 3169). Our analysis assumes that all of these project design features and BMPs will be followed for exploration activities.

Drilling support equipment will include helicopters, water trucks, crew trucks, portable mud tanks, pipe trucks or skids, portable toilets, light plants, portable generators, motor graders, excavators, and dozers. In terms of ground disturbance, the proposed action will utilize a rolling maximum of five acres of active temporary road disturbance (10,500 liner ft. of road) and eight acres of active drill pad disturbance (140 pads) within the total 65-acre authorization and will reclaim road and pad disturbance to remain below that rolling maximum of active disturbance. Where practicable, Perpetua plans to establish drill pads in reclaimed roadbeds and only open temporary roads in the vicinity of authorized mine disturbance in order to access exploration targets. Therefore, NMFS presumes some degree of minor brush clearing and minimal tree cutting may be required to clear areas for each platform. Drill pads created for truck-mounted or crawler-mounted drill rigs will require a 75 to 100-foot-long by 50 to 60-foot-wide working area (less than 0.15 acres), while drill pads supported by helicopter require working areas approximately 45 feet long by 35 feet wide working areas (less than 0.05 acres).

Drilling activities and road preparation/use will potentially cause increased sediment delivery to the EFSFSR and its tributaries, causing temporary turbidity pulses and subsequent sediment deposition. No excavation is anticipated for construction of the pads; however, excavation will be required for sump construction. The SGP will implement a variety of erosion and sediment control EDFs to minimize the amount of sediment mobilized and transported to streams. Sediment basins and traps will be used at each drill site to collect drill cuttings and to manage and circulate drilling fluids.

As in the Golden Meadows Exploration project (NMFS No.: WCR-2015-3169) and summarized in Appendix B, the risk of sediment delivery to streams should be minimized by ensuring all drill locations are at least 100 feet away from the nearest surface waterbody. And, should a drill pad be needed in an RCA, NMFS expects Perpetua will submit a written request to the PNF for approval of the location. The request must include an explanation as to why there is no reasonable alternative to siting the pad in an RCA. MGI must receive approval from the PNF prior to pad construction. Drill pads in RCAs will be sited to avoid removing any large trees to the extent possible. Any trees that are felled within the RCA will be left in the RCA. Due to

drilling pad locations, it is likely that much of the sediment not trapped by erosion control PDFs will be filtered by vegetation and other obstructions on the ground.

Reclamation of exploration disturbance will be conducted as soon as practicable following data collection and at least three growing seasons is estimated to be needed to establish vegetation and determine reclamation success. Areas disturbed for exploration will be contoured to blend into surrounding terrain; water bars and surface water channels will be retained to handle flows through the area. Compacted areas will be de-compacted as necessary prior to fertilizing and seeding.

Considering project design features and erosion control EDFs associated with exploration, these activities are unlikely to deliver any more than low-intensity, localized, infrequent, and shortlived turbidity pulses and plumes, minimally affecting the conservation value of the water quality PBF.

*Operations.* During operations, active mining will disturb, excavate, and move soil and overburden, thereby raising the potential for sediment runoff and suspended sediment increases in surface waters. However, contact water controls will be in place to capture runoff from mine facilities. During this phase (approximately 15 years), a total of 50 vehicle trips per day are anticipated on average (year-round) during operations utilizing the Burntlog Route, consisting of 17 light vehicles and 33 heavy vehicles. On the mine site and along the Burntlog Route, road maintenance, dust abatement, and proposed erosion control EDFs are expected to be effective in minimizing the amount of sediment delivered, suspended, and deposited in action area streams.

Also, during the operational period, surface discharge of treated waters has the potential to generate turbidity. To address this effect, IPDES-permitted outfalls will be constructed with energy dissipation at their discharge location, a design expected to minimize sediment delivered to and suspended in action area streams at these locations.

As previously described, the greatest potential for Project-related increases in stream sedimentation will come during storm events causing overland flow across exposed soil, excavated areas, and roads. However, proposed erosion control BMPs (e.g., mulching, planting of vegetation, silt fences, removal of legacy materials, etc.) [Appendix A] to be employed should be effective in limiting the amount of sediment transported, suspended, and deposited in action area streams.

During operations, the GRAIP-Lite model results show an increase in sediment delivery, input, and accumulation, likely because of the additional 55 km (34.1 miles) of road compared to baseline. Table 57 shows the modeled change in sediment production, delivery, and accumulation during operations. Table 58 describes the sediment delivery by road segment, where modeling indicates that sedimentation in waterways under construction will be decreased compared to baseline conditions. Table 59 breaks down sediment delivery by structure type and watershed. Additional details regarding the GRAIP-Lite model and its results are provided in Tetra Tech 2024.



### **Table 57. Sediment Production, Delivery, and Accumulation under Baseline and Operations.**

**Note:** Tons per year indicated in brackets.

### **Table 58. Predicted Sediment Loading by Road Segment under Baseline and Operations (Terra Tech 2024).**



**Note:** Tons per year indicated in brackets.

#### **Table 59. Annual Sediment Delivery to Drainage Crossings Under Baseline and Operations (kilograms and tons/year) (Stantec 2024).**



**Note:** Tons per year indicated in brackets.

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In an effort to put these sediment delivery figures from Tables 57 to 59 in context, we compared these estimated sediment budgets<sup>12</sup> found in the literature. Parkinson et al. (2003) estimated the

 $12$  As defined in Parkinson et al. (2003), a sediment budget accounts for all sediment entering, leaving, and being stored in a river system.

annual sediment budget for the SFSR as ranging from 730 to  $12,775$  tons/mi.<sup>2</sup>/year. IDEQ (2011) also estimated the sediment budget for the SFSR, where they reported logging roads as responsible for up to 2,000 tons/year, with natural sources of sediment contributing 15,900 tons/year. Considering results from the GRAIP-Lite modeling presented in Tables 57 to 59, NMFS expects that the predicted increases in sediment delivery will not degrade the water quality PBF at the stream reach, watershed, or subbasin scales.

Although fugitive dust from road use will also contribute fines during the operations period, the greatest extent of sedimentation effects from dust will be concentrated at the mine site itself. However, due to the nature of sediment transport by streams, the geographic extent of the impact could extend farther downstream in the EFSFSR depending on site- and event-specific factors. The duration for traffic-related dust and erosion/sedimentation will last throughout the mine operation period. As previously mentioned, Perpetua plans to address these potential impacts through dust abatement measures on the roads and at the SGP. In dry months, Perpetua will spray water and dust abatement chemicals on mine haul roads as necessary to mitigate dust emissions in compliance with state and USFS requirements.

Although blasting has the potential to disturb soils, generate dust, and deliver sediment to action area streams, areas requiring blasting will typically occur on steeper side slopes and in upland areas (Stantec 2024). Limiting charge sizes, controlled blasting techniques, and proposed setback distances from streams designed to be protective of fish should prove effective at minimizing the amount of sediment delivered to streams by blasting.

As described for the construction phase, underground exploration is unlikely to deliver sediment to action area streams, but surface exploration may. New surface and underground exploration activities will be conducted during the operations phase. As previously described, drilling activities and road preparation/use will potentially cause increased sediment delivery to the EFSFSR and its tributaries, causing temporary turbidity pulses and subsequent sediment deposition. However, following project design criteria for the Golden Meadows Exploration project and other proposed erosion and sediment control EDFs are expected to effectively minimize the amount of sediment mobilized and transported to streams.

In summary, during the operations period, ground-disturbing sediment impacts to DCH will include temporary turbidity increases during runoff events and localized deposition of fine sediment in stream channels. Some sediment may be later resuspended and transported to downstream areas, although not expected to greatly affect the overall sediment budget at the reach, watershed, or subbasin scales. Turbidity increases during runoff events could affect the water quality PBF to the degree that it temporarily changes fish behavior, but these increases are unlikely to be long-lasting or severe enough, to affect the long-term conservation value of the water quality PBF.

*Closure/Restoration.* During the post closure period, erosion control BMPs will remain in place as mine disturbance is covered, reclaimed, and revegetated to control runoff from mine facility areas (Stantec 2024). Stream flow will be reintroduced into restored stream segments in Meadow Creek across the TSF. Upon completion of mining in the YPP, the pit will be backfilled and covered with a geosynthetic liner. A restored segment of the EFSFSR will be routed on top of the backfilled and covered pit and flow diverted through the tunnel will be re-routed to the restored channel. EDFs and BMPs described in Table 22 and Appendix A will be employed to pre-rinse diversion channels and introduce flows slowly to limit generation of new turbidity caused by these activities.

During the closure and reclamation phase, traffic along the Burntlog Route will be reduced to a total of 27 vehicle trips per day (year-round). Once the Burntlog Route is complete, the OSV route will return to Johnson Creek Road. Closure road usage will resemble the operational conditions with usage reducing over time. Post-closure road usage will resemble the existing conditions once new portions of the Burntlog Route have been obliterated and reclaimed as part of the closure activities.

Although facility reclamation will also generate sediment, the relatively flat nature of the construction site, combined with the application of proven erosion control practices, are expected to reduce the likelihood of effects from facility construction and reclamation to negligible levels.

Surface discharge of treated waters will continue until no longer necessary. Although these discharges have the potential to generate turbidity, these outfall structures will be constructed with energy dissipation at their discharge location to which should effectively minimize potential turbidity.

The duration for traffic-related dust and erosion/sedimentation will last throughout the mine construction, operations, and post-closure periods; however, these effects will be incrementally reduced during closure and reclamation due to reduced activity at the SGP and stabilization of disturbed areas.

In summary, sediment generating activities will include frequent, sporadic turbidity effects to water quality across action area streams, with those increases occurring most frequently occurring during runoff events and resulting in localized deposition of fine sediment in action area stream channels. These effects are likely to affect action area streams from the beginning of construction through closure and reclamation. Although turbidity increases during runoff events are likely to affect water quality to the point that they periodically affect fish behavior, they are unlikely to reach levels severe enough or long enough to affect the long-term conservation value of the Water Quality PBF.

# **Water Quantity**

The changes in surface water quantity described in this section are compared to those of the simulated existing conditions. Changes in surface water flows in the analysis area are expected to result primarily from:

- Water drafting for dust suppression and test drilling for the BRGI;
- Stream diversion around mine facilities affect the path of surface water flows;
- Interception of contact water and other mine-impacted water prior to runoff;
- Development and dewatering of three open pits;
- Groundwater production for consumptive use;
- Stream water diversion above the EFSFSR tunnel for consumptive use; and,
- Discharge of treated water.

These activities have the potential to modify the location and/or flow rate of stream flows in the analysis area.

*Pre-construction.* The BRGI portion of the proposed action would reduce flow in Johnson Creek by up to 6,050 gallons for one year (M Y-3), which represents approximately 0.00002 percent of available flow during an average year. This considered, there is very little water use in the Johnson Creek drainage, and flow in Johnson Creek is nearly unimpaired. Therefore, the effects of the BRGI portion of the project is expected to have only temporary, and very minor effects on the water quantity PBF.

*Construction and Closure.* The proposed action will result in diversion of water from the EFSFSR and from groundwater wells in the project area (Table 16), which will reduce flows in Meadow Creek and the EFSFSR). The proposed action will also result in changes in hydrology within the project area, which will alter flows in Meadow Creek, Sugar Creek, and the EFSFSR. The estimated effects on flow, by MY and reach, are in described in Table 60.

<b>Mine</b> Year	<b>Net Flow Reduction (cfs) Due to Water Diversion and</b> <b>Discharge</b>				<b>Flow</b> <b>Reduction</b> (cfs) Due to <b>Drainage Area</b> <b>Reduction</b>	<b>Total Net Flow</b> <b>Reduction in the</b> <b>EFSFSR Downstream</b>
	<b>Meadow</b> <b>Creek</b>	<b>EFSFSR</b> <b>Above</b> <b>Meadow</b> <b>Creek</b>	<b>EFSFSR</b> <b>Between</b> <b>Meadow</b> <b>Creek and</b> the YPP	<b>EFSFSR</b> <b>Between</b> the <b>YPP</b> and <b>Sugar</b> <b>Creek</b>	<b>Sugar Creek</b>	from Sugar Creek and <b>The SFSR Downstream</b> from the EFSFSR
$-2$	$-0.20$	0.00	$-0.20$	$-0.21$	$\boldsymbol{0}$	$-0.21$
$-1$	0.20	0.00	0.20	0.82	$\overline{0}$	0.82
$\mathbf{1}$	0.42	0.00	0.44	1.50	0.39	1.90
$\overline{2}$	0.77	0.00	0.76	3.42	0.39	3.82
3	0.84	0.00	0.86	2.40	0.39	2.80
$\overline{\mathcal{L}}$	1.40	0.00	1.35	2.61	0.39	3.01
5	$-0.21$	0.01	$-0.15$	0.97	0.39	1.37
6	1.12	0.02	1.34	2.48	0.39	2.88
7	2.23	0.02	2.38	3.69	0.39	4.09
$\overline{8}$	1.75	0.01	1.70	3.54	0.39	3.94
9	0.55	0.00	0.56	2.10	0.39	2.50
10	0.55	0.00	0.54	2.20	0.39	2.60
11	0.54	0.00	0.56	2.40	0.39	2.80
12	0.62	0.00	0.63	1.97	0.39	2.37
13	0.67	0.08	0.75	2.14	0.39	2.54
14	0.47	0.14	0.61	1.45	0.39	1.85
15	0.53	0.10	0.48	1.02	0.39	1.42
16	0.20	0.05	0.13	0.45	0.39	0.85
17	0.21	0.14	0.30	0.54	0.39	0.94
18	0.26	0.11	0.34	0.57	0.39	0.97
19	0.18	0.08	0.22	0.34	0.39	0.74
20	0.11	0.07	0.15	0.33	0.39	0.73
$20+$	0.00	0.00	0.00	0.00	0.39	0.39

**Table 60. Estimated reduction (cfs) in average July-September, flow, by reach and Mine Year (MY).** 

We compared the estimated effects on flow, provided in the BA, to USGS gage data to determine the downstream extent of the flow analysis. The estimated flow effects summarized in Table 60 translate to reductions in average January flow (i.e., the lowest average monthly flow) of 31.6% to 33.9% in the EFSFSR below the confluence of Sugar Creek, 5.2% to 5.5% in the EFSFSR below the confluence of Johnson Creek, 2.0% to 2.2% in the SFSR below the confluence of the EFSFSR, and 0.31% to 0.33 % in the Salmon River below the confluence of the SFSR. In systems with minimal development of water resources, such as the SFSR drainage, Tehan (2014) stipulates analyzing effects on flow downstream to the point at which effects are less than one percent of the lowest base flows. The flow effects of the proposed action are substantially greater than one percent of the base flow of the SFSR below the confluence of the EFSFSR, and are less than one percent of the lowest base flows in the Salmon River at the confluence of the SFSR. We therefore analyzed the flow-related effects of the proposed action downstream to the confluence of the SFSR and the Salmon River.

The water rights associated with the proposed action include conditions that restrict diversion when flows are less than 5.0 cfs from October 1 – June 29, or 7.25 cfs from June 30 – September 30, in the EFSFSR between Meadow Creek and the YPP; and when flows are less than 3.0 cfs in lower Meadow Creek. These "minimum flows" should ensure that the proposed action will not reduce flows sufficiently to substantially impair migration PBFs. The SFSR drainage typically has a high ratio of peak to base flows, and the estimated flow effects represent a relatively small percentage of peak flows. We therefore presume that effects on channel maintenance functions of flow would be relatively small. However, reducing flow in streams affects salmonid habitat in a variety of ways in additional to passage and channel function. Reductions in flow reduces water quality (Ebersole et al. 2001; Poole and Berman 2001; Dahm et al. 2003; Ebersole et al. 2003; May and Lee 2004; Miller et al. 2008), increases the amount of fine sediment in substrates (Baker et al. 2011), reduces health of riparian vegetation (Smith et al. 1991), reduces the amount and/or availability of drifting invertebrates (Townsend and Hildrew 1976; Elliott 2002; Boulton 2003; Lake 2003; Nislow et al. 2004; Harvey et al. 2006; Hayes et al. 2007; Miller et al. 2007; Carlisle et al. 2011; Jager 2014), reduces access to escape cover (Hardy et al. 2006), and reduces the amount of cold water refugia (Ebersole et al. 2015). Through these pathways, the flow reductions resulting from the proposed action will reduce quality of SR Spring/summer Chinook salmon spawning gravel, water quality, cover/shelter, food, space, substrate, and water temperature PBFs; and SRB steelhead water quality, substrate, forage, and natural cover PBFs. Reducing flow also affects the water quantity PBFs for both SR Spring/summer Chinook salmon and SRB steelhead, the floodplain connectivity PBF for SRB steelhead, and the space PBF for SR Spring/summer Chinook salmon.

Detecting and quantifying the effects of reducing flow on individual PBFs could possibly be accomplished with intensive modeling, but even the most refined flow habitat models tend to underestimate optimal flows (Rosenfield et al. 2016) suggesting that modeling would likely underestimate the effects of flow reductions on fish habitat. Because relationships of population productivity and streamflow are well documented (see section 2.10.4.3), we used the relationships of population productivity and flow to estimate the flow-related effects of the proposed action on SR Spring/summer Chinook salmon and SRB steelhead DCH. The magnitude of effects peak at MY 7 and are greatest in lower Meadow Creek, where the estimated reduction in flow would reduce productivity of SR Spring/summer Chinook salmon and SRB steelhead DCH by 29% and 62%, respectively. Estimated reductions in productivity due to reduced flow, by stream reach and MY, are in Appendix G.

The flow-related effects of the proposed action would affect DCH within the EFSFSR Chinook salmon and SFSR Chinook salmon population areas, and in the SFSR steelhead population area. On a population scale, the flow-related effects of the proposed action would reduce productivity of DCH in the EFSFSR Chinook salmon, SFSR Chinook salmon, and in the SFSR steelhead population areas by maximum of 2.2%, 0.15%, and 0.81%. This maximum effect will occur during MY 7, and will gradually decrease between MY 7 and MY 20 (Appendix G). The average reduction in productivity for MYs -2 through 20 will be 1.1%, 0.007%, and 0.36%, respectively, for DCH in the EFSFSR Chinook salmon, SFSR Chinook salmon, and in the SFSR steelhead population areas. After mining activity stops, in MY 20, water will no longer be diverted, and streamflow throughout most of the mine footprint will return to normal. However, flows in Sugar Creek and the EFSFSR would continue to be reduced due to filling of the WEP Lake. This flow

reduction would reduce annual productivity of the EFSFSR Chinook salmon, SFSR Chinook salmon, and in the SFSR steelhead population areas by 0.13%, 0.001%, and 0.082% respectively. This reduction would last for approximately 57 years post mining (i.e., through MY 77).

# **Physical Habitat Conditions – Spawning Gravel, Passage, and Space/Cover/Shelter.**

Several PBFs (i.e., physical habitat conditions, safe passage, and space/cover/shelter) represent elements of salmon and steelhead DCH for spawning, rearing, and migration. Potential effects for these PBFs were described in the BA as potential effects to the habitat access and habitat element WCIs (i.e., interstitial sediment, LWD, pool frequency/quality, off-channel habitat, and refugia). A discussion of how the SGP may affect each of these PBFs follows.

# *2.10.1.3.1. Spawning Gravel/Substrate*

The spawning gravel/substrate PBFs are essential to steelhead and Chinook salmon spawning and rearing. The condition of the Interstitial Sediment WCI is indicative the condition of these PBFs, characterized as currently FA for both the EFSFSR and Johnson Creek watersheds (Table 30; and BA Appendix C [Stantec 2024]).

Salmonid spawning habitats are created by and depend on channel characteristics and complexities that cause hydraulic sorting and gravel accumulation into suitable spawning beds. If well established, these beds are relatively resistant to scour during periods of egg incubation. Increased sediment deposition may lead to increased embeddedness of downstream substrates. Fine, redeposited sediments have the potential to adversely affect primary and secondary productivity (Spence et al. 1996), and reduce incubation success (Bell 1991) and cover for juvenile salmonids (Bjornn and Reiser 1991).

As previously discussed, ground disturbing activities and roadwork are expected to result in sediment delivery to action area streams, beginning during the pre-construction phase, tailing off during closure and restoration. Sediment delivered by the project is generally expected to first settle out within 1,000 feet of the points of introduction. Deposited sediments are likely to later be transported farther downstream during storm events and periods of high flow. For Chinook salmon habitat, substrate-related effects are most likely to be realized in the EFSFSR and Johnson Creek during the construction phase, and the EFSFSR, Johnson Creek, Meadow, Sugar, and lower Burntlog Creeks during the operations and closure/restoration phases. For steelhead, substrate-related effects are most likely to be realized in the EFSFSR, and Johnson, Burntlog, Trapper (lower 0.75 miles), and Riordan (downstream from Riordan Lake) Creeks during all phases of the project. Steelhead substrate in Sugar Creek will be potentially affected during the operations and closure/restoration phase of the SGP.

Rio ASE (2021) identified EFMC as the single largest source of sediment delivery to the EFSFSR in the project area, contributing up to an estimated 77.4 tons per year to the upper EFSFSR watershed. The proposed action will attempt to address this chronic source of sediment delivery, by: (1) raising groundwater levels to restore wetland function and improve the incised stream channels in upper EFMC; and (2) by stabilizing the confined, high gradient channel in the lower reach.

During and after mine operations, EFMC will flow into the restored Meadow Creek channel around HFP, habitat expected to be used for rearing and spawning Chinook salmon and steelhead as soon as MY3 (table 45). Increases in fine sediment deposition within stream channels have the potential to decrease spawning gravel suitability and decrease benthic invertebrate production within gravel riffles, potentially impacting spawning/incubation and rearing/feeding life stages of Chinook salmon and steelhead. However, spawning substrates are generally FA in action area streams, and GRAIP-Lite modeling suggests that the amount of sediment delivered to action area streams will decrease in comparison to the baseline condition. With application of proposed sediment control BMPs, the deposition of sediment in action area waters is predicted to be measurable but not severe. It is not expected to degrade this PBF in the temporary or short-term; and following closure of the mine, obliteration of access roads, and restoration efforts in the EFMC, the proposed action is expected to result in a substantial decrease in annual sediment delivery into Meadow Creek and the EFSFSR. Therefore, the effects of the proposed action on substrate will be minor and will not affect the conservation value of this PBF in the short term, and may benefit this PBF in the long term.

## *2.10.1.3.2. Fish Passage*

Habitat access was characterized as FUR for the EFSFSR, and ranged between FA and FUR in Johnson Creek (Table 30). BioAnalysts (2021) evaluated fish passage for the EFSFSR and its tributaries in the project area, identifying five artificial barriers and one natural barrier in fishbearing streams. Barriers were also classified as either complete or partial (Table 32). Major barriers to fish passage at the mine site include: (1) the high gradient cascade in the EFSFSR upstream from the YPP Lake; (2) EFSFSR box culvert; and (3) the high gradient cascade in Meadow Creek upstream from the confluence with the EFMC (at the downstream end of the engineered channel). The cascade upstream from the YPP Lake is a complete barrier to natural fish passage. When Chinook salmon are not stocked upstream of this barrier, the barrier blocks about 92 percent of available Chinook salmon DCH upstream of Sugar Creek (ESS 2022). The other two major barriers, the EFSFSR box culvert and Meadow Creek barriers, are flowdependent partial barriers that can block seasonal migration, and only hinder migration of fish that reside in or were stocked upstream from the YPP Lake cascade barrier (i.e., translocated Chinook salmon) (Stantec 2024).

During the construction phase, the proposed action will begin to address fish passage in the EFSFSR subbasin. The tunnel to bypass YPP Lake, and the high gradient cascade upstream from the lake, will be complete by MY -1. Designs for the tunnel have been reviewed by NMFS' fish passage engineer, and although it is suspected that the tunnel will pass all life stages of ESAlisted salmonids, whether fish will actually use the tunnel and what the actual passage rate will be are both unknowns. However, because Perpetua will alternatively use trap and haul (facility also reviewed by NMFS' fish passage engineer) to get fish upstream of the YPP cascade should passage not work as designed, NMFS assumes that ESA-listed salmonids will access stream segments upstream of the cascade beginning in MY -1 whether the tunnel functions as intended or not.

The EFSFSR box culvert will also be removed by MY -1. Although the partial gradient barrier in Meadow Creek just upstream from EFMC will also be removed during the construction phase of the Project, a new barrier will be created in MY -2 just upstream from the existing barrier to

prevent fish passage into upper Meadow Creek where the TSF will be constructed. This new barrier, will block access to approximately 0.43 miles of otherwise accessible Chinook salmon DCH through MY 12. By MY 18, the only barrier that impacts Chinook salmon DCH will be the steep gradient section of Meadow Creek associated with the TSF buttress created as part of the SGP, a barrier that will remain indefinitely and will continue to eliminate volitional access to approximately 3.87 miles of Chinook salmon DCH (ESS 2022).

Although Chinook salmon were already using habitat upstream from the YPP cascade in years following release of excess hatchery fish, removal of these barriers will result in volitional passage into the upper EFSFSR for the first time since the 1930s. Passage upstream of the cascade will provide access to up to 12.2 miles (19.65 km) of potential Chinook DCH, 5.51 miles (8.87 km) of which was modeled as having intrinsic potential for Chinook spawning and rearing (Stantec 2024).<sup>13</sup>

The only surface water diversion associated with the proposed action will be on the EFSFSR between Sugar Creek and the YPP. This diversion will meet NMFS criteria for upstream and downstream fish passage (NMFS 2022a), and will therefore not impair upstream fish passage. However, some juveniles will migrate downstream via the bypass system, which will cause migration delay and will increase migration mortality, thereby adversely affecting the fish passage PBFs. Reduced flow, caused by surface and groundwater diversion, also has the potential to impair fish passage. However, the conditions for water rights 77-7122, 77-7293, 77- 7285, 77-14378, preclude diversion when flows are less than the stipulated minimums (i.e., 5.0 cfs from October 1 to June 29, and 7.25 cfs from June 20 to September 30) at the EFSFSR POD, and limit diversions when flows are less than 25 cfs in the EFSFSR below Sugar Creek. Also, the conditions in water right 77-14378 preclude diversion when flows in Meadow Creek are less than 3.0 cfs. Adherence to these conditions should ensure that flows are adequate for upstream and downstream migration of Chinook salmon and steelhead in Meadow Creek and in the EFSFSR within and downstream from the project area.

As outlined in the BA (Stantec 2024), during the construction of the Burntlog Route and other temporary roads, culverts will be constructed or replaced, which may affect fish access. There are 18 existing crossings along the Burntlog Road (FR-447) that will be replaced and 10 new crossings along newly constructed portions of the Burntlog Route. There are approximately 53 miles of stream segments upstream of the Burntlog Route. Currently almost all stream crossings along the Burntlog Road are impassable culverts, particularly at low flow conditions. The key perennial streams that will be crossed are Burntlog, Trapper, and Riordan Creeks, the headwaters of which do not support Chinook salmon nor steelhead. The route will cross DCH for SR Basin steelhead once in upper Burntlog Creek, and cross SR Spring/summer Chinook DCH once in upper reaches of the EFSFSR. Chinook salmon and steelhead have not been found this far upstream in either of these locations during past survey efforts. Because all new or reconstructed crossings are required to be fish passable, the proposed action will only impact the fish passage PBF temporarily as work areas are dewatered to construct the crossings. By the end of the construction phase, retrofitted crossings are likely to improve or reestablish fish passage where it had previously been reduced or blocked.

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<sup>&</sup>lt;sup>13</sup> Although steelhead will also be able to access approximately 5.42 miles (8.72 km) of intrinsic potential habitat upstream of the cascade, there is no DCH for steelhead present upstream of the cascade.

Riparian roads can also affect the passage PBF by constraining the natural migration of the stream channel where channel migration zones are present. Stream crossings can affect passage by restricting channel geometry and preventing or interfering with migration of adult and juvenile anadromous fish (Furniss et al. 1991). New road segments constructed for the Burntlog Route will cross multiple streams, but will generally not closely parallel or constrict streams. Most of the constructed and retrofitted stream crossings will occur high in tributary headwaters, well upstream of waters occupied by ESA-listed salmon and steelhead, and therefore having little if any effect on this PBF.

Transmission lines will cross streams or border streams occupied by ESA-listed fish, including the SFSR, and Cabin Creek, Trout, Johnson, Trapper, and Riordan Creeks. No new stream crossings are anticipated for transmission line access roads. Therefore, transmission line construction is not expected to affect fish passage.

Migratory fish passage could also be affected through chemical contamination resulting from an accidental spill, stormwater discharge, wastewater treatment plant discharge, or nonpoint sources of pollution associated with permanent mine facilities. Fish passage is not expected to be affected by chemical spills because spills in quantities sufficient to impede passage are not expected to occur. Determining whether a contaminant or contaminant mixtures will impede fish passage is difficult. First, it is difficult to predict avoidance responses with any degree of certainty because avoidance responses depend on many different factors such as water chemistry, species and life stage exposed, variability in metal mixtures, temperature, cover, shade, acclimation, prey availability, presence of predators, and competition with other fish. Alone, copper can elicit avoidance responses in juvenile Chinook salmon and steelhead in soft water at concentrations as low as 0.7 µg/L (Hansen et al. 1999a). Researchers have studied fish avoidance of chemical mixtures in the lab and in the field. In laboratory tests, copper and zinc mixtures have been shown to act together to cause a lower threshold of avoidance than would result from either metal alone (Giattina and Garton 1983). Hansen et al. (1999a) reported that behavioral avoidance to copper and cobalt mixtures in soft water differed greatly between the rainbow trout and Chinook salmon. Rainbow trout avoided 1.6 μg/L concentration of copper and 180 μg/L cobalt individually, but when exposed to a mixture of the two chemicals, the response occurred at significantly lower concentrations (i.e., a mix of 2.6 μg/L copper and 2.4 μg/L cobalt). Chinook salmon were more sensitive than rainbow trout, avoiding mixtures of 1.0 μg/L copper and 0.9 μg/L cobalt (Hansen et al. 1999a).

Studies of avoidance behavior to the mixture of metals in the Clark Fork River found that rainbow trout were more sensitive than brown trout, which in part may explain why rainbow trout populations appear to be more severely affected than brown trout populations (Woodward et al. 1995; Hansen et al. 1999b). Cutthroat trout avoided a metals mixture of 6 μg/L copper, 0.3 μg/L cadmium, 0.6 μg/L lead, and 28 μg/L zinc (Woodward et al. 1997). Rainbow trout avoided all metal concentrations tested from 10 to 1,000 percent of a fixed ratio of ambient metal concentrations (12 μg/L copper, 1.1 μg/L cadmium, 3.2 μg/L lead, and 5 μg/L zinc) (Hansen et al. 1999b). In a Coeur d'Alene River study (Goldstein et al. 1999), 70 percent of the adult Chinook salmon migrated up the North Fork (the control stream with zinc concentrations of around 9 µg/L) and 30 percent migrated up the South Fork (a mining impacted stream with zinc concentrations of about  $2,200 \mu g/L$ ). The authors concluded the salmon migrated up the North

Fork because it had lower ambient concentrations of zinc and other metals. However, the North and South Forks make up roughly 70 and 30 percent of the flows in the Coeur d'Alene River, respectively. Adult salmon movements in rivers in the absence of any homing cues (as was the case in this study) tend to simply follow larger flows (Anderson and Quinn 2007). Whether or not metals concentrations influenced adult salmon movement in this study is debatable. The proposed action is not expected to result in contaminant concentrations that will elicit avoidance.

Fish passage is not expected to be meaningfully affected by chemical contamination from the project. First, predicted average copper concentrations are less than 0.5 µ/L, which is below thresholds known to elicit avoidance. However, predicted maximum copper concentrations will be above concentrations that have been associated with some avoidance (e.g., lowest observed effect concentration,  $EC_{20}$ , etc.). The predicted maximum concentration may cause a few individual fish to avoid stream reaches during operations, early closure, and/or late closure, depending on the stream reach. While some fish may avoid reaches, other fish are expected to not avoid the impacted reaches. This is evidenced by juvenile fish utilizing habitats with copper concentrations currently greater than, or nearly equal to, predicted concentrations. Finally, mixing zones for copper and other contaminants are not expected to be authorized; further minimizing the potential reducing fish passage. When considering the risk of a contaminant spill and predicted contaminant concentrations, the conservation value of the safe passage PBF will not be reduced at the scale of the action area. for the following reasons: (1) individual fish appear to utilize habitat that is physically accessible to them currently and the proposed action will not substantially alter water quality relative to current conditions; (2) predicted contaminant concentrations are expected to generally be below thresholds known to elicit avoidance; (3) spills are not expected to occur; and (4) mixing zones for contaminants in point source discharges are not expected to be authorized.

Once the mining activities at the YPP are completed (MY 11), the EFSFSR will be reconstructed into a volitionally passable stream channel across the top of the YPP backfill. As soon as the stream channel restoration is complete, and EFSFSR flows are routed through the restored channel, a long-term improvement to the fish passage PBF in the upper EFSFSR watershed should be fully realized. However, NMFS' engineering review expressed a potential concern with the lack of redundancy for the liner used in the YPP backfill. The concern was raised that should the liner be damaged or ultimately fail, surface flows could go subsurface creating a new passage barrier and blocking access to habitat in the upper watershed. The ultimate depth of the geosynthetic layer is not currently clear; and it appears as though the synthetic layers may be placed as low as (or very near) the lowest calculated scour depth.

## *2.10.1.3.3. Space, Cover, and Shelter*

Space, cover, and shelter are provided by instream habitat, and the functionality of the LWD, pool frequency, pool quality, off-channel habitat, and refugia WCIs are a direct indication of the condition of these PBFs. Despite significant historical impacts to riparian vegetation caused by mining and streamside roads in the action area, the amount of LWD in action area streams is FA in both the EFSFSR and Johnson Creek watersheds. Large, landscape-scale wildfire and landslides have contributed large amounts of LWD to streams; and these large amounts of LWD have in turn created habitat complexity and quality refugia by increasing both the number and

quality of pools in action area streams (generally FAR in EFSFSR, and FA in Johnson Creek) (Table 30; and BA Appendix C [Stantec 2024]).

During construction (MYs -3 to -1), several activities have the potential to affect space, cover, and shelter available to salmon and steelhead, including: (1) Burntlog route construction; (2) transmission line construction; (3) EFSFSR tunnel; (4) Meadow Creek realignment; (5) stream enhancement in EFSFSR and lower Meadow Creek; and (6) YPP Lake dewatering.

Crossings along the Burntlog route cross DCH in only three locations, bridges at Johnson Creek (Chinook and steelhead DCH), upper Burntlog Creek (steelhead DCH), and upper EFSFSR (Chinook DCH). Space, cover, and shelter will only be temporarily affected at these stream crossing structures while they are being constructed/upgraded and the work areas are dewatered. Because all new or reconstructed crossings are proposed to be fish passable, the construction of the road network will maintain or restore passage at all crossings, and in the long term, is likely to provide access to additional steelhead DCH (e.g., upper Burntlog Creek) and therefore additional space, cover, and shelter that may have previously been seasonally reduced or blocked.

Transmission lines will cross or border streams occupied by ESA-listed fish, including the SFSR, and Trail, Curtis, Cabin Creek, Trout, Johnson, Trapper, and Riordan Creeks. No new stream crossings are anticipated for transmission line access roads. Therefore, transmission line construction is not expected to affect access to the space, cover, or shelter PBFs. Should crossings need to be upgraded, they will follow the same EDFs as for the Burntlog route, affecting access to habitat only temporarily while streams are dewatered to construct the crossing(s).

As previously mentioned, at the beginning of the construction phase, volitional fish passage for Chinook salmon and steelhead is not possible beyond the cascade immediately upstream of YPP Lake (unless planted). Until the tunnel is complete, and the YPP Lake is drained, there will be no change to the space, cover, or shelter PBFs. Construction of the EFSFSR tunnel and fishway will take two years to complete, completed in MY -1. The tunnel will be constructed around the existing YPP Lake and the cascade barrier upstream from the lake. Once access is established, Chinook salmon will gain volitional access to approximately 5.5 miles DCH with intrinsic potential, or 5.5 miles of additional space, cover, and shelter in the EFSFSR and Meadow Creek upstream of the cascade.

The dewatering of the YPP Lake, is scheduled to begin in MY -1. At this point in time, a fish weir will be constructed on the EFSFSR to redirect fish into the tunnel, and fish passage will be blocked in the EFSFSR downstream from the YPP Lake through MY 11. In addition to habitat lost in the YPP Lake itself, placement of this weir will result in the 12-year loss of approximately 500 linear feet of space, cover, and shelter in the EFSFSR for both Chinook salmon and steelhead DCH.

Also, during construction (MY -1), channel restoration and enhancement will occur in Meadow Creek downstream from the HFP. Habitat enhancement will also occur in the EFSFSR upstream from the YPP restoration reach during this same timeframe. Conceptual stream designs have

largely been developed to accommodate physical site conditions and geomorphic suitability, but biological objectives were used to refine the design within each reach to maximize habitat potential to the extent practicable. Although biological objectives and associated stream design (restoration and enhancement) actions will improve migration/passage and provide limited spawning potential for Chinook salmon DCH, designs were tailored to optimize spawning and rearing habitat for steelhead which is not DCH upstream from the YPP (Rio ASE 2021).

Restored stream segments and habitat enhancement will involve the addition of instream habitat structures, including LWD jams, boulder clusters, and excavated pools, intended to improve habitat conditions for salmonids within reaches not directly impacted by the SGP but previously impacted through historical mining. Revegetation will also take place to develop riparian, wetland, and upland zones for long-term bank stability, future LWD recruitment, overhead cover, shade, terrestrial/wetland habitat, and soil productivity. See Appendix E, Figures D-1 to D-12, in Rio ASE (2021) for typical design drawings for bank treatments, willow planting, constructed riffles, floodplain roughness, toe log and LWD structures,

Enhancement in the EFSFSR will begin in MY -1 (EF2 and EF4; Figure 34). Efforts will be focused on increasing hydraulic and geomorphic diversity while removing potential fish passage barriers. Enhancement in these stream reaches is intended to provide habitat for spawning and rearing Chinook salmon and steelhead. LWD and rock clusters will be used to enhance instream conditions, increase instream friction, sort sediment, and create localized velocity gradients. Grade control structures such as engineered riffles and/or channel-spanning rock or wood may be used to facilitate the development of relatively large pools intended to accommodate adult salmonids migrating upstream through these relatively steep reaches. The newly constructed EFSFSR, constructed across the surface of the YPP backfill, will be designed to interact with its floodplain, and will include side-channels, LWD, boulders, wetlands, and the lowest reaches of Hennessy (HC2), Garnet (MNC 1 and 2) and Midnight Creeks (MNC1 and MNC2) (Figure 34) (Rio ASE 2021).

In MY 3, a segment of lower Meadow Creek (MC4B, MC5, and MC6A; Figure 34) will be realigned around HFP in a bioengineered channel designed to provide short- and long-term function for Chinook and steelhead spawning and rearing habitat (Chinook DCH only). The channel and its floodplain will be lined to prevent excessive losses to groundwater during mining. The liner will be placed at a sufficient depth below the streambed and entire floodplain to accommodate a perched aquifer that will enable the natural function (e.g., channel migration, scour, hyporheic flow, etc.) of the stream for its geomorphic character (Rio ASE 2021). After mining, groundwater levels are expected to rebound, making the stream liner in this reach obsolete. The lowest reach of Meadow Creek will be enhanced with LWD and boulder clusters for improved habitat. Sediment controls and the rock drain will also be installed in EFMC at this time to address chronic sediment delivery to instream habitat in Meadow Creek. The lowest reach in EFMC (BC3) (Figure 34) is designed to provide additional rearing habitat for Chinook salmon and steelhead. This restoration will be complete in EFMC by MY 3 (Rio ASE 2021).

By MY 10, the YPP backfill will be complete, and by MY 11 a restored stream channel and Stibnite Lake will be created across the surface of the backfill. Once the stream channel restoration is complete, the EFSFSR flows will be routed through the restored channel and

Chinook salmon and steelhead will gain access to the space, cover and shelter PBFs not available to ESA-listed Chinook salmon since the 1930s.

The lower reaches of Midnight (MNC1 and 2) and Garnet Creeks (GC1 and 2; Figure 34), where they flow across the EFSFSR floodplain, have been designed to provide juvenile rearing habitat for Chinook salmon and steelhead. These reaches of Midnight Creek will be constructed in MYs 12 and 13. In Garnet Creek, the lowest reach (GC2) will be restored downstream from the plant from MY -3 to -1, while reach GC1 will not be restored until plant decommissioning is complete in MY 23. The uppermost reach of Meadow Creek that will be accessible to anadromous fish (MC4A), will be reconstructed in MY 18, with a biological objective of providing additional spawning and rearing habitat for Chinook salmon and steelhead (Stantec 2024; Rio ASE 2021).

Overall, although temporary and short-term effects will occur to these PBFs for Chinook salmon upstream of the YPP, and for both Chinook and steelhead from the YPP downstream to the confluence with Sugar Creek, proposed stream restoration and enhancement should ultimately increase habitat complexity in reaches accessible to anadromous fish in the upper EFSFSR and Meadow Creek, resulting in long-term improvements to the space, cover, and shelter PBFs in these portions of the action area. However, as previously mentioned, NMFS' engineering review expressed a potential concern with the lack of redundancy for liners used for engineered stream channels, concern that surface flows could go subsurface should the liners fail. These concerns also apply to all new channels in Meadow Creek, the EFSFSR, and their tributaries.



**Figure 34. Proposed Stream Design Reaches (Rio ASE 2021).**

## **Floodplain Connectivity**

The floodplain connectivity PBF, a PBF specific to rearing steelhead, is represented by the floodplain connectivity WCI. This WCI was characterized as FAR for both the EFSFSR and Johnson Creek watersheds (Table 30). Streamside roads along Johnson Creek, Burntlog Creek, Sugar Creek, and the EFSFSR negatively affect floodplain connectivity in the action area. Floodplain connectivity has also been negatively affected by past mining practices in the upper EFSFSR and Meadow Creek, where valley bottom roads and past restoration efforts have constrained the ability of the streams to function and interact with contaminated mine tailings.

Although not a Chinook salmon PBF, the DCH designation for Chinook includes the area within 300 feet of the OHWM. As soon as MY11, functional floodplains will be created for Chinook salmon DCH across Meadow Creek and much of the mining footprint upon closure (Appendix E, Figures G-5 and G-6, Rio ASE 2021). In terms of benefits realized to the floodplain connectivity PBF for steelhead, creation of functional floodplains bordering the EFSFSR constructed across the YPP backfill (i.e., the upstream extent of steelhead DCH) should result in localized improvements to the floodplain connectivity PBF for steelhead beginning MY 11.

# **Riparian Vegetation**

The riparian vegetation PBF, specific to Chinook spawning, rearing, and migration, is represented by the Watershed Conditions WCIs (i.e., road density/location, disturbance history, RCAs, and disturbance regime). These WCIs were characterized as FUR for the EFSFSR, where mining and streamside roads have heavily impacted overall watershed condition and riparian vegetation along the EFSFSR and Sugar Creek. Wildfires in 2017 have also impacted riparian vegetation to a lesser degree in the EFSFSR watershed. In the Johnson Creek drainage, these WCI have been characterized as ranging between FAR and FUR, with historic wildfire and streamside roads along Johnson and Burntlog Creeks heavily impacting this PBF (Table 30; and BA Appendix C [Stantec 2024]).

Riparian and wetland vegetation provide a diversity of ecological functions and services that include avian, wildlife, pollinator, fish, and insect habitat, refugia, and food resources, as well as shade that buffers stream temperatures. Additionally, riparian and wetland vegetation reduces scour and sediment loss and, through uptake, filtering, and microbial processing, riparian and wetland vegetation can help reduce target compounds such as nutrients, heavy metals, and undesirable effluent (Tetra Tech 2018, as cited in Rio ASE 2021).

Roads built in riparian areas often eliminate part of the riparian vegetation (Furniss et al. 1991), reducing LWD recruitment and shade. Disturbance of riparian vegetation has the potential to result in decreased shade and increased solar radiation, which potentially could further increase water temperatures in the action area. Riparian vegetation will be disturbed during construction of the Burntlog Route, during construction of the powerline corridor, and across the mine footprint in the upper EFSFSR watershed. Elevated water temperatures can affect the Chinook salmon riparian vegetation PBF to the degree that it affects salmonid physiology, growth, development, life history patterns, disease, and predator-prey interactions (Spence et al. 1996).

The BRGI portion of the proposed action will result in travel across an EFSFSR riparian wetland to access site B-22 and may result in felling trees in RCAs. Because BMPs described to protect the EFSFSR riparian wetland (Table 20) would likely be effective and few trees (possibly none) would be felled within 150 feet of streams, the overall effect of the proposed action on the riparian vegetation PBF from this portion of the action will be minor.

Riparian vegetation removal as a result of powerline construction has the potential to decrease bank stability, which may result in bank erosion and an increase in sediment to SFSR and Johnson Creek watersheds. Overall, the transmission line ROW overlaps with 132.4 acres of RCAs (including non-anadromous RCAs). However, the new portion of the powerline ROW overlaps with 14.8 acres of RCAs (Stantec 2024). For the new portion of the powerline, vegetation will need to be cleared for structure footings (size). However, the utility poles are not directly along creeks or within RCAs. Instead, vegetation removal in RCAs, if it occurs, will target taller vegetation that exceeds the maximum height specified for the wire and border zones. Because lower-growing vegetation will be maintained along the ROW, and utility poles are not directly along the creeks or within the RCAs, the localized removal of this small amount of riparian vegetation should not appreciably alter riparian functions and processes at the river reach or watershed scales.

Similarly, construction of access roads (including the Burntlog Route) will occur along headwater tributaries of Johnson Creek. Riparian vegetation will not be removed except along the new road, and where possible, removal of trees will be avoided. The construction of the road crossings may result in a loss of riparian vegetation, as well as reduced vegetation overhead cover and stream shade until riparian vegetation can be reestablished (Stantec 2024). It is anticipated that there will be limited, if any, effects to the PBF as a result of construction of the Burntlog Route.

Perpetua will seed and plant stream reaches, riparian areas, and wetlands with native plant species present currently in existing wetlands and riparian areas along streams within the SGP footprint. Seed mixes, live stakes, and nursery-grown container plants and plugs of native graminoids, forbs, shrubs, and trees would be utilized for revegetation (Tetra Tech 2023). The revegetation plan has been developed for specific riparian, wetland, and upland zones to improve long-term bank stability, LWD recruitment, overhead cover, shade, and terrestrial/wetland habitat (Brown and Caldwell 2021e). In an effort to provide shade to action area streams, riparian plantings will be 18 feet wide, with a higher percentage of taller and denser species (e.g., spruce trees) than originally planned (Brown and Caldwell, Rio ASE, and BioAnalysts 2021).

The riparian planting plan is described in Appendix F of the Stream Design Report, where local native seed, cuttings, and containerized materials will be used wherever feasible to increase the likelihood of survival (Rio ASE 2021). Riparian habitat restoration and enhancement will occur concurrently with mining where practicable, beginning early in the mining process, during construction, and would continue through site closure. The restoration of EFMC (upstream from Chinook DCH) would occur early in mine construction and operations, and the enhancement of portion of the EFSFSR would occur during the early years of mining (Brown and Caldwell, Rio ASE, and BioAnalysts 2021), with riparian plantings occurring in the EFFSR and lower Meadow Creek planted by MY 1 to 3 (Brown and Caldwell 2021e). Riparian plantings incorporated into the newly constructed channel across the YPP backfill will be planted by MY 11.

Perpetua proposes to conduct riparian monitoring on a regular basis, looking to measure changes over time to establish species survival, composition, and diversity, while controlling the influence of invasive weeds in newly restored riparian habitats. Representative sampling within the riparian area and photo-documentation will be used to provide information on the performance of newly planted riparian communities (Brown and Caldwell, Rio ASE, and BioAnalysts 2021).

Proposed weed treatment activities have the potential to affect riparian vegetation in the action area. However, by following the PNF weed treatment program, and applying PDFs and BMPs designed to be protective of non-target riparian vegetation, weed treatment activities in RCAs are expected to locally improve riparian conditions in the future through decompaction of soils and establishment of native vegetation. For these reasons, the conservation value of the riparian vegetation PBF within the action area is not expected to be meaningfully degraded by weed treatment activities.

In summary, riparian vegetation has been negatively affected by streamside roads and mining activity in the project area, and additional riparian vegetation will be disturbed as a result of project implementation. However, proposed revegetation efforts will work to begin restoration of functional riparian vegetation as soon as practicable, beginning as early as MY 1. This considered, the project will have temporary and short-term negative effects to this Chinook salmon PBF, but will result in a long-term improvement to this PBF as the Burntlog Road is obliterated and revegetation efforts at the mine site take hold. Implementation of a successful monitoring plan, along with an effective adaptive management strategy to ensure that revegetation efforts meet desired objectives, will be key to ensuring that the proposed action does not affect the long-term function of the riparian vegetation PBF.

## **Food/Forage**

The food/forage PBF can be affected by changes in water quality (e.g., stream temperature, chemical contamination), water quantity, sedimentation, clearing of riparian vegetation, and channel relocation.

Chemical contamination resulting from an accidental spill or through point and nonpoint sources of pollution has the potential to affect the quantity and quality of prey for juvenile salmon or steelhead. Mortality of aquatic invertebrates from a spill of toxic materials would be dependent upon the type and amount of material spilled. Since toxicity is expected to attenuate in a downstream direction, and spills most likely to occur are expected to be in small quantities, mortality from a spill is not likely to extend more than a mile or two downstream. These discrete events, if they were to occur, would be expected to temporarily reduce the food/forage PBF in localized area. Elevated concentrations of metals (e.g., mercury, arsenic, antimony) or TDS from point and nonpoint sources at the mine site may alter benthic community assemblages, reducing the availability of preferred invertebrate prey organisms. Such reductions or changes in prey availability could translate to adverse effects on juvenile salmonid populations.

In the Panther Creek field studies that NMFS (2014) reviewed, no obvious effects to salmonids from extinctions of macroinvertebrate species were observed. This suggests either or both that juvenile salmonids are able to switch prey when preferred prey are diminished, or that the food web effects were too subtle to tease out of the natural variability inherent in field monitoring studies using available information. Restoration efforts in Panther Creek have led to substantial reductions in copper concentrations (i.e., copper concentrations improved to below the BLMbased chronic copper criterion of  $\sim$ 2.3  $\mu$ g/L), and within a few years, stoneflies reappeared in the invertebrate assemblage. Mayflies also reappeared in the assemblage, although it took longer for mayfly species to reestablish in the system. Predictive modeling suggests that water column concentrations of contaminants are not going to be at levels that will substantially alter benthic invertebrate communities. Furthermore, concentrations of arsenic and antimony are expected to be reduced. Given macroinvertebrate communities are considered to be healthy (i.e., high MBI and O/E scores) under current water quality conditions (EcoAnalysts, Inc. and MWH Americas, Inc. 2017; LaVoie and Stantec 2019), it is reasonable to conclude that prey quantity will not be reduced and prey composition will not be altered from predicted changes in chemical concentrations. Given that juvenile salmonids are opportunistic feeders, as long as a diverse group of macroinvertebrates are protected, some loss of prey items would not be expected to reduce individual fitness of juvenile salmonids rearing in the area.

It is not certain to what extent arsenic and mercury will bioaccumulate in the prey base, indirectly affecting fish survival and growth downstream. As described in the Section 2.10.1.1, arsenic concentrations in prey of approximately 20 mg/kg dw or greater can lead to harmful concentrations of arsenic in fish. Concentrations of arsenic are currently above this threshold in macroinvertebrates collected from the EFSFSR below Meadow Creek and in Meadow Creek below the SODA. Macroinvertebrate tissue concentrations are also elevated in Sugar Creek near West End Creek. The speciation of arsenic in macroinvertebrate tissues are unknown and in light of this uncertainty, we assume the more toxic form of arsenic is prevalent. Arsenic concentrations, while decreased, are expected to remain above levels that will lead to reduced prey quality. Mercury concentrations will increase in streams within the Project area, resulting in greater potential for bioaccumulation of mercury in macroinvertebrates. As described in Appendix F, dietary mercury concentrations exceeding 2 mg/kg dw (DePew et al. 2012) can result in adverse behavioral effects. Berntssen et al. (2003) documented increased oxidative stress in Atlantic salmon parr fed diets of 4.35 mg/kg dw MeHg. Concentrations of mercury in macroinvertebrate tissues in the EFSFSR are below this threshold. Maximum mercury concentrations of 0.43 mg/kg dw were document in the EFSFSR downstream of Sugar Creek. Concentration of mercury in macroinvertebrate tissues exceeded 2 mg/kg dw in Sugar Creek. Increased water column mercury concentrations will further diminish the prey quality; however, the extent to which it is diminished will depend on the methylation potential in the habitat as described previously.

Decreased streamflows in Meadow Creek, Sugar Creek, the EFSFSR, and the SFSR during construction through closure (MYs -1 to 20) (see BA Table 4.1-35, Stantec 2024), and in Sugar Creek and downstream reaches during post-closure (MYs 20+) will cause a reduction in food availability and suitable habitat, which could increase competition for both habitat and available food/forage (Appendix G).

Proposed ground-disturbing activities could affect instream sediment levels. Increases in fine sediment deposition within stream channels have the potential to decrease spawning gravel suitability and decrease benthic invertebrate production within gravel riffles, potentially impacting spawning/incubation and rearing/feeding life stages of Chinook salmon and steelhead. However, spawning substrates are generally FA in action area streams, and GRAIP-Lite modeling suggests that the amount of sediment delivered to action area streams will decrease in comparison to the baseline condition. With application of proposed sediment control BMPs, the deposition of sediment in action area waters is predicted to be measurable but not severe. Therefore, the deposition of sediment is not expected to degrade the food/forage PBF in the temporary or short-term.

As discussed previously, proposed revegetation efforts will work to begin restoration of riparian vegetation as soon as practicable, beginning as early as MY 1. This considered, revegetation efforts should ensure that effects to forage realized by alteration of riparian vegetation are expected to be temporary and short-term, and not expected to affect the food/forage PBF in the long term.

Habitat alteration will also decrease forage available for juvenile salmonids. Re-routing portions of Midnight Creek and Hennessey Creek in MY -1 will dewater short stream segments locally affecting forage drift to downstream habitats. However, re-routing the EFSFSR through the tunnel will result in a larger, but short-term loss of about a mile of aquatic forage production through MY 11. Re-routing the channel around proposed mining efforts and piping upper Meadow Creek into a constructed ditch with a low flow pipe will also eliminate or substantially reduce forage drift. The majority of these effects from work in Meadow Creek will be realized only from MY 1 to 3, while habitat is restored in lower Meadow Creek. The constructed ditch will most likely become an intermittent channel over rock at project completion, and macroinvertebrate diversity will continue to be locally reduced. This will adversely affect forage drift to downstream habitats (Meadow Creek and the EFSFSR) up until around MY 23, following introduction of flows in the newly created Meadow Creek channel over the TSF.

## **2.10.2. Effects to Critical Habitat – Lemhi**

The Lemhi Restoration component of the Project aims to enhance habitat for ESA-listed fish species at critical life stages and restore natural stream processes for long-term habitat diversity. Specific objectives include increasing habitat quality and complexity, reducing channel W:D, enhancing floodplain connectivity, improving instream structure and velocity, increasing pool quantity and complexity, facilitating surface/groundwater interchange for temperature moderation, providing instream cover for fish, and establishing a riparian corridor for shade, cover, and bank stability. Project elements include developing a multi-threaded channel network, installing woody material for complexity, adding floodplain roughness structures, increasing floodplain activation frequency, and revegetating riparian zones with native species (see Appendix D). These efforts aim to bolster fish population abundance, productivity, and spatial structure.

Critical habitat within the action area has an associated combination of PBFs essential for supporting freshwater spawning, rearing, and migration of the Lemhi River spring/summer Chinook and steelhead populations. The action area provides very limited spawning habitat for both Chinook salmon and steelhead due to current and historical anthropogenic impacts. The project is funded with the intent of improving habitat conditions in the Lemhi River basin, which will in turn move the affected populations toward recovery. The project will enhance habitat conditions across approximately 7,000 ft. of the Lemhi River channel (RM 42.63 to 41.32), restoring floodplain access, and revegetating the floodplain within the project area. The creation of approximately 12,426 ft. (2.35 mi.) of perennial and non-perennial side channel and another 4,965 ft. (0.94 mi.) on non-perennial tertiary side-channel will locally improve spawning and rearing habitat quantity and quality and reestablish natural channel processes. Ultimately, the project will improve the conservation value of the action area's critical habitat. During construction, minor and temporary effects to critical habitat are likely to occur. These effects are primarily related to increased turbidity (water quality) and associated sedimentation of spawning gravels.

# **Water Quality**

The water quality PBF will likely be affected by short-term turbidity pulses or plumes occurring during installation of temporary cofferdams and bypass channels and when introducing water to newly constructed channel segments. In efforts to control turbidity, no work will occur in flowing water. Project staging will allow the new off-channel habitat to be created first, before introduction of flowing water. Once the side channel complex is complete, channels will be 'prewashed' into a reach equipped with sediment control structures. The Lemhi River bank will then be breached, and flow gradually routed into the side channel complex. A cofferdam will be installed to block the flow in the mainstem so work in the mainstem can also occur in the dry. Once work is complete, the channel will again be 'pre-washed' before re-watering the Lemhi River. Design criteria (e.g., in-water work window, site dewatering, pumping turbid water to the floodplain to filter rather than discharging to fish-bearing channels, pre-washing channels, staged rewatering, and erosion control BMPs) are anticipated to effectively minimize the amount of sediment delivered to and suspended in the Lemhi River.

In addition to undertaking sediment control measures, Perpetua will monitor turbidity levels during re-watering efforts. Monitoring will include collecting background turbidity samples 100 feet upstream of the disturbance area. Depending upon the width of the channel, recordings will be made, 50, 100, or 200 feet downstream from the disturbance, monitored every 4 hours while work is being conducted. Should exceedances occur for more than two consecutive monitoring intervals (after 8 hours), the activity will stop until the turbidity level returns to background, and OSC will be notified. If turbidity controls (cofferdams, wattles, fencing, etc.) are determined ineffective, crews will be mobilized to modify, as necessary.

Temporary increases in turbidity are most likely to occur when adding water to newly constructed off-channel habitat, installation of the cofferdam on the Lemhi River, and when rewatering the Lemhi River channel (3 defined plumes or pulses). Based on review of the literature and the proposed sediment control measures, NMFS expects: (1) resulting sediment plumes will not exceed 100 NTUs; (2) should not extend farther than 1,000 feet downstream; (3) should dissipate within a few minutes to hours; and (4) progressively diminish as they progress downstream (Casselli et al. 2000; Jakober 2002; USFS 2005). In addition to effects remaining within 1,000 feet of the disturbance, suspended sediment levels are likely to quickly return to

background considering the expected small volume of sediment likely to be introduced and suspended (Jakober 2002; Casselli et al. 2000).

Turbidity pulses are expected to be infrequent, and short-lived (a few hours), before returning to background levels. These levels may temporarily cause up to 1,000 feet of the Lemhi River to be temporarily less suitable for fish. Because the Lemhi River is relatively large, the plumes are likely to be confined to one bank and many exposed fish will be able to easily move to adjacent non-turbid habitats, thereby avoiding exposure. Although these levels may be sufficient to cause minor behavioral modifications or an increase in foraging rates for some fish, these effects are expected to be minor and temporary in nature, and will not degrade the water quality PBF.

The use of heavy machinery adjacent to the stream channel increases the risk for the potential of an accidental spill of fuel, lubricants, antifreeze, hydraulic fluid, or similar contaminant into the riparian zone, or directly into the water where they could adversely affect the water quality PBF. However, all equipment performing work will workprimarily from streambanks and within dewatered stream channels, significantly reducing the likelihood of toxic materials entering live water. In addition, all equipment operated within 150 feet of any waterbody will be inspected daily for leaks and, if necessary, repaired before leaving the staging and refueling areas. A SPCC plan will be developed for the project, and all equipment will have spill containment kits onsite.

Considering the described EDFs, it is unlikely that fuel, lubricants, antifreeze, hydraulic fluid, or similar contaminant will be present on-site or spilled in volumes or concentrations large enough to harm water quality in or downstream from the project site. NMFS believes that fuel spill and equipment leak contingencies and preventions described in the proposed action are sufficient to effectively minimize the risk of negative impacts to water quality from toxic contamination; therefore, potential for adverse effects from chemical contamination is unlikely to occur to this PBF.

## **Substrate/Spawning Gravel**

Temporary pulses of sediment and turbidity plumes are expected to cause small increases in downstream sediment deposition (increased surface fines), negatively affecting substrate in the short term. Fine, re-deposited sediments have the potential to adversely affect primary and secondary productivity (Spence et al. 1996), and reduce incubation success (Young and Hubert 1991; Henley et al. 2000; Wu 2000) and cover for juvenile salmonids (Bjornn and Reiser 1991). As described above, project design and conservation measures (e.g., work will occur in dewatered work areas, during low flow periods, pre-washing and slowly re-watering channels) should effectively minimize the amount of sediment generated and delivered to action area waters. Additionally, all disturbed sites will be replanted with native vegetation to reduce any potential long-term sediment inputs. This considered, only minor amounts of fine sediment deposition are expected as a result of the proposed action. In the long term, projected increases in pool to riffle ratio, increased habitat complexity, and access to new off-channel habitat are expected to significantly improve the quantity and quality of spawning substrate locally available. This would be a localized, long-term improvement in the conservation value of this PBF.

# **Floodplain Connectivity**

Historic channel straightening and grazing has occurred throughout the project reach. These impacts have led to loss of floodplain connectivity. Reestablishing connectivity along 7,000 ft. of the Lemhi River will greatly improve the ability of the river to interact with its floodplain in the action area, locally improving the conservation value of this PBF.

### **Space, Cover, and Shelter**

Channel confinement and straightening, bank erosion, lack of riparian vegetation, and sedimentation have contributed to low pool frequencies and decreased pool quality (e.g., lack of cover, complexity, and depth) in this stream segment. The stream channel design increases access to approximately 3.29 mi. of new off-channel perennial and seasonal habitat. NMFS expects a significant corresponding improvement in the quality and quantity of fish cover and shelter within the subject reach. Overall the results of the action are a significant, long-term, localized improvement in the conservation value of this PBF.

## **Riparian Vegetation, Water Temperature, Food/Forage**

The project area is in a disturbed location with riparian vegetative cover lacking in much of the stream reach. Existing streambank vegetation shall be preserved and protected to the extent practical, with the contractor only removing trees and shrubs necessary for the execution of the work. No tree or shrub shall be removed unless approved by the contracting officer. The contractor shall not disturb the roots of woody vegetation in this area during project excavations to the extent practical. Riparian vegetation will be replanted prior to or at the beginning of the first growing season following construction. Reestablishment of vegetation will be achieved in disturbed areas to at least 70 percent of the pre-project conditions within three years. The W:D in the mainstem will be reduced, which will locally provide long-term improvements to water temperature by creating a deeper, narrower stream channel, and creating a condition that is more conducive to stream shading over a larger percentage of the channel width.

Naturally functioning riparian vegetation and improved floodplain connectivity will increase the amount of available habitat and provide better habitat complexity. The planting, seeding, and transplanting of riparian vegetation will be a critical component of the long-term restoration process of the project area. Improved riparian conditions will help stabilize streambanks, decrease sediment loading, reduce stream temperatures, and improve habitat complexity. The planting of riparian vegetation could result in a localized long-term benefit to streamside shade, bank stability, cover for fish, water temperatures, and habitat for macroinvertebrates in the project area. Therefore, this project is anticipated to improve the conservation value of these PBFs in the action area.

#### **Fish Passage**

There will be no physical barriers created or removed other than cofferdams to temporarily block or redirect flows depending on the construction stage. Low flows may result in temporary passage barriers during construction; however, fish will not have access to the construction area

during construction activities. Once the side channel habitat is constructed, the Lemhi River bank will be breached to redirect flow away from the mainstem and into the new channels.

## **2.10.3. Summary of Effects to Designated Critical Habitat**

The SGP will affect SR Basin steelhead DCH for the Salmon River MPG, specifically the SFSR population; while effects of the Lemhi Restoration portion of the project will also affect the same MPG, but the Lemhi River population. Effects to steelhead DCH are most likely to occur in the Lemhi River, SFSR, the EFSFSR (downstream from YPP), Johnson Creek, Cabin Creek, and Sugar Creek.

The SGP will affect SR Spring/summer Chinook salmon DCH for the SFSR MPG, specifically the SFSR and EFSFSR populations; while the Lemhi Restoration portion of the proposed action will affect DCH for the Upper Salmon MPG, specifically the Lemhi River population. Effects to Chinook salmon DCH are most likely to occur in the Lemhi River, SFSR, EFSFSR, Johnson Creek, Cabin Creek, Sugar Creek, lower reaches of Burntlog Creek, Trapper Creek, Riordan Creek, and EFMC.

Adverse effects will be primarily related to potential effects to: (1) water quality (turbidity, contaminants, or temperature); (2) water quantity; (3) riparian vegetation; (4) natural cover/space; (5) food and forage (due to chemical contaminants and water quantity effects); (6) substrate/spawning gravels; (7) fish passage; and (8) floodplain connectivity. Modification of these PBFs may affect freshwater spawning, incubation, rearing, or migration in the action area.

Sediment generating activities will include frequent, sporadic turbidity effects to water quality across action area streams, with those increases occurring most frequently occurring during runoff events and resulting in localized deposition of fine sediment in action area stream channels. These effects are likely to affect DCH from the beginning of construction through closure and reclamation. Although turbidity increases are likely to affect water quality to the point that they periodically affect fish behavior (particularly during runoff events), they are unlikely to reach levels severe enough or long enough to affect the long-term conservation value of the Water Quality PBF.

Potential effects could be realized to the water quality PBFs for both steelhead and Chinook should an accidental spill of contaminants occur. However, the risk of a hazardous material spill is low, and proposed transport EDFs should effectively minimize the potential for accidental spills and their resulting effects to occur.

Water quality will be affected by both point and nonpoint sources of pollution during construction, operations, and closure. Stormwater discharges from site facilities and roadways will degrade the water quality PBF during and immediately following storm events, and this disturbance is expected to continue to occur sporadically throughout the construction and operations period. Because site facilities will be removed and mine-related traffic will be substantially reduced following closure, this impact will not persist in perpetuity. The sanitary wastewater treatment facility discharge is also expected to reduce the water quality PBF in the EFSFSR during construction and operations.

Removal of the historical, unlined mine waste disposal material will have an early, positive impact on water quality; however, placement of permanent mine facilities (e.g., TSF embankment and buttress, and waste rock backfill into the HFP) and discharge of treated mine contact water will negatively impact water quality. The net effect of these actions during operations will be an overall reduction in antimony, arsenic, and copper concentrations in Meadow Creek and the EFSFSR during operations. However, during operations concentrations of these contaminants will remain at levels that will reduce the ability of the water quality PBF to support incubation and early life stages (in the EFSFSR when maximum predicted concentrations of antimony and arsenic are realized for extended periods of time), juvenile rearing, and adult/juvenile migration. Total mercury contributions from West End Creek to Sugar Creek will increase by at least an order of magnitude. During early closure, the water quality PBF will continue to be impacted by permanent mine facilities (e.g., TSF embankment and buttress and waste rock backfills) and discharge of treated mine contact water. Arsenic, antimony, and copper concentrations are expected to continue to be lower than existing conditions in Meadow Creek and the EFSFSR. Mercury concentrations will decrease in Meadow Creek and the EFSFSR; however, concentrations will remain slightly elevated in Sugar Creek. Mercury concentrations in lower West End Creek will decrease substantially as the WEP Lake fills. Although concentrations are decreased, they will remain above levels that can contribute to harmful bioaccumulation of mercury. Ultimately, individual contaminants and contaminant mixtures are expected to continue to negatively impact the ability of the water quality PBF to support spawning and incubation, juvenile rearing, and adult/juvenile migration. During late closure, the water quality PBF will continue to be impacted by permanent mine facilities. Mine contact water will no longer need treatment because any seepage from the TSF and TSF embankment and buttress is predicted to meet water quality criteria. Arsenic concentrations will continue to decrease in Meadow Creek and will be below concentrations associated with sublethal effects from waterborne exposures. Similarly, arsenic concentrations will continue to decrease in the EFSFSR; however, maximum predicted concentrations are likely to cause some mortality of embryos. In addition, arsenic concentrations will contribute to sublethal effects to rearing juveniles feeding as a result of dietary exposure. Antimony concentrations will either decrease slightly or remain the same, and will be below levels associated with lethal or sublethal effects. Copper concentrations will decrease, though are still associated with potential sublethal olfaction effects. Mercury concentrations will generally decrease in stream reaches upstream of Sugar Creek. Concentrations will remain above 2 ng/L, and depending on methylation potential, could accumulate in fish tissue and result in some sublethal effects such as reduced growth or altered behavior. Overall, the ability of the water quality PBF in Meadow Creek, EFSFSR, and Sugar Creek to support incubating embryos, rearing juvenile salmonids, and migrating adult and juvenile salmonids will be negatively impacted by the proposed action.

Water temperatures are not expected to be negatively impacted in streams near or overlapped by the transportation or transmission line corridors. Similarly, temperatures will not be negatively impacted in Sugar Creek or EFMC to a degree that would impact the ability of stream temperatures to support rearing (both streams), or migration and spawning (Sugar Creek). At the mine site, water temperatures in the EFSFSR and Meadow Creek will be impacted as a result of channel relocation and reconstruction, EFSFSR diversion tunnel construction, riparian vegetation removal, and point source discharges. During the construction phase, stream temperatures are expected to be similar to baseline conditions and the ability of the designated critical habitats to

support spawning, rearing, and migration will not be reduced. During operations, closure, and post-closure, stream reaches will respond differently to SGP activities as summarized below.

Stream temperatures in Meadow Creek upstream of the EFMC as a result of the SGP are predicted to improve relative to baseline, during operations and will be warmer than baseline conditions during early and late closure. Stream temperatures in Meadow Creek downstream of the EFMC will improve relative to baseline conditions during all years of project implementation. Yet, the habitat's ability to support Chinook salmon juvenile rearing will continue to be diminished during portions of the warmer summer months for a period of 25 years. Temperatures are predicted to increase in between MYs 23 and 27; particularly in the reach upstream of the EFMC. This reach will be further impacted by thermal loading from the wastewater treatment plant discharge. During this 5-year period, the maximum weekly maximum summer temperature in this reach is predicted to be 20.8°C, which will limit juvenile rearing as well as adult migration and spawning (particularly during the early weeks of the spawning season). During the fall, the maximum weekly maximum temperature is predicted to be 16°C which will reduce survival of incubating embryos. Considering the rapid decline of stream temperatures in the fall months, late spawners utilizing this habitat will have more success. While there will be a long-term reduction in the temperature PBF of Chinook salmon critical habitat in Meadow Creek, these reaches represent a small fraction of the designated critical habitat for Chinook salmon in the action area. Temperatures in the EFSFSR between Meadow Creek and the YPP are expected to improve relative to existing baseline conditions during all years of project implementation, and will be less than 18°C. The predicted temperatures in this reach will fully support adult Chinook salmon migration and holding. Predicted temperatures will remain above 16<sup>o</sup>C, diminishing the value of the temperature PBF for Chinook salmon juvenile rearing, particularly in MYs 6 and 23 through 27. Maximum fall temperatures will be 13.5°C, which is slightly above optimal levels for Chinook salmon spawning and incubation at least for the few weeks of the season in MY 6. Below the YPP, stream temperatures in the EFSFSR are expected to fully support salmon and steelhead adult migration (temperatures less than 18°C), spawning (temperatures less than 13°C), and incubation (temperatures less than 13°C). Temperatures in this reach are expected to be slightly above 16°C during the summer, which will slightly decrease in the ability of stream temperatures to support salmon and steelhead juvenile rearing during MY 6. That is, some individuals are likely to experience sublethal effects as a result of increased stream temperatures. Temperatures in Sugar Creek and the EFSFSR below Sugar Creek are expected to increase. While the conservation value of the temperature PBF for rearing will be diminished through MY 32 and some rearing juveniles may experience sublethal effects, the habitat will support adult migration, spawning, incubation, and rearing for Chinook salmon and steelhead.

Climate change with or without implementation of the SGP will make it more difficult to maintain temperatures that are protective of all anadromous salmonid life stages within the action area. Climate change is expected to exacerbate the effects of the proposed action on stream temperatures directly as well as indirectly through impacts to air temperature, streamflow, and vegetative recovery.

Physical habitat conditions, represented by the spawning gravel, passage, and space/cover/shelter PBFs, will be affected by the proposed actions. Modeling suggests that the amount of sediment
delivered to action area streams will be effectively minimized and decrease in comparison to the baseline condition – considering proposed sediment control BMPs. Therefore, the deposition of sediment in action area waters is predicted to be measurable but not severe, and is not expected to degrade this PBF in the temporary or short term. Following closure of the mine, obliteration of access roads, and restoration efforts addressing chronic sediment delivery in the EFMC, the proposed action is expected to result in a substantial decrease in annual sediment delivery into Meadow Creek and the EFSFSR, and may benefit this PBF in the long term.

Passage, and access to additional space, cover, and shelter, will be improved for steelhead DCH in East Fork Burntlog Creek when the Burntlog Route stream crossings is retrofitted to ensure season-long fish passage. Volitional passage will be restored for Chinook DCH upstream of the YPP, with Chinook able to access an additional 12.2 miles (19.65 km) usable habitat, of which 5.5 miles is identified as having intrinsic potential for Chinook salmon. These increases in Chinook space, cover, and shelter will be realized as soon as tunnel construction is complete in MY -1. Restored and enhanced stream channels in Meadow Creek and the EFSFSR upstream and downstream from the YPP (MY -1), and across the YPP backfill (MY -11), will increase habitat complexity in stream reaches in the project area, resulting in long-term improvements to the space, cover and shelter PBFs for Chinook upstream and downstream from the YPP, and downstream of the YPP for steelhead DCH. Passage will not be affected for steelhead or Chinook in the Lemhi River portion of the action, although both species will gain additional space, cover, and shelter as a result of the proposed restoration project in the Lemhi River basin.

The floodplain connectivity PBF is specific to rearing steelhead. Reconnecting the EFSFSR to its floodplain across the YPP backfill by MY -11, will locally improve floodplain connectivity for the portion of this stream reach that is formally steelhead DCH (EFSFSR upstream to the base of the YPP cascade). Floodplain connectivity will be restored across a 7,000 linear ft. section of the Lemhi River, locally restoring this steelhead PBF by MY -1.

Riparian vegetation has been heavily impacted by historic mining and streamside roads (and wildfire to a lesser extent) in the EFSFSR and Sugar Creek. This PBF has also been heavily impacted by historic wildfire and streamside roads along Johnson Creek and Burntlog Creek. Localized riparian vegetation may be cleared or trimmed to accommodate the new access roads for mining and the powerline. However, riparian habitat restoration and enhancement will occur concurrently with mining, beginning early in the mining process, during construction, and would continue through site closure. The SGP will have temporary and short-term negative effects to this Chinook salmon PBF, but will result in a long-term improvement as the Burntlog Road is obliterated and revegetation efforts at the mine site take hold.

Riparian vegetation in the Lemhi River channel has also been heavily impacted, but impacted by channelization, grazing, development, and agriculture. The Lemhi portion of the project will reestablish connectivity with the river's floodplain, and the revegetation plan will locally restore riparian habitat in the long term.

Proposed ground-disturbing activities could affect instream sediment levels, which could in turn affect benthic invertebrate production. However, spawning substrates are generally FA in action area streams, and modeling suggests that the amount of sediment delivered to action area streams is predicted to be measurable but not severe. Therefore, the deposition of sediment in action area waters is not expected to degrade the food/forage PBF in the temporary or short-term.

The food/forage PBF can also be affected by changes in water quality (e.g., stream temperature, dissolved oxygen, chemical contamination), water quantity, sedimentation, and clearing of riparian vegetation. Reduction in stream flow will reduce food availability in all affected stream reaches during construction through closure, and in Sugar Creek and downstream reaches for approximately 135 years post closure. Increased contamination is not expected to reduce the quantity or diversity of available prey; however, the quality of prey items will be reduced at the mine site and in downstream reaches of the EFSFSR. We expect concentrations of arsenic in prey to decrease as a result of project implementation; however, mercury concentrations are expected to slightly increase in prey items.

## **2.10.4. Effects to Species – SFSR**

With the exception of disturbance and fish sampling effects, all of the proposed action's potential effects to species are directly related to fish responses to the previously described effects on DCH. For this reason, this section incorporates the previous effects to DCH by reference given individual fish and fish populations' response to the habitat in the manner described above. Doing so eliminates redundancy while enabling more clear focus on the effects to the species under consideration. In some cases, we determined stressors that have potential to occur but are unlikely to materialize given the type of activities, the location of activities relative to aquatic habitats, the effectiveness of proposed avoidance and mitigation measures contained in the action, or a combination of these factors. For these instances, the low likelihood of the stressor being generated effectively reduces exposure of ESA-listed species to the stressor itself and thus there is no direct effect anticipated. For these stressors, we do not discuss the mechanisms of effect to the species because exposure is considered unlikely. Changes in water quantity and water quality are both anticipated to occur. Therefore, the species will be exposed to these habitat changes and respond. The following sections discuss these impacts from such exposure and responses.

#### **Water Quality**

#### *2.10.4.1.1. Chemical Contamination*

Effects to ESA-listed Chinook salmon and steelhead could be affected through chemical contamination, which could occur: (1) should a spill occur onsite or in the transportation corridor; (2) when construction equipment is working within or adjacent to the stream channel; (3) from herbicide treatment; (4) as a result of effluent discharge; or (5) through groundwater contamination.

*Hazardous Materials Spill Risk Analysis.* Potential for transportation-related spills of substances associated with construction, operation, and closure of the SGP were discussed in Section 2.10.1.1. That discussion found a low likelihood for diesel, or other chemical spills, to affect critical habitats in the action area during transportation to and from the mine site. Similarly, in Section 2.10.1.2, we concluded the SGP is unlikely to lead to onsite chemical spills that reach designated critical habitats. These conclusions were reached after considering: (1) the

transportation routes used; (2) accident probability; (3) historical accident rates; (4) location of mine facilities relative to occupied habitat; (5) onsite access controls; (6) site collection and handling requirements; and (7) anticipated effectiveness of transport EDFs, the SPCC Plan, and the Hazardous Materials Handling and Emergency Response Plan. These measures outline proper material inventory, packaging, transportation, storage, use, disposal, and clean up procedures, which, when implemented properly, collectively reduce the potential for spilled materials reaching waters utilized by ESA-listed salmon and steelhead.

*Herbicide Application.* As previously described, noxious weed and invasive species plant control will be conducted according to methods approved in the 2020 Programmatic Activities Biological Assessment (PNF 2020) and NMFS opinion (NMFS 2020) on this action. The potential effects of weed treatment when applied as proposed have been described in detail in Section 2.5.4.1 of NMFS' opinion (NMFS 2020), incorporated here by reference and summarized below.

Application of herbicides can contaminate surface waters, harming individual fish. Risks to salmon and steelhead from herbicides are likely to occur primarily through the direct toxicological effects of herbicides and adjuvants on the fish, rather than indirectly through physical changes in fish habitat or effects on aquatic vegetation or prey species. Weed treatments typically occur between April and November, depending on elevation. During this period, all life stages of Chinook salmon and steelhead could potentially be exposed to herbicides, including incubating eggs, rearing juveniles, and migrating/holding adults. Herbicides will not be directly applied to water, with only spot spraying or hand application techniques allowed within 100 ft. of streams, making direct application to waters unlikely.

Herbicides (including the active ingredient, inert ingredients, and adjuvants) can potentially harm fish directly or indirectly. Herbicides can directly affect fish by killing them outright or causing sublethal changes in behavior or physiology. Indirect effects to fish may occur when herbicides alter the aquatic environment by way of causing changes in cover, shade, runoff, and prey availability (Scholz et al. 2005).

Herbicide exposure may directly result in one or more of the toxicological endpoints identified below. These endpoints are generally considered to be important for the fitness of salmonids, and include:

- Direct mortality at any life history stage;
- An increase or decrease in growth;
- Changes in reproductive behavior;
- A reduction in the number of eggs produced, fertilized, or hatched;
- Developmental abnormalities, including behavioral deficits or physical deformities;
- Reduced ability to osmoregulate or adapt to salinity gradients;
- Reduced ability to tolerate shifts in other environmental variables (e.g., temperature or increased stress);
- An increased susceptibility to disease;
- An increased susceptibility to predation; and,
- Changes in migratory behavior.

Herbicides applied by the applicant are not expected to reach streams in concentrations that kill fish; but, weed treatment may result in concentrations of sufficient magnitude to elicit short-term sublethal effects. However, implementation of weed control PDFs, including BMPs designed to minimize the amounts of herbicide getting to active water, will effectively reduce the risk of chemical contamination associated with weed treatment activities.

Application of many herbicides proposed for use could potentially drift to waterways or leach into soils, contaminate groundwater, and eventually show up in streams where ESA-listed fish spawn and rear. This risk depends on a number of variables, including but not limited to the rate of application, concurrent precipitation, herbicide degradation, solubility, and distance to water. Again, identified design criteria minimize risks of these occurrences. Although unlikely to occur, accidental spill of herbicides directly to waterways could result in conditions toxic to salmonid fish species and result in harm/death of ESA-listed fish.

In spite of the PDFs, herbicides (and adjuvants) cannot be kept entirely out of the water. Although direct lethal effects are not expected, the types of chemicals used are ones that can be capable of causing harmful sublethal effects. It is possible that individual or smaller groups of fish could be exposed and experience sublethal effects, potentially experiencing changes in behavior, reduced growth, survival, etc. However, herbicide applications will not occur over large contiguous areas (less than 500 acres of chemical application per year in the entire SFSR subbasin) (PNF 2020), and most of the action area will not be subjected to spraying in any given year. Considering the low level of effects that may occur, coupled with the very small impact area, we do not expect there to be any realized reductions in the abundance or productivity of any potentially affected populations.

*Surface and Groundwater Quality Analysis.* Section 2.10.1.1 describes in detail how the SGP is predicted to alter water quality in the action area. Appendix F and Section 2.10.1.1 summarize the best available information regarding the toxicity of contaminants that present the greatest risk of reducing individual fitness. The primary pathways of effect for SR Spring/summer Chinook salmon and SR Basin steelhead is through waterborne exposure to contaminants during all life stages and through dietary exposure as juveniles are feeding as they rear in and migrate through the action area.

*Waterborne Exposures.* As previously described, historic sources of contamination will be removed and new sources chemical contaminants will be introduced to action area streams as a result of stormwater runoff, point source discharges from a sanitary wastewater treatment plant and mine contact water treatment plant, and nonpoint source contributions from mine facilities. Juvenile fish are generally the most sensitive to contaminants through waterborne exposures, although in some cases, early life stages (incubation through early exogenous feeding) may be even more sensitive. All life stages of fish are expected to be exposed to contaminants originating in stormwater runoff from roadways and maintenance and housing facilities as well as to contaminants in sanitary wastewater treatment plant effluent. These exposures are not expected to cause outright lethality; however, some sublethal effects (e.g., altered behavior) from exposures to individual contaminants or contaminant mixtures may occur. There is substantial uncertainty about the degree to which these sublethal effects lead to latent mortality.

Adult SR Spring/summer Chinook salmon and SR Basin steelhead are expected to be exposed to arsenic, mercury, antimony, copper, and other contaminants that leach from mine facilities as they migrate through and hold in the action area (access to habitat upstream of the YPP is expected to occur by MY -1). Predicted concentrations of these contaminants are not expected to reach levels that will kill adult salmonids or reduce spawning activities. Copper is the only contaminant that may affect adult life stages at the mine site. Copper concentrations may reach levels that cause adult salmonids to avoid habitats; however, we consider there to be low risk of this type of response. This is because copper concentrations are expected to decrease relative to existing conditions and adult SR spring/summer Chinook salmon redds have been documented in the EFSFSR, Sugar Creek, and Meadow Creek (when adults are outplanted) in recent years, suggesting copper concentrations are not preventing, and will not prevent, seeding of available spawning habitat. Although SR Basin steelhead redd surveys have not been conducted, we assume this species would respond similarly.

Juvenile life stages are generally more sensitive than adult life stages. As such, there is greater risk of adverse responses for juvenile exposures to arsenic, antimony, and copper. There is a very low risk of mortality of some developing embryos or newly emerged fish from exposures to concentrations of antimony and arsenic in some reaches during project implementation. Maximum predicted antimony concentrations in the EFSFSR below the YPP during operations are near concentrations associated with low levels (1 percent) of mortality in incubating embryos (Birge et al. 1978). If redds are exposed to this maximum predicted concentration for extended periods of time, then it is possible some low levels of mortality could occur. Similarly, predicted arsenic concentrations in the EFSFSR are near concentrations associated with low levels of mortality in incubating embryos (Birge et al. 1978). If redds are exposed to this maximum predicted concentration for extended periods of time, then it is possible some low levels of mortality could occur. This risk exists in the EFSFSR upstream of the YPP only during operations; however, this risk exists in the EFSFSR downstream of the YPP through late closure. Juvenile fish are not expected to experience any other lethal or sublethal effects as a result of waterborne exposures to antimony or arsenic at the predicted concentrations.

There is some risk of juvenile fish experiencing reduced olfactory function as a result of exposure to the maximum predicted copper concentrations. Such reductions in olfaction may lead to death through impairment of the ability of salmonids to avoid predators or find prey. Individual fish rearing in Meadow Creek, EFSFSR, and Sugar Creek are at risk of experiencing this type of response. Because copper concentrations are expected to decrease or remain essentially the same relative to existing conditions in all stream reaches during all mine phases, the proposed action is not expected to result in reduced abundance, productivity or spatial structure.

Considering evidence of successful incubation and early rearing in Meadow Creek and the EFSFSR under current water quality conditions, and considering the SGP will substantially reduce antimony and arsenic concentrations in these streams and copper concentrations will remain substantively the same, we do not believe these potential low levels of mortality or sublethal effects will reduce abundance or productivity in the population.

*Dietary Exposure.* Juvenile Chinook salmon and steelhead primarily feed on macroinvertebrates and are at risk of accumulating arsenic and mercury. Adult salmon and steelhead are not expected to be feeding in the project area, therefore, the risk of arsenic or mercury bioaccumulation in adults from on-site exposure is negligible.

As described in Section 2.10.1.1, dietary concentrations of arsenic of around 20 mg/kg dw are associated with reduced growth, organ damage, and other physiological effects (Cockell et al. 1991; Hansen et al. 2004; Erickson et al. 2010, 2011). These sublethal effects are thought to occur when arsenic concentrations in prey are approximately 20 mg/kg dw or greater, which can accumulate when water levels are about 10 µg/L. Average arsenic concentrations are predicted to substantially decrease and average concentrations are expected to generally be below the 10  $\mu$ g/L threshold thought to be associated with dietary toxicity. Yet, predicted maximum concentrations will occasionally exceed the 10  $\mu$ g/L (i.e., for some years at the start of operations and during early closure). As such, there is some risk of individual fish experiencing sublethal effects as a result of dietary arsenic exposure.

As described in Section 2.10.1.1, the proposed action is expected to cause incremental increases in total mercury in the water column of stream inhabited by salmon and steelhead. These increases could lead to greater bioaccumulation of mercury in fish tissues. Additionally, an incremental increase in organic carbon content in the EFSFSR due to sanitary wastewater effluent could increase in methylation potential in the EFSFSR. Bioaccumulation of mercury in fish tissue is not expected to reach lethal levels; however, some fish may experience sublethal effects. The degree to which these sublethal effects may manifest to reduced survival is unknown.

Baseline data (presence of redds, juvenile Chinook salmon, and juvenile steelhead) in the EFSFSR, Sugar Creek, and Meadow Creek (Chinook salmon only due to outplanting efforts) indicates spawning, rearing, and migration currently occurs in streams with elevated arsenic and detectable, and in some cases elevated, concentrations of total mercury. While arsenic and mercury body burdens may continue to cause sublethal effects in some individual juvenile fish as a result of project implementation, we do not have adequate information to be reasonably certain that these effects will result in population-level effects.

#### *2.10.4.1.2. Water Temperature.*

Water temperatures is an important factor affecting the survival of each Chinook salmon and steelhead life stage. Sublethal water temperatures may influence behavior, respiration, growth rates, metabolism, and ecological interactions such as predation, competition, or disease, migration timing, and egg viability. Elevated water temperatures may also trigger avoidance of areas in which may result in crowding of other rearing or holding habitat that offers more suitable temperatures. A number of factors influence how individual fish respond to elevated temperatures, including thermal variation, access to thermal refugia, prey availability, and presence of other stressors (such as low dissolved oxygen). Streams in the mine site area exhibit significant seasonal and diurnal variations, and for mobile life stages (i.e., adults and juveniles), if MWMTs are above the optimal thresholds, fish may seek suitable habitat nearby (e.g., EFMC or EFSFSR). Through stream restoration and enhancement actions, stream cover and instream structures may also provide thermal refugia that does not currently exist. A summary of adverse

effects that each life stage may experience when exposed to elevated temperatures is provided in Table 61.

Section 2.10.1.1 details the temperatures changes that may occur as a result of SGP implementation and the effects the predicted stream temperatures are likely to have on the various life stages. As previously described, temperature impacts in parts of the action area that are not within the mine site area are not expected to be great enough to elicit lethal or sublethal responses from individuals. Similarly predicted temperature impacts in Sugar Creek and the EFSFSR below Sugar Creek are not expected to cause any lethal or sublethal responses in exposed individuals. Chinook salmon and steelhead will have volitional access to habitat upstream of the YPP as soon as the EFSFSR diversion tunnel is completed, which is anticipated to occur in MY -1. As such, our analysis considered exposure and response of all life stages of salmon and steelhead to these habitats.

Steelhead adult migration and spawning life history stages are not expected to be adversely affected by the SGP given they occupy the habitat during the spring when flows are higher and temperatures are lower. Incubating embryos could potentially be impacted by the proposed action if development is advanced due to elevated temperatures in June/July. Given the elevation of the mine site and the anticipated climate conditions at the site, the risk of this potential exposure pathway leading to effects that negatively impact individual fitness is quite low. Juvenile steelhead rearing at the mine site will be adversely affected by changes in stream temperature and their exposure and response is expected to be similar to that of Chinook salmon.

Chinook salmon migrate, spawn, and rear at the mine site, and all life stages are expected to be negatively affected in one or more reaches as a result of stream temperature changes cause by the SGP.

Individual salmon and steelhead are expected to respond to temperature changes in Meadow Creek and the EFSFSR (reaches between Meadow Creek and Sugar Creek) during mine operations and closure. The biological objectives underlying the stream designs for Meadow Creek reaches MC4, MC5, and MC6 are to support both Chinook salmon and steelhead spawning and rearing. Temperatures in Meadow Creek are expected to be reduced from existing conditions during operations and early closure. Although reduced, temperatures will exceed 16°C for many years. It is likely that some individuals will experience sublethal effects (e.g., reduced growth, increase risk of disease) during their rearing time period, particularly if macroinvertebrate drift is reduced from upstream sources. Warmer temperatures result in higher metabolic rates, and without adequate forage, fish may not grow sufficiently. Individual fish that fail to grow adequately in natal and rearing streams have less chance of successfully returning as adults (Mebane and Arthaud 2010).

<b>Life Stage</b>	<b>Temperature</b> $(^\circ C)$	<b>Biological Response</b>		
<b>Adult Migration</b>	$21 - 22$	Direct mortality (sustained exposures)		
		Migration blockage and delay		
	>20	Reduced swimming performance		
		Reduced individual survival		
	>18	Elevated disease risk		
		Reduced gamete viability		
		Overall reduction in migration fitness		
Spawning/Incubation	>18	Mortality		
		Development abnormalities		
	>13	Increased risk of embryo mortality ٠		
Juvenile Rearing	>23	Direct mortality		
	>20	Reduced competitive success		
		Increased risk of predation		
	>16	Elevated disease risk		
		Reduced growth		

**Table 61. Potential biological responses, by life stage, that may occur if individuals are exposed to elevated temperatures.**

Source: EPA 2003; EPA 2023

Temperatures in Meadow Creek above EFMC are expected to experience the greatest thermal change and may reach MWMT of 20.8°C during MY 27. It is possible that adult Chinook salmon, if they elect to remain in this reach for extended periods of time could experience prespawn mortality as a result of elevated temperatures coupled with other stressors (e.g., elevated contaminants) in the reach. In addition, gamete viability will likely be reduced. Embryos will also experience mortality as a result of redd exposure to elevated temperatures during the first few weeks of incubation. Juvenile fish rearing in this reach may experience reduced competitive success and may be at greater risk of predation from predators (e.g., bull trout) for MYs 23 through 27. Predicted stream temperatures may also reduce embryo survival, particularly at the early stages of spawning. Overall, temperature conditions in Meadow Creek are expected to support production of Chinook salmon and steelhead during SGP implementation; however, the contribution of salmon or steelhead to population productivity is expected to vary over time, with the lowest contributions occurring in MYs 23 through 27. The Meadow Creek reaches that may be used for spawning and rearing represent less than 0.1 percent of the total intrinsic potential habitat for the SFSR steelhead population and less than 1 percent of total intrinsic potential habitat for the EFSFSR Chinook salmon population. Considering the habitat represents a small fraction of the population production potential and considering only a few individuals are likely to experience lethal effects (e.g., Chinook salmon embryos; or delayed mortality of juvenile salmon or steelhead due to sublethal effects from exposure to elevated temperatures), we do not believe the productivity of the populations will be reduced to a degree that will preclude the populations from achieving their desired status.

The biological objectives underlying the stream restoration and enhancement designs for the EFSFSR are to support both steelhead (primary) and Chinook salmon (secondary) spawning and rearing. These reaches are expected to support adult migration, spawning, and rearing for both species. During MY 6, MWMT are expected to be slightly above that which is considered "optimal" for juvenile rearing. It is possible that individual fish may experience some limited

sublethal effects as a result of these slightly elevated temperatures. Sublethal impacts to a few individuals inhabiting this reach of the EFSFSR is not expected to alter the productivity of the SFSR steelhead or EFSFSR Chinook salmon populations.

#### **Suspended Sediment, Spawning Gravel/Substrate**

As previously described in the DCH effects analysis, action area streams could be affected through project generated turbidity and subsequent sediment deposition. However, the proposed action includes measures to reduce or avoid sediment delivery and turbidity impacts, including silt fences, straw waddles, graveling and applying dust control chemicals on roads, limiting equipment stream crossings, operating in dewatered work areas, etc.

Increased sediment delivery can cause turbidity pulses and lead to excessive deposition on the channel bottom. Elevated turbidity can cause lethal, sublethal, and behavioral effects in juvenile and adult salmonids; depending on the duration, frequency, and intensity of the exposure (Newcombe and Jensen 1996). Increased turbidity levels in the action area may result in temporary displacement of fish from preferred habitat or potentially sublethal effects such as gill flaring, coughing, avoidance, and increase in blood sugar levels (Bisson and Bilby 1982; Sigler et al. 1984; Berg and Northcote 1985; Servizi and Martens 1992). Literature reviewed in Rowe et al. (2003) indicated that NTU levels below 50 generally elicit only behavioral responses from salmonids, and Lloyd (1987) suggested that salmonids reacted negatively, by moving away, when turbidity reached 50 NTU. Although elevated turbidity levels may cause stress, Gregory and Northcote (1993) have shown that moderate levels of turbidity (35 to 150 NTU) can also accelerate foraging rates among juvenile Chinook salmon, likely because of reduced vulnerability to predators (camouflaging effect).

Sediments suspended in the water column reduce light penetration, increase water temperature, and modify water chemistry. Once in streams, fine sediment is transported downstream and is ultimately deposited in slow water areas and behind obstructions. Sediment deposition can locally alter fish habitat conditions through partly or completely filling pools, increasing the width to depth ratio of streams, and changing the distribution of pools, riffles, and glides. In particular, fine sediment has been shown to fill the interstitial spaces among larger streambed particles, which can eliminate the living space for various microorganisms, aquatic macroinvertebrates (i.e., prey items for juvenile salmon and steelhead), and juvenile fish (Bjornn and Reiser 1991).

As previously described, the proposed action will cause increased sediment delivery to streams from pre-construction through closure and restoration. The resultant turbidity plumes will be sufficient in magnitude and duration for fish to experience biologically meaningful behavioral changes or ill effects as previously described. It is likely that turbidity spikes, especially those associated with instream work will cause fish to find refuge away from the turbid water, which may expose them to predation. Fish unable to escape turbid waters may experience short-term behavioral changes described above. Turbidity plumes associated with instream work are anticipated to travel up to 1,000 feet downstream prior to dissipating to levels that are no longer harmful to aquatic species. These plumes are expected to be short-lived (lasting only a matter of minutes to hours), and no turbidity related injury or morality is expected to occur as a result of the project.

When sediment delivery exceeds the sediment transport capability of the stream, the amount of fine sediments will increase on and within stream substrates. Potential problems associated with excessive instream sediment have long been recognized for a variety of salmonid species and at all life stages, from possible suffocation and entrapment of incubating embryos (Coble 1961; Phillips et al. 1975; Hausle and Coble 1976; McCuddin 1977; Cederholm and Salo 1979; Peterson and Metcalfe 1981; Tagart 1984; Reiser and White 1988; Lisle and Lewis 1992), to loss of summer rearing and overwintering cover for juveniles (Bjornn et al. 1977; Hillman et al. 1987; Griffith and Smith 1993), to reduced availability of invertebrate food for resident adults (Tebo 1955; Cederholm and Lestelle 1974; Bjornn et al. 1977; Alexander and Hansen 1986). Salmonid populations are typically negatively correlated with the amount of fine sediment in stream substrate (Chapman and McCleod 1987).

The BRGI portion of the action could affect the substrate/spawning gravel PBFs through mobilization of fine sediments. However, as described above, proven erosion control measures are expected to effectively limit the amount of sediment delivered to action area streams. In addition to BMPs mentioned above in the critical habitat effects section, streams near drilling sites will be monitored during drilling, and drilling will cease if turbidity is detected; and, because drilling activities typically do not result in mobilization of sediment, sediment effects on Chinook salmon or steelhead due to the proposed drilling activities are expected to be minor and not expected to reach levels causing harm.

Initial vegetation clearing and the construction, use, and maintenance of access roads may increase delivery of sediment to waterways and increase sediment deposition. New construction of access roads, upgrades of existing access roads, use of roads, and construction of facilities and transmission line foundations have the greatest potential for sediment delivery. Localized sediment deposition is expected to occur from these activities from pre-construction through closure and restoration. It is likely that some localized rearing habitat may be negatively impacted by sediment deposition to a degree that may contribute to sublethal effects (e.g., reduced growth, density dependence effects due to reduced habitat space, etc.) to juvenile fish rearing in the action area. Whether sediment delivery will cause direct mortality of incubating embryos depends on whether sediment is deposited directly on top of redds in sufficient amounts to cause suffocation or entrapment.

Overall, the magnitude of the increase in sediment delivery and its impact on fish spawning, incubation, and rearing through elevated turbidity and subsequent sediment deposition is difficult to predict. These effects will occur throughout the construction and the active mining periods (approximately 20 years), and over that timeframe NMFS expects that sediment-related effects will extend downstream to the confluence with the SFSR. However, implementation of BMPs and PDFs should effectively minimize the amount of sediment being delivered, and because these PDFs are known to be both proven and effective, turbidity pulses from with projectgenerated sediment are expected to be localized, low-intensity, infrequent, and last for only minutes to hours. Channel dewatering and fish salvage will remove fish from streams where inwater work and mining overlap. Direct impacts from sediment runoff will be restricted to access routes and areas at the edge of the active mine; and sediment impacts in these areas will be further limited by erosion control BMPs (e.g., silt fences, straw waddles, etc.).

Any increase in fine sediment deposition within stream channels has the potential to decrease spawning gravel suitability and decrease benthic invertebrate production within gravel riffles, effects that would impact spawning/incubation and rearing/feeding life stages of Chinook salmon and steelhead. However, with the application of sediment control BMPs and stormwater treatment techniques, the impacts of sediment in surface water, as well as interstitial sediment, to fish are predicted to be measurable but not severe, and should not meaningfully affect salmonid spawning and rearing success.

Provided GRAIP-Lite modeling (Tetra Tech 2024) is accurate, restoration efforts to address chronic sources of sediment delivery in the EFMC, planned road closures, and mine site revegetation efforts will lead to a substantial reduction in sediment delivery to action area streams. Site restoration should ultimately benefit spawning and rearing Chinook salmon and steelhead with an overall, long-term, localized decrease in sediment input into Johnson Creek, Meadow Creek, and the EFSFSR.

# **Water Quantity**

During the pre-construction phase, water will need to be withdrawn to support drilling activities associated with the BRGI portion of the project. However, because flow in Johnson Creek is nearly unimpaired, the flow reduction due to the BRGI portion of the proposed action is very small, and the reduction would occur for only one year, the overall effect of reducing flow in occupied Chinook salmon and steelhead habitat will be minor and not likely to result in harm of ESA-listed salmon and steelhead.

During operations, water rights associated with the proposed action authorize a total of 9.9 cfs of diversion from PODs in the EFSFSR drainage, 9.6 cfs of which will be used for mining, ore processing, etc., with the remainder used for domestic purposes and support of a vehicle maintenance facility (Table 16). The four water rights authorizing diversion of the 9.6 cfs for mining, ore processing, etc., include conditions that will protect "minimum flows" in lower Meadow Creek, in the EFSFSR between the YPP and Sugar Creek, and in the EFSFSR below Sugar Creek (Water Right details for 77-7122, 77-7293, 77-14378,77-7285). These "minimum flows" should ensure that the flow effects of proposed action will not appreciably impair upstream and downstream fish passage. However, the anticipated reductions in flow will affect other aspects of SR Spring/summer Chinook salmon and SRB steelhead biology.

Reducing stream flow affects stream dwelling salmonids in a variety of ways. Food availability for stream dwelling salmonids is generally positively related to streamflow across the entire range of base flows (Harvey et al. 2006; Hayes et al. 2007; Davidson et al. 2010) and reducing stream flow reduces growth of individual salmonids (Harvey et al. 2006) and productivity of salmonid populations (Nislow et al. 2004). Reducing streamflow reduces access to escape cover (Hardy et al. 2006a) which could increase predation risk and could reduce the amount of suitable habitat, thereby increasing competition for suitable habitat. Reducing flow can cause long-term increases in fine sediments in stream substrates (Baker et al. 2011), thereby reducing food production, availability of cover for rearing juveniles, and survival of eggs in redds. Reducing flow increases summer water temperature (Tate et al. 2005; Miller et al. 2007), which will reduce growth and survival of rearing juveniles and could reduce survival of adult holding Chinook salmon. Cold water refugia are important for rearing juvenile Chinook salmon and steelhead

(Sauter et al. 2001; Richter and Kolmes 2005) and for pre-spawning adult Chinook salmon (Berman and Quinn 1991; Torgersen et al. 1999), suggesting that reducing water from cold tributary streams will adversely affect rearing Chinook salmon and steelhead, and possibly holding adult Chinook salmon.

Modeling effects of flow changes on fish populations has historically been problematic (Shirvell 1986; Bourgeios et al. 1996; Hardy et al. 2006; Beecher et al. 2010) and even the most refined flow habitat models tend to underestimate optimal flows (Rosenfield et al. 2016) suggesting that modeling would likely underestimate the effects of flow reductions on SR Spring/summer Chinook salmon and SRB steelhead. Year class strength of many salmonid populations is positively related to streamflow (Ricker 1975; Mathews and Olson 1980; Mitro et al*.* 2003; Elliott et al.1997; Nislow et al. 2004; Arthaud et al. 2010; Beecher et al. 2010; Warkentin et al. 2022; Morrow and Arthaud 2024). A review of 46 studies found that salmonid demography was typically positively related to flow (Kovach et al. 2016). Because relationships of population productivity and streamflow are well documented, we used the relationships productivity and flow for the EFSFSR and SFSR Chinook salmon populations, and the SFSR steelhead population, to estimate the flow-related effects of the proposed action on SR Spring/summer Chinook salmon and SRB steelhead. The regression models also incorporated population density and, because the proposed action would affect flow in only a portion of the EFSFSR Chinook salmon and the SFSR steelhead population areas, we scaled the estimated effects based on the proportion of habitat affected. The flow leverage plots from the regression models are in Figure 34. The methods used to calculate flow effects, and the complete regression results, are in Appendix G.

According to the information provided in the BA, the proposed action would result in a slight increase in flow during MY -2, but flows would be reduced after MY -2. During MY -1 through MY 7, flow reductions will generally increase, each year, reaching a maximum during MY 7, and then will generally decrease through MY 20. A small reduction in flow will persists from MY 21 through approximately MY 78. The analysis in Appendix G uses the relationships of population productivity versus rearing flow (Figure 35) to translate the flow-related effects into effects on population productivity. The maximum flow effect, which will occur during MY 7, will reduce productivity of the EFSFSR Chinook salmon population, the SFSR Chinook salmon population, and the SFSR steelhead population, by 2.2%, 0.015%, and 0.82% respectively. Presuming average population size, these reductions in productivity will result in approximately seven fewer returning EFSFSR Chinook salmon, substantially less than one returning SFSR Chinook salmon, and approximately eight returning SFSR steelhead. The average productivity reduction for MYs -2 through 20 will be 1.1%, 0.007%, and 0.36%, respectively, for the EFSFSR Chinook salmon, SFSR Chinook salmon, and in the SFSR steelhead populations. The estimated reductions in productivity for MYs -2 through 20, and the resultant reductions in returns for MYs one through 23, are in Figures 36 and 37. Although water diversion for mining will cease after MY 20, the effects on flow will persist, due to the filling of the WEP Lake. From MY 21 through approximately MY 78, the flow reduction due to filling the WEP Lake will reduce productivity of the EFSFSR Chinook salmon, SFSR Chinook salmon, and in the SFSR steelhead populations by 0.13%, 0.001% and 0.082%, respectively, and will result in an average of 0.4 fewer returning EFSFSR Chinook salmon, 0.006 fewer returning SFSR Chinook salmon, and 0.8 fewer returning SFSR steelhead, each year.



**Figure 35**. **Flow leverage plots for multivariate regressions of whole life cycle population productivity and average July – September flow (percent of average) for the EFSFSR Chinook salmon population (A), the SFSR Chinook salmon population (B), the SFSR steelhead population (C), and the Secesh River Chinook salmon population (D)**.

Based on the estimated effects on whole life cycle population productivity, the flow related effects of the proposed action would result in 819 fewer juvenile EFSFSR Chinook salmon, ten fewer juvenile SFSR Chinook salmon, and 370 fewer juvenile SFSR steelhead migrating downstream past Lower Granite Dam during MY 7. The estimated annual average reduction for MYs -2 through MY 20, is 406 fewer juvenile EFSFSR Chinook salmon, five fewer juvenile SFSR Chinook salmon, and 163 fewer SFSR steelhead. The long-term effects (i.e., year 23 – 100+) would result in an average of 45, 0.6, and 38 fewer juvenile EFSFSR Chinook salmon, SFSR Chinook salmon, and SFSR steelhead, respectively, migrating downstream past Lower Granite Dam, each year.



**Figure 36**. **Change in population productivity for the EFSFSR Chinook salmon population, the SFSR Chinook salmon population, and the SFSR steelhead population, due to flow reductions caused by the proposed action for MYs -2 through 20.**



**Figure 37. Reduction in adult returns assuming average returning adult population sizes of 341 for the EFSFSR Chinook salmon population, 649 for the SFSR Chinook salmon population, and 945 for the SFSR steelhead population. The EFSFSR Chinook salmon population size is based on Nez Perce Tribe redd survey data, the SFSR Chinook salmon population size is based on IDFG redd survey data, and the SFSR steelhead population size is, respectively, redd survey data, and the SFSR steelhead population size is based on the IDFG steelhead return estimates.** 

**Note:** These population sizes represent averages for 1998-2022, for the EFSFSR Chinook salmon population, 2008- 2022 for the SFSR Chinook salmon population, and 2011-2020 for the SFSR steelhead population.

#### **Fish Passage, Habitat Access, and Habitat Condition**

As described in the DCH effects section, the proposed action will not likely reduce flows to levels that impair upstream or downstream fish passage and the only surface water diversion (EFSFSR between Sugar Creek and the YPP) will meet NMFS criteria for fish passage (NMFS 2022a). Upstream migration is not likely to be impaired by the water diversion, but some downstream migrants will pass via the fish screen/bypass system and will experience migration delay and increased chance of migration mortality.

During the construction of the Burntlog Route or of temporary roads, culverts will be constructed or replaced, which may affect fish access in different sections of streams. Any new or reconstructed crossing will be designed to be fish passable, which will increase or reestablish fish access where it had been reduced or blocked unless there is a risk of passing non-native fish species. As mentioned in the DCH effects section, there are 18 existing crossings along the Burntlog Road (FR-447) that will be replaced and 10 new crossings along newly constructed portions of the Burntlog Route. There is a total of approximately 53 miles of stream segments upstream of the Burntlog Route. Currently almost all stream crossings along the Burntlog Road are impassable culverts, particularly at low flow conditions. The key perennial streams that will be crossed are Burntlog, Trapper, and Riordan Creeks. Access roads to the new transmission line

cross some creeks; however, alteration to most of these crossings may not be necessary. If the crossing needs to be upgraded, the same BMPs will be followed as for the Burntlog Route.

Also, during the construction phase, the proposed action will begin to address fish passage in the EFSFSR subbasin. Construction of the EFSFSR tunnel to bypass the barrier upstream from YPP Lake and removal of the EFSFSR box culvert will both be completed by MY -1 and will facilitate fish passage. Although the partial gradient barrier in Meadow Creek just upstream from EFMC will also be removed during the construction phase of the Project, a new barrier will be created in MY -2 just upstream from the existing barrier to prevent fish passage into upper Meadow Creek where the TSF will be constructed. This new barrier, will block access to approximately 0.43 miles of habitat otherwise accessible to Chinook salmon and steelhead through MY 12. By MY 18, the only barrier impacting ESA-listed salmon and steelhead will be the steep gradient section of Meadow Creek associated with the TSF buttress created as part of the SGP, a barrier that will remain indefinitely and will permanently block volitional access to approximately 3.87 miles of suitable habitat. Because the water rights authorizing water diversion for SGP operations include conditions that protect minimum flows, reduction in streamflow, due to the proposed action, is not likely to further impair fish movement.

Although Chinook salmon were already using habitat upstream from the YPP cascade in years following release of excess hatchery fish, removal of these barriers will result in volitional passage into the upper EFSFSR for the first time since the 1930s. Passage upstream of the cascade will provide access to up to 5.51 miles of habitat with intrinsic potential for Chinook and 5.42 miles of intrinsic potential for steelhead spawning and rearing.

As discussed in the DCH effects section, the amount of instream LWD, pool frequency, pool quality, and off-channel habitat are often indicators of quality fish habitat. Large wildfires in the action area have contributed large amounts of LWD to action area streams, which in turn has increased habitat complexity and created quality refugia by increasing both the number and quality of pools in action area streams. Channel restoration and enhancement will occur in Meadow Creek and the EFSFSR as early as MY -1, work that will involve the addition of instream habitat structures, including LWD jams, boulder clusters, and excavated pools, intended to improve spawning and rearing habitat conditions for salmonids. By MY 11, a restored stream channel and Stibnite Lake will be created across the surface of the YPP backfill. Once this portion of the stream channel restoration is complete, the EFSFSR flows will be routed through the restored channel and Chinook salmon and steelhead will gain access to additional spawning and rearing habitat. Revegetation efforts will also take place to develop riparian vegetation to provide for long-term bank stability, future LWD recruitment, overhead cover, and shade.

Overall, although temporary and short-term effects to Chinook and steelhead could occur from work to restore and enhance these stream reaches, this work will be completed in dewatered work areas, and other than small, temporary, localized, brief turbidity pulses, impacts from these activities will be confined to the effects of fish salvage and handling (see below). Proposed stream restoration and enhancement should ultimately increase habitat complexity in reaches accessible to anadromous fish in the upper EFSFSR and Meadow Creek, resulting in long-term, localized improvements to fish productivity.

#### **Food/Forage**

Quantities of available food/forage can be affected by changes in water quality (e.g., stream temperature, dissolved oxygen, chemical contamination), water quantity, sedimentation, and clearing of riparian vegetation.

Chemical contamination resulting from an accidental spill or through point and nonpoint sources of pollution has the potential to affect the quantity and quality of prey for juvenile salmon or steelhead. Mortality of aquatic invertebrates from a spill of toxic materials would be dependent upon the type and amount of material spilled. Since toxicity is expected to attenuate in a downstream direction, mortality from a spill is not likely to extend more than a mile or two downstream. As discussed in the DCH effects section, contaminant concentrations are not expected to rise to a level acutely toxic to aquatic invertebrates. Furthermore, contaminant concentrations are not expected to substantially alter the quantity or diversity of macroinvertebrate assemblages. Given that juvenile salmonids are opportunistic feeders, as long as a diverse group of macroinvertebrates are protected, some loss of prey items would not be expected to reduce individual fitness of juvenile salmonids rearing in the area.

While the quantity of prey is not expected to be negatively affected, predicted concentrations of arsenic and mercury are expected to impact prey quality. As previously described, arsenic concentrations are predicted to substantially decrease over the life of the project; therefore, we believe accumulation of arsenic in prey tissues will also decrease over time. This conclusion is not unreasonable, as arsenic concentrations in macroinvertebrate tissues were lower in Meadow Creek when measured about 10 years after reclamation. Even though arsenic accumulation is expected to decrease, juvenile fish are still expected to experience sublethal effects as a result of dietary exposures. The proposed action will result in higher mercury concentrations in streams within the Stibnite project area. In addition, the SGP will produce an incremental increase in organic carbon content in the EFSFSR due to sanitary wastewater effluent. Taken together, there will be an increased risk of bioaccumulation of mercury in salmonid prey items and a concomitant increase in bioaccumulation of mercury in fish. Whether, and the extent to which increased mercury bioaccumulation will reduce individual fitness is unknown.

Decreased streamflows in Meadow Creek, Sugar Creek, the EFSFSR, and the SFSR during the construction and post closure (MYs -1 to 20) will cause a reduction in food availability and suitable habitat, which could increase competition for both habitat and available food/forage. This will likely reduce growth and survival of SR Spring/summer Chinook salmon and SR Basin steelhead (Appendix G).

Proposed ground-disturbing activities could affect instream sediment levels, which could in turn affect benthic invertebrate production. However, as discussed in the DCH effects section, the amount of sediment delivered to action area streams is predicted to be measurable but not severe. Therefore, increases in sediment are not expected to impact the quantity or types of food available to ESA-listed salmonids in the action area, and are therefore not expected to impact juvenile Chinook salmon or steelhead growth and survival.

#### 2.10.4.6. Noise and Vibration

Noise and vibration from equipment operation and blasting have the potential to disturb or harm ESA-listed Chinook salmon and steelhead. The potential effects from both of these activities will be discussed in more detail below.

#### *2.10.4.6.1. Equipment Operation.*

Noise and vibration from heavy equipment operating adjacent to live water may disturb fish in the immediate vicinity causing short-term displacement. As described in the BA (Stantec 2024), to avoid injury, instantaneous sound levels should be less than 206 peak dB and sound emitted over extended time (sound emitted repeatedly) should be less than 187 dB (183 dB for fish less than 2 grams) exposure level, referenced at 1 micropascal for sound traveling through water, measured at a distance of 10 meters (Fisheries Hydroacoustic Working Group 2008). However, sound levels over 150 dB can trigger behavioral effects, such as moving to other locations.

Machinery operation adjacent to streams will be intermittent, with actual activity near the stream occurring at various times on any given day. The Federal Highway Administration (2024) indicates backhoe, grader, drill rigs, rock drill, loader, and dump truck noise production ranging between 80 and 98 dB. According to Bauer and Babich (2006), equipment used for surface limestone mining (i.e., jaw and cone crushers, screens, conveyor belts, etc.) produced noise levels ranging from 67 to 111 dB. For mining related noise, Mintek (2024) reported pneumatic and percussion tools operating at levels ranging from 114 to 120 dB, while loaders and haulage truck levels ranging from 90 to 110 bB. Rock crushing equipment, such as crushers and screens, can generate noise levels that can exceed 85 dB. Noise levels can vary depending on the process, with blasting activities ranging from 102.8 to 130.8 dB, and crushing activities ranging from 97 to 116.2 dB (Pal and Mandal 2021; Hebbal and Kadadevaru 2017).

Because the decibel scale is logarithmic, there is nearly a 100-fold difference between noise levels expected from the action and noise levels known to have generated adverse effects to fish species, as discussed above. Therefore, noise related disturbances of this magnitude are unlikely to result in injury or death. It is unknown if the expected dB levels will cause fish to temporarily move away from the disturbance or if fish will remain present. Even if fish move, they are expected to migrate only short distances to an area they feel more secure and only for a few hours in any given day. NMFS does not anticipate short-term movements caused by equipment noise will result in effects of a duration or magnitude that is expected to result in harm of ESAlisted salmon or steelhead.

In addition to sound effects, excessive ground vibrations have the potential to affect salmonids, particularly the sensitive egg life stage (Timothy 2013; Kolden and Aimone-Martin 2013). Smirnov (1954, as cited in ADFG 1991) found significant egg mortality caused by ground vibrations with a peak particle velocity (PPV) of 2 inches per second (ips). Jensen and Collins (2003) found that a PPV of 5.8 ips resulted in 10 percent mortality of Chinook salmon embryos. Faulkner et al. (2008) found that PPVs up to 9.7 ips resulted in significantly higher mortality in *O. mykiss* eggs but there was no increase in mortality when exposed to PPVs of 5.2 or less. The Alaska Department of Fish and Game (ADFG) have PPV restrictions of 2.0 ips to protect salmonids (Timothy 2013).

The reported PPV value for an in-situ soil sampling rig at a distance of 100 feet is 0.011 ips (ATS Consulting 2013). Dowding (2002, as cited in Aimone-Martin and Kolden 2019) reported vibrations causing PPVs ranging from 0.04 to 0.39 ips for most types of equipment, and up to 0.59 ips for higher impact energy sources like pile drivers. This considered, PPV associated with equipment operation in and around the SGP is not expected to create levels or expected to reach levels known to have generated adverse effects to fish species, as discussed above, and are unlikely to result in injury or death.

### *2.10.4.6.2. Blasting.*

Explosives will be used to fracture rock from mine operations. Explosives detonated near water produce shock waves that may be lethal or damaging to fish, fish eggs, or other aquatic organisms. Outside of the zone of lethal or harmful shock waves, the vibrations caused by drilling and blasting have the potential to disturb fish causing stress or altering behavior.

Blasting has the potential to affect Chinook salmon, steelhead, or salmonid redds during operations. In fish, sudden changes in overpressure can cause injury, behavioral changes, temporary or permanent hearing loss, and mortality in extreme cases. Injuries can include hemorrhaging, embolism, and damage to the swimbladder, liver, or other internal organs. In eggs, blasting can cause ground vibrations in substrate that can physically deform, dislodge, and tear or crush the eggs (Aimone-Martin and Kolden 2019).

However, as previously mentioned, most of the blasting will be away from streams, in and near the Yellow Pine, Hangar Flats, and West End Pits, although some may also be required for construction of stream diversions at the YPP, TSF, and TSF Buttress. According to the BA, areas requiring blasting will typically occur on steeper side slopes and in upland areas (Stantec 2024), but the proposed action further states that where the setback distance cannot be met and alterations to the blasting protocol will not adequately mitigate potential harm to fish communities, Perpetua would implement measures to isolate, capture, and relocate ESA-listed fish species from the stream segment where potential for impact exists. Because the BA did not identify where, how often, and how much area could be affected, and they did not identify how many fish might be salvaged for blasting, NMFS is not able to predict how many fish might be affected by salvage activities associated with blasting. Therefore, we do not factor any fish salvage associated with blasting into our effect's analysis, and presume that all blasting EDFs will be applied alongside any streams occupied by ESA-listed salmon or steelhead. The proposed action includes limits on charge sizes, controlled blasting techniques, and using setback distances restricting use by streams (section 1.7.2). These EDFs have been designed according to standards established in Wright and Hopky (1998), ADFG (1991), and Timothy (2013). These EDFs have been shown to be protective of fish (Timothy 2013) and should effectively limit the potential for project-related blasting from affecting spawning/rearing Chinook salmon, steelhead, or their redds.

#### **Fish Salvage and Handling**

In-water work associated with the proposed action will take place during in-water work windows, in dewatered work areas, and only after fish salvage has been conducted. Water will be slowly removed from work areas to allow some fish to leave volitionally, and pumps used for

dewatering will be screened to meet NOAA Fisheries and IDFG standards to avoid entrainment of juvenile fish. Fish removal methods will use blocknets, seining, minnow traps, dip-netting, and electrofishing. Captured fish will be relocated as quickly as possible to pre-planned release areas using aerated and shaded transport buckets holding limited numbers of fish of comparable size to minimize predation (Stantec 2024). Fish salvage will be necessary beginning in the construction phase (i.e., Burntlog Route and YPP dewatering), continuing through operations and closure (i.e., stream restoration and enhancement).

The proposed action did not identify how many electrofishing passes would be employed during fish salvage, but NMFS assumes that a three-pass electrofishing effort will be employed because it is a standard practice. The proposed action stated that NMFS' electrofishing guidelines (NMFS 2000) will be followed when salvaging fish. The guidelines require that field crews be trained in observing animals for signs of stress and shown how to adjust electrofishing equipment to minimize that stress. Electrofishing is not done in the vicinity of redds or spawning adults. All electrofishing equipment operators will be trained by qualified personnel to be familiar with equipment handling, settings, maintenance, and safety. Only direct current units will be used, and the equipment will be regularly maintained to ensure proper operating condition. Voltage, pulse width, and rate will be kept at minimal levels and water conductivity will be tested at the start of every electrofishing session so those minimal levels can be determined. When such low settings are used, shocked fish normally revive instantaneously. Fish requiring revivification will receive immediate care.

In an effort to quantify effects of fish salvage for the SGP, NMFS included the following information in our calculations. We applied fish abundances provided in the BA, and where not provided, used mean fish densities for intensive snorkeling surveys in the EFSFSR drainage reported by IDFG (Poole et al. 2019). Because all sites will be dewatered in stages, NMFS assumes that approximately half of the fish present in a work area will move out of each work area volitionally and not be subject to any form of fish handling. By seining and dip-netting project areas next, NMFS expects that many fish in the area will flee as the block-nets are maneuvered into place. We estimated a 70% capture rate of those fish that remain in work areas with netting. Although seining of fish is likely to cause some elevated stress levels from the contact with the seine and personnel, these effects are not expected to result in injury or death of juvenile salmonids. NMFS also expects that some fish will simply retreat into cover within the work areas and will be subject to electrofishing.

The effects of electrofishing on juvenile steelhead and spring/summer Chinook salmon will be limited to the direct and indirect effects of exposure to an electric field, capture by netting, and the effects of handling associated with transferring the fish back to the streams. Most of the studies on the effects of electrofishing have been conducted on adult fish >12 inches in length (Dalbey et al. 1996). The relatively few studies that have been conducted on juvenile salmonids indicate that spinal injury rates are substantially lower than for larger fish. Smaller fish intercept a smaller head-to-tail potential than larger fish (Sharber and Carothers 1988) and may therefore be subject to lower injury rates (Dalbey et al. 1996; Thompson et al. 1997).

McMichael et al. (1998) found a 5.1% injury rate for juvenile middle Columbia River steelhead captured by electrofishing in the Yakima River subbasin, while Ainslie et al. (1998) reported

injury rates of 15% for direct current electrofishing on juvenile rainbow trout. The incidence and severity of electrofishing damage is partly related to the type of equipment used and the waveform produced (Sharber and Carothers 1988; Dalbey et al. 1996; Dwyer and White 1997). Continuous direct current or low-frequency (equal or less than 30 Hz) pulsed direct current have been recommended for electrofishing because lower spinal injury rates occur with these waveforms, particularly in salmonids (Fredenberg 1992; Dalbey et al. 1996; Ainslie et al. 1998). Only a few recent studies have examined the long-term effects of electrofishing on salmonid survival and growth (Ainslie et al. 1998; Dalbey et al. 1996), indicating that although some fish suffer spinal injury, few dies as a result. However, severely injured fish grow at slower rates and sometimes show no growth at all (Dalbey et al. 1996).

In an effort to estimate the number of fish expected to be injured or killed as a result of this project, NMFS used fish abundances described in the BA (Stantec 2024) and where not provided, used mean fish densities for intensive snorkeling surveys in the EFSFSR drainage reported by IDFG (Poole et al. 2019) to estimate how many fish would be exposed to electrofishing. NMFS assumed a 10% injury rate for fish subject to electrofishing. We have also factored in a 5% electroshocking mortality rate, and a stranding rate of 5% for all project sites. Table 62 estimates the number of Chinook salmon and steelhead likely to be injured, killed, or handled as part of fish salvage operations, by MY and major activity.

## *2.10.4.7.1. Construction.*

Construction of the Burntlog Route will occur from MY -3 to -1, work that will require in-water work associated with bridges and culverts along the route. All bridge work crosses stream segments with fish presence, while only 9 of 25 culverts occur in streams with fish present. Fish salvage is expected to be necessary for anadromous fish as follows:

- Both Chinook and Steelhead: Johnson Creek (Station 1013+00), Burntlog Bridge (Sta. 1597+00), and East Fork Burntlog Bridge (Sta. 1873+00) sites.
- Steelhead Only: Unnamed Trib. to East Fork Burntlog Bridge (Sta. 1803+00), Peanut Creek culvert.
- Chinook Only: EFSFSR Bridge.

Before draining the YPP Lake, a fish barrier will be constructed downstream from the lake (Construction period MY -2 to -1). The barrier will be designed to allow fish to leave the lake but will not allow fish to migrate upstream into the lake. It will be constructed in a window between September 15 to April 1, working to avoid migrating adult salmonids. Once the fish tunnel is complete, EFSFSR water will gradually be diverted into the tunnel and away from the YPP Lake.



### **Table 62 Estimated Number of Juvenile Chinook Salmon and Steelhead Present, Handled, Injured, or Killed from Fish Salvage Efforts by Stibnite Activity and Mine Year.**



Once most of the flow has been re-routed into the tunnel, fish salvage of the Lake will begin, and it will take approximately a week to complete. Based on population estimates when the YPP lake was sampled from 2018 to 2019, very few Chinook salmon or steelhead are expected to be salvaged from the lake. Although the BA anticipated an abundance of no more than five of either species (BA Table 4.1-38), fish sampling from 2018 to 2019 caught 52 Chinook and 4 *O. mykiss*, although they were not able to complete a population estimate (Brown and Caldwell, Rio ASE, and BioAnalysts 2021). Therefore, to make sure we do not underestimate the number of Chinook and steelhead present, NMFS assumed up to 85 Chinook and 25 steelhead could be present in the lake when fish salvage begins.

The box culvert on the EFSFSR and enhancement of the EFSFSR upstream from YPP will also take place in the construction period, completed in MY -1.

Combined, NMFS estimates that fish salvage for the construction period (MYs -3 to -1) will: handle approximately 2,729 Chinook salmon, with 819 subject to electrofishing, injuring 82, and stranding or killing another 82; and will handle approximately 24 steelhead, injuring one, and stranding or killing another one (Table 62). These effects will all occur to the EFSFSR Chinook salmon population and the SFSR steelhead population. However, the injuries and mortalities will be spread across the three-year period, and will not be realized all at once to a single year class of either species.

#### *2.10.4.7.2. Operations and Closure.*

Fish salvage will also be required during operations and closure, beginning with stream enhancement in Lower Meadow Creek in MY 3, decommissioning the tunnel around MY 17, and wrapping up around MY 18 when lower Meadow Creek is restored from the toe of the TSF buttress. Although additional fish salvage may be required upon closure of the Burntlog Route beginning in MY 23+, the segments of the Burntlog Route where Chinook and steelhead are expected to be present will not be decommissioned and these stream crossings are expected to remain in place. Therefore, no additional fish salvage of ESA-listed Chinook salmon or steelhead is anticipated at mine closure for decommissioning of this route.

Combined, NMFS estimates that fish salvage for the operations and closure period (MYs 0 to 23+) will: handle approximately 4,374 Chinook salmon, with 1,312 subject to electrofishing, injuring 131, and stranding or killing another 131; and will handle approximately 438 steelhead, 131 of which will be subject to electrofishing, injuring 13, and stranding or killing another 13. These effects will all occur to the EFSFSR Chinook salmon population and the SFSR steelhead population. However, the injuries and mortalities will be spread across different years of mine operation, and will not be realized all at once to a single year class of either species.

## *2.10.4.7.3. Trap and Haul.*

Trap and haul are an alternative that will only be used if fish arrive volitionally at the downstream end of the tunnel, but do not ascend the fishway (see FOMP Appendix C Table 1). Trap and haul could occur each year the EFSFSR tunnel fishway is operational (from MY -1 to 11), but would only occur when deemed necessary to avoid delay of adult passage near or during the spawning period. Migration periods (i.e., Chinook salmon: Jul. 7 – Sep. 15; steelhead: Apr 1. – May 31) will be monitored via fishway video and Passive Integrated Transponders (PIT) tag detections in the fishway. The period that trap and haul would occur coincides with 1-week prior to the spawning period (all species). If necessary, trap and haul frequency each year is only expected to occur 1-2 times per day depending on multi-species presence.

As described in the FOMP (Brown and Caldwell, McMillen Jacobs, and BioAnalysts 2021), the trap and haul facility has been designed to handle up to 100 Chinook and 100 steelhead annually. Given the short transport distance (i.e., to pool habitat upstream of the tunnel), and if managed correctly, NMFS expects mortality associated with trap and haul to be low (0.1 to 1%) (R. Graves, NMFS Branch Chief, pers. comm., June 2024). If handling up to 100 Chinook or steelhead per year, no more than one adult Chinook salmon or steelhead are expected to be killed by trap and haul in any given year. Because it is not known how effective the tunnel will be at providing unassisted fish passage, it is not clear if this trap and haul mortality will take place, or how frequently. However, if required each year, it will only be required through MY 11, when unimpaired fish passage is restored across the YPP backfill.

# **2.10.5. Effects to Species - Lemhi**

This component of the proposed action could potentially affect SR spring/summer Chinook salmon and SR Basin steelhead. This portion of the project is being implemented with the intent of improving habitat conditions in the Lemhi River subbasin. As described above in the critical

habitat effects analysis, anticipated beneficial effects of this project include increased bank stability, enhanced instream habitat for rearing, spawning, and migrating fish, improved floodplain access, and improved riparian conditions. Potential short-term negative effects to fish include disturbance or salvage of individuals during construction, sediment-related impacts, and possible chemical contamination.

If successfully implemented, this proposed project will immediately increase floodplain connectivity to a 7,000-ft. reach of the Lemhi River. These physical changes are anticipated to improve habitat conditions in the project reach (e.g., increased habitat complexity, reduced stream temperatures, etc.), which in turn is likely to generate improvements in distribution of spawning and rearing fish. Considered in its entirety, this proposed project is anticipated to generate a positive response to ESA-listed fish production (e.g., salmon redds) and growth. Since this segment occurs within the primary Lemhi River Chinook salmon production area, adults will likely select the restored reach for spawning within a few years of project completion, and juvenile densities are expected to increase immediately. Steelhead densities are also expected to increase immediately within this stream reach. Although improved fish production and growth may lead to better survival of Lemhi River populations, this action in itself will not likely be enough to restore either population, due to other limiting factors both in and out of the basin.

Perptua is proposing to complete all in-water work during the locally approved in-water work window (July 1 to 3<sup>rd</sup> week of August). Adult steelhead and steelhead embryos will be absent from the action area at this time and not be exposed to any project impacts. Adult Chinook salmon may be present during the proposed work window and could be affected by the work completed during this time. Chinook spawning is expected to begin in late August.

Reviewing IDFG redd survey data from 2008 to 2023, Chinook salmon redds have periodically been observed in the Lemhi project area, with one found each year in years 2008, 2012, 2015, 2017, 2019, and 2022. The higher value and more frequently used spawning areas in the Lemhi River are found upstream with some additional spawning farther downstream. Juvenile SR Basin steelhead and SR Chinook salmon may be present throughout this portion of the action area. Reach-specific juvenile linear fish density data were provided in the BA from 2015 to 2020, averaging approximately 1,166 Chinook per mi. and approximately 2,916 steelhead per mi.

# **Fish Salvage**

As previously described, construction of the off-channel habitat will take place in the dry, before being connected to the mainstem Lemhi River. Fish salvage will be required in the mainstem once the cofferdam is in place and water is diverted into the newly constructed off-channel habitat.

Considering the number of fish estimated present in the BA, and the same fish salvage rates used previously for fish handling efforts on the Stibnite portion of the effects analysis, NMFS estimates that fish salvage for the construction period (MY -2) will: exposing approximately 774 Chinook salmon to handling, 224 of which will be subject to electrofishing, injuring 22, and stranding or killing another 22; and will handle approximately 1,867 steelhead, approximately 560 of which will be subject to electrofishing, injuring 56, and stranding or killing another 56

(Table 63). Fish salvage for the Lemhi portion of the proposed action will occur once, in MY -2, impacting a single year class of each species.

These effects will all occur to the Lemhi Chinook salmon Lemhi steelhead populations. Although part of the same ESU, the Lemhi Chinook salmon population is part of the Upper Salmon MPG versus the SFSR MPG impacted by the rest of the SGP. However, the Lemhi steelhead population is not only part of the same ESU, but also part of the Salmon River MPG, the same MPG affected by the SGP.

#### **Table 63. Estimated Number of Juvenile Chinook Salmon and Steelhead Present, Handled, Injured, or Killed from Fish Salvage Efforts at the Lemhi River Restoration Project by Mine Year.**



# **Water Quality (Turbidity and Chemical Contamination)**

Juvenile fish will likely be affected by short-term turbidity pulses occurring during installation of temporary cofferdam installation and when introducing water to newly constructed channel segments. Elevated turbidity can cause lethal, sublethal, and behavioral effects in juvenile and adult salmonids depending on the duration and intensity (Newcombe and Jensen 1996). Increased turbidity levels in the action area may result in temporary displacement of all fish from preferred habitat or potential sublethal effects such as gill flaring, coughing, avoidance, and increase in blood sugar levels (Bisson and Bilby 1982; Sigler et al. 1984; Berg and Northcote 1985; Servizi and Martens 1992). However, as previously described in the critical habitat effects analysis, completing excavation in the dry and while implementing a large number of sediment control measures will limit both the frequency and the magnitude of turbidity pulses in the Lemhi River. Turbidity pulses are expected to be short-lived (a few hours) before returning to background levels. Because the Lemhi River is relatively large, the plumes are likely to be confined to one bank and many exposed fish will be able to easily move to adjacent non-turbid habitats, thereby avoiding exposure without harm. Operations will cease should turbidity monitoring show exceedances for more than two monitoring intervals.

The use of heavy machinery adjacent to the stream channel increases the risk for the potential of an accidental spill of fuel, lubricants, hydraulic fluid or similar contaminant into the riparian zone, or directly into the water where they could injure or kill aquatic food organisms, or directly impact ESA-listed species. However, as described in the critical habitat effects analysis, NMFS believes that the proposed fuel spill and equipment leak contingencies and preventions described in the proposed action are sufficient to effectively minimize the risk of negative impacts to ESAlisted fish from toxic contamination, and the risk of chemical contamination is unlikely to occur.

#### 2.10.5.3. Noise and Disturbance

Noise and vibration from equipment operation have the potential to disturb or harm ESA-listed Chinook salmon and steelhead. Noise and vibration from heavy equipment operating adjacent to live water may disturb fish in the immediate vicinity causing short-term displacement.

Increased noise caused by construction vehicles will be similar to that described previously for the Stibnite mine site, except there will be no blasting associated with the Lemhi portion of the project. Specialty mufflers will be used for continuously running generators, pumps, and other stationary equipment, and most of the construction activities will occur outside the flowing channel. While construction equipment will be used to construct the channels, other than the placement of the blockage weir, the construction equipment will be used alongside channels in which fish have no access or have been salvaged. As described in the critical habitat effects analysis, the sound level for a typical construction vehicle (such as an excavator) is typically between 80 and 120 dB (FHWA 2024; Bauer and Babich 2006; Mintek 2024; Pal and Mandal 2021; and Hebbal and Kadadevaru 2017). These levels are below those that should impair fish health or behavior. Effects are therefore unlikely, but those that do occur, will be limited to localized behavioral impacts such as movement disruption or area avoidance for Chinook salmon, steelhead. NMFS does not anticipate short-term displacement caused by equipment noise will result in effects of a duration or magnitude that is expected to result in harm of ESAlisted salmon or steelhead.

#### **2.10.6. Summary of Effects to Chinook Salmon and Steelhead**

The SGP will affect SR Basin steelhead in the Salmon River MPG, specifically the SFSR population; while effects of the Lemhi Restoration portion of the project will also affect the Lemhi River steelhead population in the Salmon River MPG. Effects to steelhead are most likely to occur in the Lemhi River, SFSR, EFSFSR (downstream from YPP), Johnson Creek, Burntlog Creek, Cabin Creek, Meadow Creek, and Sugar Creek.

The SGP will affect SR Spring/summer Chinook salmon for the SFSR MPG, specifically the SFSR and EFSFSR populations; while the Lemhi Restoration portion of the proposed action will affect the Upper Salmon MPG, specifically the Lemhi River population. Effects to Chinook salmon are most likely to occur in the Lemhi River (Lemhi population); SFSR and Cabin Creek (SFSR population); and the EFSFSR, Johnson Creek, Meadow Creek, Sugar Creek, lower reaches of Burntlog Creek, Trapper Creek, Riordan Creek, and EFMC (EFSFSR population).

With the exception of disturbance and fish sampling effects, all of the action's potential effects to species are directly related to fish responses to the previously described effects on DCH. A summary of the various effects follows.

Chinook salmon and steelhead could potentially be affected by chemical contamination. We concluded that the SGP is unlikely to lead to onsite or transport-related chemical spills that reach designated critical habitats, and EDFs and BMPs for herbicide treatment are expected to ensure that weed treatment does not result in more than sublethal effects. Stormwater discharges from site facilities and roadways during construction and operations will contribute contaminants such as PAHs, tire wear particles, and metals during and immediately following storm events.

Discharge from the sanitary wastewater treatment plant is expected to contain phosphorus, nitrogen, caffeine, pharmaceuticals, personal care products, plasticizers, food additives, flame retardants, microparticles, and per and polyflouryl alkyl substances (PFAS) (Bothfeld 2021). All life stages of fish are expected to be exposed to contaminants originating in stormwater runoff from roadways and maintenance and housing facilities as well as to contaminants in sanitary wastewater treatment plant effluent. These exposures are not expected to cause outright lethality; however, some sublethal effects (e.g., altered behavior) may occur.

Discharge of mine-contact water, either through point sources or nonpoint sources is expected to alter water quality in the project area, with arsenic, antimony, copper, and mercury being the contaminants of greatest concern. These contaminants are expected to elicit sublethal effects (e.g., reduced growth, tissue damage, altered behavior, reduced olfaction) in fish that develop or rear in project area streams through either waterborne or dietary exposures. There is some risk that early life stage fish (incubation through early exogenous feeding) may experience low levels of mortality (<1 percent) if a redd is continually exposed to predicted maximum concentrations of antimony in Meadow Creek during the early years of operation. Similarly, it is possible that early life stage fish may experience some mortality (< 1 percent) if a redd is continually exposed to predicted maximum concentrations of arsenic in the EFSFSR downstream of the YPP during some years in early closure. Because the proposed action is expected to improve arsenic and antimony concentrations in these streams, the risk of mortality of early life stage fish, although not eliminated, will be reduced. The proposed action is expected to cause incremental increases in total mercury in the water column of streams inhabited by salmon and steelhead. These increases could lead to greater bioaccumulation of mercury in fish tissues. Additionally, an incremental increase in organic carbon content in the EFSFSR due to sanitary wastewater effluent could increase in methylation potential in the EFSFSR. Bioaccumulation of mercury in fish tissue is not expected to reach lethal levels; however, some fish may experience sublethal effects. The degree to which these sublethal effects may manifest to reduced survival is unknown.

Considering evidence of successful incubation and early rearing in Meadow Creek and the EFSFSR under current water quality conditions, and considering the SGP will reduce antimony and arsenic concentrations in these streams and copper concentrations will remain substantively the same, we do not believe these potential low levels of mortality or sublethal effects will reduce abundance or productivity in the population throughout the life of the project.

The SGP will alter temperatures of streams within the mine site in perpetuity. Climate change is expected to exacerbate the impacts of the proposed action; however, adherence to the performance objectives relative to stream temperatures will reduce the magnitude of thermal impacts. The degree to which temperatures are altered varies by reach. Steelhead adult migration and spawning are unlikely to be significantly impacted by the SGP since these life stages utilize the habitat in spring and early summer when thermal conditions are not expected to be impacted by the SGP. Juvenile steelhead rearing will be adversely affected by temperature changes, similar to Chinook salmon. All life stages of Chinook salmon will experience negative effects from temperature changes induced by the SGP. Meadow Creek will experience the greatest thermal impacts as a result of the SGP and will experience the greatest change between MYs 23 and 27. This is due to reactivation of the reconstructed Meadow Creek channel atop the TSF.

Meadow Creek will also be further impacted by thermal loading from the wastewater treatment plant discharge, which will exert its greatest effects in MYs 4 through 6.

Meadow Creek temperatures are predicted to reach critical levels upstream of EFMC during MYs 23 through 27, potentially causing issues such as prespawn mortality for adult Chinook salmon and reduced survival for Chinook salmon embryos and both Chinook salmon and steelhead juveniles. Temperatures in Meadow Creek downstream of EFMC will be reduced relative to existing conditions; however, temperatures will still be elevated to a level that may cause sublethal effects in some juvenile steelhead and Chinook salmon. The Meadow Creek reaches that may be used for spawning and rearing represent less than 0.1 percent of the total intrinsic potential habitat for the SFSR steelhead population and less than 1 percent of total intrinsic potential habitat for the EFSFSR Chinook salmon population. When considering: (1) the habitat represents a small fraction of the population production potential; (2) presence of nearby suitable habitat (i.e., EFSFSR, EFMC, downstream reach of Meadow Creek); and (3) only a few individuals are likely to experience lethal effects (e.g., Chinook salmon embryos; or delayed mortality of juvenile salmon or steelhead due to sublethal effects from exposure to elevated temperatures), we do not anticipate population-level impacts as a result of these localized temperature changes.

The proposed action includes measures to reduce or avoid sediment delivery and turbidity impacts, including silt fences, straw waddles, graveling and applying dust control chemicals on roads, limiting equipment stream crossings, operating in dewatered work areas, slowly rewatering work areas, etc. Therefore, turbidity levels are not expected to reach lethal levels, but are expected to reach levels that generally only elicit behavioral responses from salmonids. Vegetation clearing and the construction, use, and maintenance of access roads may increase delivery of sediment to waterways, which may increase sediment deposition. However, with the application of sediment control BMPs and stormwater treatment techniques, the impacts of sediment in surface water, interstitial spaces, etc. as well as the effects to fish, are not predicted to be measurable but not severe. GRAIP-Lite modeling suggests that the combination of restoration efforts to address chronic sources of sediment delivery in the EFMC, planned road closures, and mine site revegetation efforts, should ultimately result in an overall decrease in sediment input into Johnson Creek, Meadow Creek, and the EFSFSR.

The proposed action will reduce flow within and downstream from the project area from the construction through closure stages, and in Sugar Creek and downstream reaches for 100+ years post closure. The flow effects will be very small during the construction phase, will be largest during mine operation, and will be very small during post closure. The flow effects will peak in MY 7, resulting in approximately eight fewer returning SR Spring/Summer Chinook salmon and eight fewer returning SR Basin steelhead. The flow effects will generally decrease from MY 7 through MY 20, and will decrease to approximately 0.4 cfs during post closure. The long-term flow effects will reduce adult returns of SR Spring/summer Chinook salmon and SR Basin steelhead by an average of approximately 0.43 and 0.78, respectively.

Fish passage will be improved across the action area in both the temporary and the long term, allowing increased volitional access to stream segments not occupied since the 1930s. Both Chinook salmon and steelhead will also gain access to new spawning and rearing habitat in the EFSFSR and Meadow Creek as soon as the tunnel is complete in MY -1. Minimum flows,

stipulated in the water rights conditions, should protect fish passage within and downstream from the project area.

The quantity and quality of forage available to salmonids will be reduced in localized areas at the Stibnite site due to channel relocation, changes in water quality and quantity, sedimentation, and clearing of riparian vegetation. The quantity of forage will be reduced downstream of channels that are dewatered or relocated due to the loss of wetted habitat and loss of riparian vegetation. Revegetation efforts are expected to forage reductions in localized areas are temporary or shortterm in nature. In addition, flow reductions will also cause reductions in benthic invertebrates as well as reduce the rate of benthic invertebrate drift. In the long-term, given restoration of stream channels and increased channel complexity, we expect the quantity of invertebrates to be similar to, or greater than current conditions. Short-term losses of forage are expected to increase competition among rearing juveniles. Predicted concentrations of arsenic and mercury are expected to impact prey quality. As previously described, arsenic concentrations are predicted to substantially decrease over the life of the project; therefore, we believe accumulation of arsenic in prey tissues will also decrease over time. The proposed action will result in slightly higher mercury concentrations in streams inhabited by salmonids. In addition, the SGP will produce an incremental increase in organic carbon content in the EFSFSR due to domestic sanitary wastewater effluent. Taken together, there will be an increased risk of bioaccumulation of mercury in salmonid prey items. Because mercury methylation is dependent upon a myriad of factors, it is not possible to quantify the potential decrease in prey quality associated with mercury accumulation.

Noise and vibration from blasting and heavy equipment operating adjacent to live water may disturb fish in the immediate vicinity causing short-term displacement. However, equipment operation in and around the SGP is not expected to create levels expected to reach levels known to have generated adverse effects to fish species, and is unlikely to result in injury or death. Similarly, design criteria and BMPs for blasting are expected to be protective of salmon, steelhead, and their redds.

In-water work associated with the proposed action will take place during in-water work windows, in dewatered work areas, and only after fish salvage has been conducted. Most of the project-related fish salvage will occur during the construction period (MY -3 to -1), during construction of stream crossings along the Burntlog Route and in the EFSFSR, channel reconstruction in Meadow Creek and the EFSFSR, and when draining the YPP Lake. Fish salvage will also be required during operations and closure, beginning with stream enhancement in Lower Meadow Creek in MY 3, decommissioning the tunnel around MY 17, and wrapping up around MY 18 when lower Meadow Creek is restored from the toe of the TSF buttress.

For the Stibnite portion of the project, NMFS estimates that fish salvage for the construction period (MYs -3 to -1) will handle approximately 2,729 Chinook salmon, subjecting approximately 819 Chinook salmon to electrofishing, injuring 82, and stranding or killing another 82. It will also handle approximately 24 steelhead, subjecting approximately 8 steelhead to electrofishing, injuring one, and stranding or killing another one. These effects will be split between the SFSR and EFSFSR populations for Chinook and the SFSR population for steelhead, and the effects will be spread across three-year classes for each species.

NMFS estimates that fish salvage for the operations and closure period (MYs 0 to 23+) will: handle approximately 4,374 Chinook salmon, subjecting 1,312 to electrofishing, injuring 131, and stranding or killing another 131; and will handle approximately 438 steelhead, electrofishing 131, injuring 13, and stranding or killing another 13. These effects will all occur to the EFSFSR Chinook salmon population and the SFSR steelhead population. However, the injuries and mortalities will be spread across three different years of mine operation, and will not be realized all at once to a single year class of either species.

For the Lemhi fish salvage efforts, fish will also be salvaged during the construction period (MY -2). NMFS estimates that approximately 774 Chinook salmon will be handled, 224 of which will be subject to electrofishing, injuring 22, and stranding or killing another 22; and approximately 1,867 steelhead will be handled, 560 of which will be subject to electrofishing, injuring 56, and stranding or killing another 56. Fish salvage for this portion of the proposed action will occur once, in MY -2.

For the Lemhi portion of the project, fish salvage could result in the loss of up to two returning steelhead, but is not likely to result in the loss of more than one adult equivalent Chinook salmon. These effects would only occur during a single year.

Trap and haul is an alternative that will only be used if fish arrive volitionally at the downstream end of the tunnel, but do not ascend the fishway. NMFS expects mortality associated with trap and haul to be low (0.1 to 1%), and if up to 100 Chinook or steelhead are handled in this manner each year, no more than one adult Chinook salmon or steelhead are expected to be killed by trap and haul in any given year. These effects would be realized to the EFSFSR population of Chinook salmon and to the SFSR population of steelhead.

For water quality effects, it is not possible to quantify the number of juvenile fish that may suffer sublethal effects to an extent that will reduce their survival; therefore, it is not possible to express water quality effects in terms of fish population productivity or production. However, based on available information population-level abundance or productivity reductions are not expected to occur given predicted water column concentrations. As described above, construction-related effects (e.g., fish salvage, stranding, crushing, etc.) are based on a known area that will be affected, fish population data are available, and the effects of fish salvage are well documented. Therefore, the number of SR spring/summer Chinook salmon and SR Basin steelhead killed by construction/salvage activities, can be calculated. Likewise, because population trend data and streamflow gage are available, the effects of changing streamflow on population productivity can be estimated, which can be used to calculate the effects on fish production. We used natal stream to Lower Granite Dam (LGD) survival and SAR data from the Idaho Anadromous Monitoring Annual Reports (2015-2022) to express construction/salvage and the flow-related effects as the number of juveniles migrating downstream past LGD, and as the number of adult returns. These results are summarized in Tables 64 and 65, and constitute the quantifiable take that will result from the proposed action.

<b>Mine</b> Year	<b>Chinook Salmon</b> <b>Production Loses Due to</b> <b>Construction Activities</b>		<b>Chinook Salmon</b> <b>Production Loses Due to</b> <b>Due to Flow Reductions</b>		<b>Total Chinook Salmon</b> <b>Production Losses</b>	
	<b>Juveniles</b> At LGD	<b>Adult</b> <b>Returns</b>	<b>Juveniles</b> At LGD	<b>Adult</b> <b>Returns</b>	<b>Juveniles At</b> <b>LGD</b>	<b>Adult Returns</b>
$-3$	0.121	0.00121			0.04	0.0004
$-2$	$\overline{a}$		$-40.8$	$-0.39$	$-40.7$	$-0.39$
$-1$			160	1.5	160	1.5
$\mathbf{1}$	$\blacksquare$		373	3.5	373	3.5
$\overline{2}$	20.4	0.19	770	7.3	790	7.5
$\overline{3}$	27.0	0.26	556	$\overline{5.3}$	583	5.5
$\overline{4}$			600	5.7	600	5.7
5	$\overline{a}$	$\overline{\phantom{0}}$	267	2.5	267	2.5
6			574	5.4	574	5.4
$\overline{7}$	$\overline{\phantom{0}}$	$\qquad \qquad \blacksquare$	829	7.8	829	7.8
8			796	7.5	796	7.5
9			495	4.7	495	4.7
10			515	4.9	515	4.9
11			556	5.3	556	5.3
12			467	4.4	467	4.4
13			502	4.8	502	4.8
14			362	3.4	362	3.4
15			275	2.6	275	2.6
16			164	1.6	164	1.6
17	0.25	0.0024	181	1.7	181	1.7
18	11.1	0.11	187	1.8	198	1.9
19			142	1.3	142	1.3
20	$\overline{\phantom{a}}$	$\overline{\phantom{0}}$	140	1.3	140	1.3
$20+$	$\overline{\phantom{a}}$		45.9	0.43	45.9	0.43
$Ave^2$	2.6	0.02	403	3.8	406	3.8

**Table 64. Reduction in SR spring/summer Chinook salmon juveniles migrating downstream past Lower Granite Dam (LGD), and reduction in returning adults, due to construction/salvage and flow effects of the proposed action.**

**Notes:** 1. Some or all of the take attributed to construction in mine year (MY) -3 could actually occur in MYs -3, -2, or -1. 2. Average for MYs -3 through 20, flow effects described for 20+ will persists for approximately 57 years.

<b>Mine</b> Year	<b>Steelhead Production</b> <b>Loses Due to</b> <b>Construction Activities</b>		$\alpha$ now encode or the proposed action. <b>Steelhead Production</b> <b>Loses Due To Flow</b> <b>Reductions</b>		<b>Total Steelhead Production</b> <b>Losses</b>	
	<b>Juveniles</b> At LGD	<b>Adult</b> <b>Returns</b>	<b>Juveniles</b> At LGD	<b>Adult</b> <b>Returns</b>	<b>Juveniles At</b> <b>LGD</b>	<b>Adult Returns</b>
$-3$	0.101	0.00211	$\equiv$		0.03	0.0007
$-2$	0.03	0.0006	$-20.4$	$-0.42$	$-20.3$	$-0.42$
$-1$	0.14	0.0028	61.7	1.3	61.9	1.3
1	$\blacksquare$		145	3.0	145	3.0
$\overline{2}$	$\frac{1}{2}$		289	6.0	289	6.0
$\overline{3}$	3.3	0.07	221	4.6	224	4.7
$\overline{4}$	$\overline{\phantom{0}}$		255	5.3	255	5.3
5	$\overline{\phantom{a}}$		90.2	1.9	90	1.9
6	$\equiv$		242	5.0	242	5.0
$\overline{7}$			370	7.7	370	7.7
8			334	6.9	334	6.9
9	$\blacksquare$	$\overline{\phantom{a}}$	190	3.9	190	3.9
10			197	4.1	197	4.1
11	$\overline{\phantom{a}}$	$\overline{\phantom{0}}$	211	4.4	211	4.4
12			184	3.8	184	3.8
13			202	4.2	202	4.2
14			151	3.1	151	3.1
15			118	2.5	118	2.5
16			67.1	1.4	67.1	1.4
17	0.18	0.0037	80.0	1.7	80.2	1.7
18	1.3	0.03	82.2	1.7	83.5	1.7
19	$\overline{\phantom{0}}$		62.3	1.3	62.3	1.3
20	$\overline{\phantom{0}}$		58.5	1.2	58.5	1.2
$20+$			37.5	0.78	37.5	0.78
$Ave^2$	0.22	0.005	163	3.4	163	3.4

**Table 65. Reduction in SRB steelhead juveniles migrating downstream past Lower Granite Dam, and reduction in returning adults, due to construction/salvage and flow effects of the proposed action.**

**Note:** 1. Some or all of the take attributed to construction in mine year (MY) -3 could actually occur in MYs -3, -2, or -1. 2. Average for MYs -3 through 20, flow effects described for 20+ will persists for approximately 57 years.

#### **2.11. Cumulative Effects**

"Cumulative effects" are those effects of future state or private activities, not involving federal activities, that are reasonably certain to occur within the action area of the federal action subject to consultation [50 CFR 402.02 and 402.17(a)]. Future federal actions that are unrelated to the proposed action are not considered in this section because they require separate consultation pursuant to section 7 of the ESA.

Some continuing non-federal activities are reasonably certain to contribute to climate effects within the action area. However, it is difficult if not impossible to distinguish between the action area's future environmental conditions caused by global climate change that are properly part of

the environmental baseline versus cumulative effects. Therefore, all relevant future climaterelated environmental conditions in the action area are described earlier in the discussion of environmental baseline (Section 2.4).

The Stibnite portion of the action area is primarily managed by the BNF and PNF. A few small parcels of private property and state-administered lands are scattered throughout the action area. Uses on these non-federal lands are not expected to change in the foreseeable future. Activities in the action area include road/trail maintenance performed by non-Federal entities (e.g., Valley County, Idaho State Parks and Recreation) and recreation (e.g., camping, fishing, hiking, etc.). These activities will continue to influence water quality and habitat conditions for anadromous fish in the action area. Riparian and stream corridors have been negatively impacted by roads and trails and these impacts will continue in the future.

The Lemhi River portion of the action area is entirely privately owned, and is currently operating as a ranch. Land use at the ranch has typically revolved around agriculture and grazing. However, the LRLT is negotiating a perpetual conservation easement for the property to ensure the restoration benefits and associated mitigation credits persist for at least the length of the predicted temporal loss for the SGP. If signed, the agreement is expected to limit the property's uses in order to protect its conservation values. In conservation easements, the LRLT typically commits to a perpetual working partnership with all present and future landowners to ensure that the conservation easement is honored, and assists and encourages landowners to engage in conservation focused land management practices.

The impacts of these activities on the current condition of ESA-listed species and designated critical habitats within the action area was described in the Status of the Species, Status of Critical Habitat, and Environmental Baseline sections of this opinion. Current levels of these activities are likely to continue into the future and are unlikely to be substantially more severe than they currently are.

# **2.12. Integration and Synthesis**

The Integration and Synthesis section is the final step assessing the risk that the proposed action poses to species and critical habitat. In this section, we add the effects of the action (Section 2.5) to the environmental baseline (Section 2.4) and the cumulative effects (Section 2.6), taking into account the status of the species and critical habitat (Section 2.2), to formulate the agency's biological opinion as to whether the proposed action is likely to: (1) reduce appreciably the likelihood of both the survival and recovery of a listed species in the wild by reducing its numbers, reproduction, or distribution; or (2) appreciably diminish the value of designated or proposed critical habitat as a whole for the conservation of the species.

# **2.12.1. Critical Habitat**

Critical habitat throughout the SR Spring/summer Chinook salmon and SR Basin steelhead designations, ranges from excellent in wilderness areas, to degraded in areas of human activity. Historical mining pollution, sediment delivery from historical logging practices, and degraded riparian conditions from past grazing were major factors in the decline of anadromous fish populations in the action area. Habitat-related limiting factors for recovery of one or more

populations within the action area include excess sediment, degraded riparian conditions, passage barriers, and high-water temperatures (NMFS 2017). Climate change is likely to exacerbate several of the ongoing habitat issues, in particular, increased summer temperatures.

The impacts of federal and non-federal land use activities on critical habitat are reflected in the environmental baseline section of this document. Current levels of these uses are likely to continue into the future and are unlikely to be substantially more severe than they currently are. It is difficult if not impossible to distinguish between the action area's future environmental conditions caused by global climate change that are properly part of the environmental baseline versus cumulative effects. Therefore, all relevant future climate-related environmental conditions in the action area are described in the environmental baseline.

The SGP portion of the action is located in headwater tributaries of the SFSR, with proposed mining activities to take place primarily in areas previously disturbed by mining beginning around 1919. Excavation of the YPP began in 1938, and volitional passage of Chinook salmon and steelhead to the EFSFSR and Meadow Creek was eliminated at that time. Additional mining continued at the site in the 1940s, 1970s, 1980s, and 1990s. Mining, milling, smelting, and leaching activities left behind impacts including underground mine workings, multiple open pits, development rock dumps, mill tailings deposits, cyanidation heap leach pads, neutralized (spent) heap leach ore piles, a mill and smelter site, three town sites, camp sites, a washed-out earthen dam (with its associated erosion and downstream sedimentation), haul roads, an abandoned water diversion tunnel, an airstrip, and other disturbances.

Past mining activities have resulted in ongoing releases of contaminants to surface water and groundwater at the site including elevated concentrations of antimony, arsenic, copper, lead, mercury, and cyanide. Most notable are elevated concentrations of arsenic and antimony. Past mining activities have also caused alterations to stream configurations and habitat including formation of the YPP lake, creation of a fish passage barrier immediately upstream of the YPP lake, sediment and tailings deposits, development rock dumps, and channel diversions.

Although not formally identified as a CERCLA Superfund site, some mining operators at the site conducted activities to reduce the release of hazardous substances before 2001. Notable work included diverting Meadow Creek and stabilizing the Bradley Tailings/SODA disposal area, which was completed in 1999. In addition, the USFS began using its CERCLA authorities to address legacy mining impacts at the site since 2001. In 2002, the USFS removed tailings from a pond and soils located at the former smelter stack area, and the Meadow Creek floodplain was reconstructed in the former pond area. In 2004 and 2005, the USFS reconstructed Meadow Creek directly downstream of Smelter Flats. This included the removal of tailings from the channel and depositing this material in a new containment cell located on the SODA. The new channel banks were revegetated with willow plants and the old channel was backfilled and reclaimed. In 2009, the USFS regraded and covered a portion of the remaining tailings at Smelter Flats to prevent further erosion and exposure risk. It is in these restored stream reaches, that surplus hatchery Chinook salmon have periodically been placed by the NPT and IDFG and have successfully spawned and reared for the first time in decades.

Population-specific habitat concerns for the SR Spring/summer Chinook salmon SFSR MPG, which includes the EFSFSR and SFSR populations, include: fine sediment; high stream temperatures; and passage barriers. Limiting factors for the SFSR steelhead population include fine sediment, migration barriers (of which the YPP is identified as a barrier), and degraded riparian conditions, which contribute to elevated stream temperatures. Based on NMFS' most recent 2022 5-Year Reviews (NMFS 2022b; 2022c), recommended actions specific to SR spring/summer Chinook salmon and SR Basin steelhead in the EFSFSR include improving water quality by reclaiming abandoned mine sites such as the Cinnabar Mine; continuing to conduct appropriate road maintenance, road obliteration, road relocation, and road resurfacing; improving riparian conditions in disturbed areas; eliminating passage barriers; restoring floodplains; and improving planning for potential climate change effects by continuing to monitor stream temperature and validate fish distribution in modeled cold water refugia.

Within the Lemhi River portion of the action area, fish habitat has been affected by construction and maintenance of roads, livestock grazing, channelization and straightening of stream channels, conversion of uplands and wetlands into agriculture land, construction and maintenance of water diversions, and extensive water use for irrigated agriculture. Riparian and instream habitat in the mainstem and most tributaries have been affected, with resultant increases in water temperature and sediment; reduced access to riparian wetlands, side channels, and tributary stream habitat; and reduced cold water refugia.

Important activities to advance the recovery of Chinook salmon populations in the Upper Salmon River MPG include: (1) increasing habitat complexity by creating multi-threaded channels; (2) increasing rearing habitat by increasing floodplain connectivity; (3) reducing the W:D, stabilizing streambanks, and improving willow-dominated riparian areas; and (4) maintaining and improving instream flows and tributary connectivity.

Recommended actions for the Lemhi River population of SR Basin steelhead include; (1) improving riparian conditions in disturbed areas; (2) eliminating passage barriers; (3) increasing winter juvenile rearing habitat by increasing floodplain connectivity and complex habitat structure; (4) reducing W:D; (5) increasing low- to zero-velocity pool habitat with cover; (6) providing more side channel and multi-threaded channel habitat; and (7) reducing fine sediment delivery to streams. Specifically, in the upper Lemhi River, recommended actions to improve conditions for steelhead include increasing habitat complexity by creating multi-threaded channels, narrowing the W:D, stabilizing banks, increasing willow-dominated riparian areas, maintaining and improving instream flows, and improving tributary stream connections to the to the mainstem Lemhi River.

Various habitat restoration projects have locally improved riparian and stream channel habitat and improved flow in the lower reaches of some of the tributaries and in portions of the mainstem. These improvements have resulting in localized improvements in salmonid productivity.

Streams within the action area are vitally important to the recovery of anadromous fish species. There are a number of heavily used Chinook salmon and steelhead spawning areas in the action area in Johnson Creek, the EFSFSR, and the Lemhi River. Tributary habitat will likely become
even more important for thermal refugia in the face of climate change. Recreation, use of the existing road system, mining, and agriculture are the primary human activities in the action area. Roads from legacy logging remain on the landscape and are a threat to the aquatic ecosystem. In more recent times, wildfire has become the largest disturbance mechanism in the SFSR subbasin. Sediment conditions have generally been on an improving trend, likely due to restoration actions and changes to land management approaches in the action area. Water temperatures are currently warmer than optimal and will likely continue to warm into the future as a result of climate change. Riparian conditions are degraded in areas where roads are located in the RCA and in areas used for mining and agriculture. Although there are some localized areas heavily impacted habitat as described above, habitat conditions in mainstem rivers and tributary streams within the action area are good overall.

At this time, Chinook salmon only access portions of historical habitat upstream of the YPP if adults are outplanted by IDFG and the NPT. The upper reaches of Meadow Creek, above the SODA are not used by adult Chinook salmon and it is unlikely that juvenile Chinook salmon can migrate through the Meadow Creek reach adjacent to the SODA to access upstream habitat. Steelhead, on the other hand, have not had access to these upper reaches since around 1938, locally affecting overall SFSR steelhead spatial structure. Although both species currently have access to the Lemhi River portion of the action area, Chinook and steelhead abundance and productivity are both hampered due to degraded habitat conditions in the mainstem Lemhi River.

Designated critical habitat within the action area will be negatively impacted in the temporary (less than 3 years), short-term (less than 15 years), and long-term (greater than 15 years) timeframes. Negative impacts are associated with: (1) increased sediment delivery, (2) impeded passage; (3) increased water temperatures; (4) contaminant contributions; (5) reduced flows; (6) altered riparian vegetation; (7) reduced floodplain connectivity, and (8) reduced forage. The proposed action will also have some positive impacts including elimination of migration barriers, addressing chronic sources of sediment delivery in EFMC, increased channel complexity, reduced contaminant concentrations, and improved stream temperatures.

Increased sediment delivery to action area streams from activities associated with construction, stream restoration, and overall ground disturbance will have temporary and short-term effects. Increases in fine sediment deposition within stream channels have the potential to decrease spawning gravel suitability and decrease benthic invertebrate production within gravel riffles, potentially impacting spawning/incubation and rearing/feeding life stages of Chinook salmon and steelhead. However, spawning substrates are generally FA in action area streams, and GRAIP-Lite modeling suggests that the amount of sediment delivered to action area streams will decrease in comparison to the baseline condition. With application of proposed sediment control BMPs, the deposition of sediment in action area waters is not expected to degrade habitat in the temporary or short-term; and following closure of the mine, obliteration of access roads, and restoration efforts in the EFMC to address chronic sources of sediment delivery, the proposed action is expected to result in a substantial decrease in annual sediment delivery into Meadow Creek and the EFSFSR. Therefore, we anticipate that spawning and rearing habitat will be negatively impacted in small, localized areas immediately following instream and grounddisturbing activities as a result of turbidity pulses and subsequent sediment deposition.

Passage will be impaired for short periods of time at discrete locations (i.e., where channels are being dewatered to facilitate stream crossing construction or channel relocation), but will be locally restored across most of the SGP portion of the action area as early as MY-1. Construction of the fish tunnel, providing passage to habitat in the EFSFSR and lower Meadow Creek, will restore access to spawning and rearing habitat not volitionally accessible to Chinook salmon or steelhead since 1938. This will locally restore lost spatial structure to the EFSFSR Chinook and SFSR steelhead populations. Similarly, new off-channel habitat created in the Lemhi River portion of the project in MY-2 will immediately restore habitat complexity, locally improving abundance and productivity by creating new migration, spawning, and rearing habitat for both populations Chinook salmon and steelhead in the Lemhi River mainstem.

At the SGP site, stream temperature will be negatively impacted in Meadow Creek and the EFSFSR at different spatial and temporal scales. These impacts are primarily due to channel relocation because it takes years for riparian vegetation to grow to a height that will provide sufficient shade and cooler microclimates adjacent to streams. In addition, draining of the YPP and pit dewatering also impact stream temperature by removing the cooling influences of groundwater recharge. Modeled negative impacts will primarily occur during early closure, following reintroduction of flow into the newly constructed Meadow Creek channel on the TSF. Increased solar radiation along this long reach of Meadow Creek will contribute to warm downstream temperatures. Meadow Creek, below the TSF, will experience the most severe effects. In order to support adult migration and spawning, incubation, and juvenile rearing, we believe 7DADM water temperatures should be below 18, 13, and 16°C, respectively. Some embryo mortality may occur in redds constructed early in the spawning season for up to 32 years due to average weekly maximum temperatures being greater than 13°C in Meadow Creek below the TSF. Similarly, elevated summer temperatures in Meadow Creek during this extended time period may cause rearing juveniles to experience sublethal effects such as reduced growth or increased susceptibility to disease due to average weekly maximum temperatures being greater than 16°C. Over the longer term (beyond MY 52), average weekly maximum temperatures are expected to be below the applicable temperature thresholds for summer and fall. Meadow Creek, below the TSF, represents a small fraction of the overall habitat available for spawning and rearing and juvenile fish and potential spawning adults are able to utilize nearby suitable habitat if conditions in Meadow Creek are not favorable. Temperature impacts will lessen with distance downstream considering colder water contributions from tributary streams. For example, the EFSFSR upstream of Meadow Creek is cold and will dampen the temperature impacts downstream of its confluence with Meadow Creek. In the EFSFSR, below Meadow Creek, there will be a couple of years where summer and fall stream temperatures will be slightly higher than 16°C and 13°C, causing some individual fish to experience sublethal effects and potentially some embryo mortality in redds constructed early in the season. Modeled stream temperatures in the EFSFSR below Sugar Creek will be below protective temperature thresholds and are expected to support spawning, rearing, and migratory life stages of salmon and steelhead. In the Lemhi River, allowing the River to better interact with its floodplain, in addition to revegetation efforts associated with this restoration project, will address local habitat restoration priorities, locally buffering solar inputs along this reach of the Lemhi River.

Removal of the historical, unlined mine waste disposal material will have an early, positive impact on water quality; however, placement of permanent mine facilities (e.g., TSF

embankment and buttress, and waste rock backfill into the pits) and discharge of treated mine contact water will negatively impact water quality during operations and early closure. During late closure, mine contact water is not expected to require treatment in perpetuity because seepage from the TSF and TSF embankment and buttress is predicted to meet water quality criteria. The net effect of these actions will be an overall reduction in antimony, arsenic, and copper concentrations in Meadow Creek and the EFSFSR. Yet, concentrations of one or more these contaminants are predicted to be at levels that may cause some mortality of embryos and sublethal effects in adult and juvenile fish. Mercury concentrations will slightly increase in Meadow Creek, EFSFSR, and Sugar Creek. Total mercury contributions from West End Creek will increase by at least an order of magnitude. Mercury concentrations will be above levels that have been associated with sublethal effects caused by bioaccumulation of mercury (NMFS 2014). Whether harmful bioaccumulation occurs in the action area will depend on methylation potential. Overall, although there is risk of harmful effects to individuals from contaminant exposures, we believe the water quality PBF will continue to support spawning, incubation, rearing, and migration for salmon and steelhead.

The proposed action will reduce flow in SR spring/summer Chinook salmon and SR Basin steelhead DCH during MYs  $-2 - 20$ , with smaller reductions in MYs  $21 - 78$ . These reductions in flow will reduce average productivity of DCH by 1.1%, 0.007%, and 0.36%, respectively, for DCH in the EFSFSR Chinook salmon, SFSR Chinook salmon, and in the SFSR steelhead population areas, for MYs  $-2 - 20$ . For MYs  $21 - 78$ , flow reductions will reduce productivity of DCH by 0.13%, 0.001% and 0.082%, respectively, in the SFSR Chinook salmon, and in the SFSR steelhead population areas.

The quantity and quality of forage available to salmonids will be reduced in localized areas at the Stibnite site due to channel relocation, changes in water quality and quantity, sedimentation, and clearing of riparian vegetation. The quantity of forage will be reduced downstream of channels that are dewatered or relocated due to the loss of wetted habitat and loss of riparian vegetation. Revegetation efforts are expected to forage reductions in localized areas are temporary or shortterm in nature. In addition, flow reductions will also cause reductions in benthic invertebrates as well as reduce the rate of benthic invertebrate drift. In the long-term, given restoration of stream channels and increased channel complexity, we expect the quantity of invertebrates to be similar to, or greater than current conditions. Predicted concentrations of arsenic and mercury are expected to impact prey quality. As previously described, arsenic concentrations are predicted to substantially decrease over the life of the project; therefore, we believe accumulation of arsenic in prey tissues will also decrease over time. The proposed action will result in slightly higher mercury concentrations in streams inhabited by salmonids. In addition, the SGP will produce an incremental increase in organic carbon content in the EFSFSR due to domestic sanitary wastewater effluent. Taken together, there will be an increased risk of bioaccumulation of mercury in salmonid prey items. Because mercury methylation is dependent upon a myriad of factors, it is not possible to quantify the potential decrease in prey quality associated with mercury accumulation.

While the proposed action will negatively impact some PBFs, as summarized above, implementation of the proposed action is expected to positively impact many physical aspects of DCH in the long term, including eliminating chronic sources of sediment delivery, removing fish passage barriers, increasing habitat complexity, restoring floodplain connectivity and function, and improving RCA conditions. These actions directly address the passage barrier, elevated sediment, and degraded riparian condition limiting factors identified in the recovery plans.

The Stibnite portion of the project will address the road maintenance, road resurfacing, and elimination of passage barriers called for in NMFS' most recent 5-Year Review (NMFS 2022c) for SR Basin steelhead in the EFSFSR, benefits realized during the pre-construction phase of the SGP. Other recommendations, including improving water quality by reclaiming abandoned mine sites; access road obliteration; improving riparian conditions in disturbed areas; and restoring floodplains will begin early in the pre-construction phase but will not be fully realized until closure and restoration elements of the project have been fully implemented. The SGP will similarly address population-specific habitat concerns (i.e., fine sediment, water quality, and passage barriers) for the SR Spring/summer Chinook salmon SFSR MPG, specifically the EFSFSR and SFSR populations. Addressing chronic sources of chemical contamination and fine sediment, while restoring historical access to complex, instream habitat where streams interact with a fully functioning floodplain, is expected to lead to localized, long-term improvements to spawning, incubation, and rearing habitat for the EFSFSR Chinook salmon population and the SFSR populations of Chinook salmon and steelhead. These improvements will locally benefit DCH PBFs long term, but will not be at the scale to fully address habitat issues and threats at the designation scale.

In the Lemhi River, the restoration project has been designed to specifically address degraded habitat conditions, limiting factors, and threats identified for this stream reach. This element of the proposed action will reestablish connectivity with the river's floodplain, creating multi-thread channels, reducing width-to-depth ratios, stabilizing streambanks, and will include a vegetation plan to restore riparian habitat in the long term. Floodplain connectivity will be restored across a 7,000 linear ft. section of the Lemhi River, locally restoring DCH in this stream reach by MY -1. Although these activities will not address these limiting factors and threats to DCH at the designation scale, they will locally benefit DCH for Lemhi River populations by improving habitat conditions locally at the river reach scale.

The SGP project will occur at a previously disturbed mine site where anadromous fish access has been anthropogenically blocked for almost 80 years. The upper EFSFSR subwatershed is currently degraded as a result of past mining activities. The proposed action will both positively and negatively impact PBFs of DCH as previously described, and the Lemhi restoration effort was incorporated into the proposed action as a means for offsetting the unavoidable adverse effects that will occur in the upper EFSFSR. When considering the status of the critical habitat, environmental baseline, effects of the action, and cumulative effects as discussed above, NMFS concludes that implementation of this proposed action will not appreciably diminish the value of DCH as a whole for the conservation of both species.

## **2.12.2. Species**

As described in Section 2.2, individuals belonging to two MPGs (i.e., SFSR and Upper Salmon) and three populations (i.e., SFSR, EFSFSR, and Lemhi) within the SR Spring/summer Chinook salmon ESU; and individuals belonging to two populations (i.e., SFSR and Lemhi) within the Salmon River MPG of the SR Basin steelhead DPS use the action area to fully complete the

migration, spawning, and rearing parts of their life cycle. The SR Spring/summer Chinook salmon ESU is currently at a high risk of extinction. Similarly, the SR Basin steelhead DPS is not currently meeting its VSP criteria and is at a moderate risk of extinction. Since the last 5-year review, there has been a substantial downturn in adult abundance for both species. This downturn is thought to be driven primarily by marine environmental conditions and a decline in ocean productivity. Very large improvements in abundance will be needed to bridge the gap between the current status and proposed status for recovery for many of the ESU/DPS component populations.

The regional tributary habitat strategy set forth in the final recovery plans (NMFS 2017) is to protect, conserve, and restore natural ecological processes at the watershed scale that support population viability. Ongoing actions to support recovery of these two species include, but are not limited to, conserving existing high-quality habitat and restoring degraded (and maintaining properly functioning) upland processes to minimize unnatural rates of erosion and runoff. Natal habitat recovery strategies and actions for populations within the action area include: (1) reduce road-related impacts (e.g., sediment delivery) on streams; (2) inventory stream crossings and replace any that are barriers to passage; (3) reduce floodplain and channel encroachment; and (4) restore floodplain function.

The environmental baseline incorporates effects of restoration actions implemented to date. It also reflects impacts that have occurred as a result of land management and implementation of various programmatic activities. In addition, impacts from existing state and private actions are reflected in the environmental baseline. Cumulative effects from state and private actions in the action area are expected to continue into the future and are unlikely to be substantially more severe than they currently are. The environmental baseline also incorporates the impacts of climate change on both the species and the habitat they depend on. Several of the ongoing habitat issues that impact VSP parameters, in particular, increased summer temperatures and decreased summer flows, will continue to be affected by climate change.

All three populations of SR spring/summer Chinook salmon occupying the action area are at a high risk of extinction. NMFS' preferred recovery scenario for the SR spring/summer Chinook salmon ESU targets the SFSR and Lemhi populations to achieve viable status, and the EFSFSR population to be maintained. Both populations of SR Basin steelhead are at a moderate risk of extinction. The preferred recovery scenario for the SR Basin steelhead DPS targets both the SFSR population and the Lemhi population be viable. Within the action area, the most heavily used Chinook salmon and steelhead spawning/rearing habitat occurs in Johnson Creek, Burntlog Creek, the EFSFSR, Sugar Creek, and the Lemhi River. Meadow Creek is also used for Chinook spawning and rearing, but only in years in which surplus hatchery fish are stocked above YPP Lake. SR Basin steelhead have not had access to historical spawning and rearing in the upper EFSFSR or Meadow Creek since around 1938. In order to achieve these preferred recovery scenario goals, it is vitally important to preserve habitat conditions that are FA and improve habitat conditions that are FAR or FUR.

Implementation of the Stibnite portion of the proposed action is expected to adversely impact Chinook salmon and steelhead as a result of fish salvage efforts and impacts to their habitat (i.e., reduced water quantity, chemical contamination, increased water temperature, increased

sediment delivery, and reduced forage. It will also improve fish habitat in the long term, through localized improvements to water quality, chronic sediment delivery, fish passage, floodplain connectivity, riparian vegetation, and instream habitat complexity.

The proposed action includes construction activities, with fish salvage employed to reduce effects on ESA listed fishes. The effects associated with construction, and the resultant fish salvage, will lead to mortality of ESA-listed salmonids in the action area. Given the inwater work window and application of the NMFS electrofishing guidelines under which electrofishing will not be done in the vicinity of redds or spawning adults, these effects will be on rearing juveniles. Additionally, these effects will be confined to the project area and will mostly occur in MYs – 3 to 3. Based on our estimates of the number of juvenile fish that will be killed and considering these effects will be spread across multiple years, fish salvage is unlikely to reduce the number of returning adult SR spring/summer Chinook salmon or SR Basin steelhead.

Flow-related effects are likely to impact fish production and population productivity while the mine is in operation, with smaller impacts for approximately 57 years after mining ceases. The proposed action will reduce flow in the occupied SR spring/summer Chinook salmon and SR Basin steelhead habitat for MYs -2 - 20, with smaller reductions during MYs  $21 - 78$ . These reductions in flow will reduce productivity by an average of 1.1%, 0.007%, and 0.36%, respectively, for the EFSFSR Chinook salmon, SFSR Chinook salmon, and in the SFSR steelhead populations, for MYs  $-2 - 20$ . From MYs 21 - 78, the reduction in flow will reduce productivity of the EFSFSR Chinook salmon, SFSR Chinook salmon, and in the SFSR steelhead populations by 0.13%, 0.001% and 0.082%, respectively.

Discharge of mine-contact water, either through point sources or nonpoint sources is expected to alter water quality in the project area, with arsenic, antimony, copper, and mercury being the contaminants of greatest concern. These contaminants are expected to elicit sublethal effects (e.g., reduced growth, tissue damage, altered behavior, reduced olfaction) in fish that develop or rear in project area streams through either waterborne or dietary exposures. There is some risk that early life stage fish (incubation through early exogenous feeding) may experience low levels of mortality (<1 percent) if a redd is continually exposed to predicted maximum concentrations of antimony in Meadow Creek during the early years of operation. Similarly, it is possible that early life stage fish may experience some mortality (< 1 percent) if a redd is continually exposed to predicted maximum concentrations of arsenic in the EFSFSR downstream of the YPP during some years in early closure. It is reasonable to conclude that implementation of the proposed action is reducing the risk of mortality of early life stage fish (though not eliminating that risk) because arsenic and antimony concentrations will be reduced relative to baseline conditions. The proposed action is expected to cause incremental increases in total mercury in the water column of streams inhabited by salmon and steelhead. These increases could lead to greater bioaccumulation of mercury in fish tissues. Additionally, an incremental increase in organic carbon content in the EFSFSR due to sanitary wastewater effluent could increase in methylation potential in the EFSFSR. Bioaccumulation of mercury in fish tissue is not expected to reach lethal levels; however, some fish may experience sublethal effects. The degree to which these sublethal effects may manifest to reduced survival is unknown. Baseline data (presence of redds, juvenile Chinook salmon, and juvenile steelhead) in the EFSFSR, Sugar Creek, and Meadow Creek (Chinook salmon only due to outplanting efforts) indicates spawning, rearing, and

migration currently occurs in streams with elevated arsenic and detectable, and in some cases elevated, concentrations of total mercury. When considering this, coupled with the projected incremental increases in mercury in habitat that is or will be occupied and the predicted decreases in antimony and arsenic concentrations in these streams, we do not believe these potential low levels of mortality or sublethal effects will reduce abundance or productivity in the population throughout the life of the project.

Stream temperature will be negatively impacted in Meadow Creek and the EFSFSR at different spatial and temporal scales. In order to support adult migration and spawning, incubation, and juvenile rearing, we believe 7DADM water temperatures should be below 18, 13, and 16°C, respectively. These temperature thresholds are predicted to be exceeded in Meadow Creek and the EFSFSR above the Sugar Creek confluence. Meadow Creek will experience the greatest temperature increase for the greatest amount of time. In this reach (from its mouth to the TSF), anadromous salmonids may be impacted as follows: (1) some Chinook salmon embryo mortality is expected to occur in redds that are constructed early in the season; (2) some adult Chinook salmon, may experience pre-spawn mortality or reduced gamete viability if they hold in water temperatures exceeding 18°C; (3) and some rearing juvenile Chinook salmon and steelhead may experience sublethal effects such as reduced growth or increased disease risk. In the EFSFSR, there may be a few years where some Chinook salmon embryo mortality occurs in redds that are constructed early in the season, and some rearing juvenile Chinook salmon and steelhead may experience sublethal effects. Taken together, these reaches of Meadow Creek and the EFSFSR represent a small fraction of the overall habitat available for spawning and rearing. Both juvenile and adult fish may seek alternative, nearby habitats with more suitable thermal conditions. For these reasons, any impacts to individuals from exposure to elevated temperatures are not expected to cause population-level reductions in productivity nor will it diminish spatial structure.

Sediment generating activities will include frequent, sporadic turbidity effects to water quality across action area streams, with those increases occurring most frequently occurring during runoff events and resulting in localized deposition of fine sediment in action area stream channels. Although these effects will likely temporarily affect fish behavior (particularly during runoff events), they are unlikely to reach levels severe enough to result in harm. Localized deposition of fine sediment in action area streams has the potential to decrease spawning gravel suitability and decrease benthic invertebrate production within gravel riffles, potentially impacting spawning/incubation and rearing/feeding life stages of Chinook salmon and steelhead. However, spawning substrates are generally FA in action area streams, and GRAIP-Lite modeling suggests that the amount of sediment delivered to action area streams will decrease in comparison to the baseline condition. With application of proposed sediment control BMPs, the deposition of sediment in action area waters is predicted to be measurable but not severe. It is not expected to affect spawning success in the temporary or short-term; and following closure of the mine, obliteration of access roads, and restoration efforts in the EFMC to address chronic sources of sediment delivery, the proposed action is expected to result in a substantial decrease in improved spawning conditions in Meadow Creek and the EFSFSR.

The quantity and quality of forage available to salmonids will be reduced in localized areas at the Stibnite site due to channel relocation, changes in water quality and quantity, and clearing of

riparian vegetation. The quantity of forage will be reduced downstream of channels that are dewatered or relocated due to the loss of wetted habitat and loss of riparian vegetation. Revegetation efforts are expected to forage reductions in localized areas are temporary or shortterm in nature. In addition, flow reductions will also cause reductions in benthic invertebrates as well as reduce the rate of benthic invertebrate drift. In the long-term, given restoration of stream channels and increased channel complexity, we expect the quantity of invertebrates to be similar to, or greater than current conditions. Short-term losses of forage are expected to increase competition among rearing juveniles. Predicted concentrations of arsenic and mercury are expected to impact prey quality. As previously described, arsenic concentrations are predicted to substantially decrease over the life of the project; therefore, we believe accumulation of arsenic in prey tissues will also decrease over time. The proposed action will result in slightly higher mercury concentrations in streams inhabited by salmonids. In addition, the SGP will produce an incremental increase in organic carbon content in the EFSFSR due to domestic sanitary wastewater effluent. Taken together, there will be an increased risk of bioaccumulation of mercury in salmonid prey items. Because mercury methylation is dependent upon a myriad of factors, it is not possible to quantify the potential decrease in prey quality associated with mercury accumulation.

The Lemhi Restoration portion of the proposed action is proposed as compensatory mitigation for stream and wetland impacts of the SGP, with the intent of improving habitat conditions in the Lemhi River basin. The project will enhance habitat conditions across approximately 7,000 ft. of the Lemhi River channel; increasing habitat quality and complexity; reducing channel W:D; enhancing floodplain connectivity; improving instream structure and velocity; increasing pool quantity and complexity; facilitating surface/groundwater interchange for temperature moderation; providing instream cover for fish; and establishing a riparian corridor for shade, cover, and bank stability. These largely beneficial effects will begin to take effect by MY-2, benefitting the Lemhi populations of SR spring/summer Chinook and SR Basin steelhead before active mining begins at the SGP. The potential adverse effects of this element of the proposed action on ESA-listed salmon and steelhead and their critical habitat can be broadly categorized into effects from fish salvage, and temporary habitat-related effects to water quality (i.e., turbidity, chemical contamination) and habitat (i.e., sedimentation of spawning gravels). In the long term, reconnecting the Lemhi River to its floodplain is expected to improve habitat complexity and will improve habitat conditions across the reach, likely leading to localized improvements in Lemhi River Chinook salmon and steelhead population abundance and productivity.

NMFS expects that Perpetua and their and contractors will implement the proposed action as proposed, with full adherence to the design criteria and the EDFs. NMFS also expects that we will be involved in the IARB process to review project elements with potential effects to ESAlisted species or critical habitat to ensure that: (1) no changes to the proposed action occur that will result in effects not previously considered in the BA or this opinion; and (2) the assumptions underlying the analysis of this opinion remain valid. Given this, we expect that adverse effects to ESA-listed species will be avoided or minimized and that unavoidable adverse effects are appropriately addressed. As described in the Effects of the Action (Section 2.5), potential impacts to water quality, food/forage, sedimentation, and effects from dewatering and fish handling/salvage all have the potential to harm individual SR Spring/summer Chinook salmon

and SR Basin steelhead. The majority of these effects are expected to be sublethal. A few individual embryos may experience mortality if exposed to elevated concentrations of arsenic or antimony or if redds are constructed early in the spawning season when temperatures are elevated. Mortality within the project area is expected to primarily occur as a result of fish salvage activities during dewatering events associated with stream crossing structures and stream channel restoration efforts.

NMFS quantified the effects of flow alterations and fish salvage. These quantified effects, expressed as adult returns, will range from slightly positive during MY 2 (due to increased flow from discharge of treated contact water) to a reduction of approximately eight returning adult SR spring/summer Chinook salmon, and eight returning adult SR Basin steelhead, in MY 7. After MY 7, the effects steadily decrease to approximately one returning adult, for each species, in MY 19 and 20. After mine closure, long-term effects on flow, and fish, will result from the filling of the West End Pit Lake, and will reduce adult returns by less than one SR Snake River Chinook salmon and one SRB steelhead, per year.

Should trap and haul become necessary, NMFS expects that up to one more adult Chinook salmon and steelhead is expected to be killed annually through this process. These losses would most likely occur from MY -1 to MY 11, years in which the tunnel will be in operation.

For the Lemhi portion of the project, fish salvage could result in the loss of up to two returning steelhead, but is not likely to result in the loss of more than one adult equivalent Chinook salmon. These effects would occur during one year only.

We were unable to estimate the number of juveniles that might be affected by turbidity, or from disturbance associated with blasting or heavy equipment operation. However, given proposed EDFs and BMPs, we do not expect any more than behavioral responses, and do not otherwise expect any harm to ESA-listed salmonids. Other than when installing bypass structures, no inwater work will occur, with inchannel work occurring only in dewatered work areas. Turbidity pulses and plumes will occur as a result of inchannel work; road construction, reconstruction, and use; and mine and exploratory operations. But, erosion control BMPs and turbidity monitoring are expected to ensure that turbidity levels do not reach a magnitude or persist long enough to result in more than behavioral responses by salmonids. The magnitude of sediment increases and its impact on fish spawning, incubation, rearing is also difficult to predict; however, implementation of erosion control EDFs should effectively minimize the amount of sediment being delivered, and should improve conditions as project area vegetation becomes reestablished and chronic sediment sources are addressed. Considering the minimum amount of fish handling that will occur, spread across time and populations, and considering the short duration of adverse effects associated with elevated sediment delivery, we do not expect climate change to amplify any of these adverse effects. Juvenile fish passage will be temporarily impaired in discrete locations; however, long-standing fish barriers will be removed, which should locally improve overall conditions and population spatial structure. We are unable to estimate the number of juvenile salmonids that may experience sublethal effects as a result of waterborne or dietary exposures to contaminants or elevated temperatures. Implementation of the EDFs, monitoring water quality and biological conditions, and adaptively managing the project

in response to monitoring data will avoid and minimize future reductions of water quality and associated adverse effects to ESA-listed species.

Sediment introduced into and subsequently deposited in the EFSFSR, Johnson Creek, their tributaries; or the Lemhi River, as a result of project implementation, is not expected to reduce the current productivity of the EFSFSR, SFSR, and Lemhi River Chinook salmon populations; or the SFSR and Lemhi River steelhead populations. This is primarily because: (1) turbidity pulses are expected to be short-lived (lasting only a matter of minutes to hours) and small in both magnitude and their downstream extent; (2) sediment will not be delivered to streams simultaneously, rather sediment will be delivered over discrete periods of time (e.g., during rainstorms following ground-disturbing activities or during channel re-watering; and (3) sources of sediment will be dispersed along the stream network so not all of the sediment will end up in a single location within the stream channel. Our assessment assumes the USFS, USACE, and any applicant or contractor will properly implement appropriate EDFs and BMPs during project implementation.

The effects of construction, fish salvage, and reduced flow, due to the proposed action, will result in the annual loss of one to eight adult EFSFSR Chinook salmon equivalents and one to eight adult SFSR steelhead equivalents, during MY -1 through 20. These losses will reduce abundance and productivity of these two populations by 1.1% and 0.36%, slightly increasing extinction risk through MY 20. Facilitating passage past the YPP will slightly improve spatial structure, slightly reducing risk for both populations. Overall, the proposed action will have a negligible effect on extinction risk of the EFSFSR Chinook salmon and the SFSR steelhead populations during MY -1 through 20. After MY 20, average annual losses will be less than one half of a returning adult EFSFSR Chinook salmon and less than one returning adult SFSR steelhead, which will not appreciably reduce population abundance or productivity of either population. The flow effects of the proposed action will reduce size and productivity of the SFSR Chinook salmon population by less than 0.01% during MY -1 through 20, and by less than 0.0001% after MY 20, which will not appreciably increase the extinction risk for that population. Because these impacts will not appreciably increase the extinction risk for the affected populations, the viabilities of the MPGs and the ESU/DPS are not likely to be appreciably reduced.

The SGP project will occur at a previously disturbed mine site where anadromous fish access has been anthropogenically blocked for almost 80 years. The proposed action will locally improve spatial structure, abundance, and productivity by improving habitat in the mine area footprint, restoring floodplain connectivity, habitat complexity, improving water quality, and restoring access to new spawning and rearing habitat for both Chinook salmon and steelhead. The Lemhi restoration effort incorporated into the proposed action as a means for offsetting the unavoidable adverse effects and that will occur in the upper EFSFSR, further benefitting abundance and productivity of the SR spring/summer Chinook salmon ESU, and the Salmon River MPG of the SR Basin steelhead DPS. When considering the status of the species, and adding in the environmental baseline, and cumulative effects, implementation of the proposed action will not appreciably reduce the likelihood of survival and recovery of SR Spring/summer Chinook salmon or SR Basin steelhead.

### **2.13. Conclusion**

After reviewing and analyzing the current status of the listed species and critical habitat, the environmental baseline within the action area, the effects of the proposed action, the effects of other activities caused by the proposed action, and the cumulative effects, it is NMFS' biological opinion that the proposed action is not likely to jeopardize the continued existence of SR Spring/summer Chinook salmon, SR Basin steelhead, or destroy or adversely modify their designated critical habitat.

## **2.14. Incidental Take Statement**

Section 9 of the ESA and Federal regulations pursuant to section 4(d) of the ESA prohibit the take of endangered and threatened species, respectively, without a special exemption. "Take" is defined as to harass, harm, pursue, hunt, shoot, wound, kill, trap, capture or collect, or to attempt to engage in any such conduct. "Harm" is further defined by regulation to include significant habitat modification or degradation that actually kills or injures fish or wildlife by significantly impairing essential behavioral patterns, including breeding, spawning, rearing, migrating, feeding, or sheltering (50 CFR 222.102). "Harass" is further defined by interim guidance as to "create the likelihood of injury to wildlife by annoying it to such an extent as to significantly disrupt normal behavioral patterns, which include but are not limited to, breeding, feeding, or sheltering." "Incidental take" is defined by regulation as takings that result from, but are not the purpose of, carrying out an otherwise lawful activity conducted by the Federal agency or applicant (50 CFR 402.02). Section 7(b)(4) and section 7(o)(2) provide that taking that is incidental to an otherwise lawful agency action is not considered to be prohibited taking under the ESA if that action is performed in compliance with the terms and conditions of this ITS.

## **2.15. Amount or Extent of Take**

The proposed action is reasonably certain to result in incidental take of ESA-listed species. NMFS is reasonably certain the incidental take described here will occur because adult and juvenile Chinook salmon and steelhead will be present in the action area during project implementation, and those fish will be exposed to effects of the proposed action. In some instances, NMFS is able to quantify the amount of take; however, where available information precludes our ability to quantify take, we use surrogates to describe the incidental take pursuant to 50 CFR 402.14 (I). The surrogates used to describe incidental take are explained in sections 2.1.5.1 through 2.5.1.3.

## **2.15.1. Construction and Fish Salvage**

In this opinion, NMFS determined that incidental take is reasonably certain to occur due to construction and associated fish salvage activities. Disturbance associated with equipment operation (i.e., from noise), blasting and drilling (i.e., from noise and vibration) is likely to harass rearing juvenile salmonids (i.e., annoying juveniles sufficiently to disrupt normal behavioral patterns). As described in the species effects analysis, NMFS is unable to quantify the take associated with disturbance due to equipment operation in close proximity to action area streams. It is not possible to tell whether fish are present and have been disturbed, and it is not possible to determine how many, if any, juvenile fish are subject to predation as a result of these activities.

Equipment operation will produce noise and vibration at sufficient intensities to cause behavioral modifications. The degree that juvenile salmon and steelhead will be harassed is directly correlated with both the amount of work done and the proximity of that work to streams. For these reasons, we are not able to identify an amount, extent, or surrogate of take associated with general equipment operation. However, we have defined a surrogate for blasting and drilling. When working alongside streams or lakes occupied by ESA-listed Chinook salmon or steelhead, NMFS will consider take exceeded if: (1) blasting occurs closer than 239-ft. on 20-ft. benches, or closer than 419 ft. on 40-ft. benches; or (2) if drilling occurs closer than 100 ft. Although these surrogates could be considered coextensive with the proposed action, they function as effective reinitiation triggers because our analysis is based upon impacts resulting from such measures, they can be readily monitored, and thus will serve as a regular check on the proposed action.

Similarly, take caused by the increased sediment delivery into action area streams cannot be accurately quantified as number of fish for a variety of reasons. The distribution and abundance of fish within the action area is dependent upon a number of environmental factors that vary over time and space, potentially including exposure of both juvenile and adult salmon and steelhead to resulting turbidity plumes. It is not possible to monitor the number of fish that may be displaced by turbidity plumes. In these circumstances, NMFS can use the causal link established between the activity and the likely changes in habitat conditions affecting the listed species to describe the extent of take as a numerical level of habitat disturbance.

The best available indicators for the extent of take caused by increased sediment delivery is the magnitude and extent of turbidity plumes in the receiving waters during project implementation. The magnitude and extent of the turbidity plume is proportional to the amount of harm that the proposed action is likely to cause through short-term degradation of water quality and instream habitat. Sediment levels are expected to rapidly peak and then steadily decrease in intensity within 1,000 feet downstream of construction areas that are immediately adjacent to or within the stream channel. Although we recognize the limitations of using turbidity as a surrogate for suspended sediment, monitoring for turbidity is a reasonable and cost-effective measure that can be readily implemented in the field. Most of the time turbidity measurements take 30 seconds, can be done on site, and therefore allow for rapid adjustments in project activities if turbidity approaches unacceptable levels. For these reasons, we have chosen turbidity as a surrogate for incidental take from sediment-related effects.

NMFS will consider the extent of take exceeded if turbidity readings, taken approximately 1,000 feet downstream of inwater work areas, reveal turbidity concentrations greater than 50 NTU above background for more than 90 minutes, or 100 NTUs instantaneously. Literature reviewed in Rowe et al. (2003) indicated that NTU levels below 50 generally elicit only behavioral responses from salmonids thereby making this a suitable surrogate for sublethal incidental take monitoring. This take indicator functions as effective reinitiation trigger because it can be readily monitored, and thus will serve as a regular check on the proposed action.

NMFS estimated a quantity of fish potentially handled/harassed during fish salvage. NMFS will consider the amount of take exceeded if:

- At Stibnite, during the construction phase of the project (MYs -3 to -1), more than 819 juvenile Chinook salmon captured via electrofishing, or more than 82 juveniles are killed; or more than 8 juvenile steelhead are captured via electrofishing, or more than one juvenile steelhead is killed.
- At Stibnite, during the operations and closure period (MYs 0 to 23+), more than 1,312 juvenile Chinook salmon are captured via electrofishing, or more than 131 juveniles are killed.
- At Lemhi, if more than 224 juvenile Chinook salmon are captured via electrofishing, or more than 22 juveniles are killed; or more than 560 juvenile steelhead are captured via electrofishing, or more than 56 juveniles are killed.

NMFS will also consider the extent of take exceeded if trap and haul activities handle more than 100 adult Chinook salmon or 100 adult steelhead in any given year, or if more than one of either species is killed by trap and haul in any given year. These take indicators function as effective reinitiation triggers because they can be readily monitored, and thus will serve as a regular check on the proposed action.

#### **2.15.2. Water Diversion, Use, Management, and Drainage Area Reduction**

Water diversion, water use and management, and long-term reduction in drainage area available for streamflow is reasonably certain to result in incidental take of ESA-listed species. NMFS determined that incidental take is reasonably certain to occur as follows: (1) the proposed action will reduce flow, within the project area in Meadow Creek, Sugar Creek, and the EFSFSR; and downstream from the project area in the EFSFSR and SFSR; (2) the affected habitat is occupied by SR spring/summer Chinook salmon and SR Basin steelhead; (3) reducing flow in occupied SR spring/summer Chinook salmon and SR Basin steelhead habitat will reduce growth and survival of rearing Chinook salmon and steelhead; (4) juvenile SR spring/summer Chinook salmon and SR Basin steelhead will enter the EFSFSR diversion and will have to exit via the juvenile bypass system, which will delay downstream migration and increase mortality. The take exempted by this ITS is the loss of SR spring/summer Chinook salmon and SR Basin steelhead from these circumstances. NMFS has quantified an average annual reduction in production, due to the proposed action, of 416 Chinook salmon outmigrants and 168 steelhead outmigrants, for MYs -2 through MY 20, and 46 Chinook salmon outmigrants and 38 steelhead migrants after MY 20. However, changes in production cannot be monitored sufficiently to ensure that amount and extent of take is not exceeded. This is because: (1) production estimates for SR spring/summer Chinook salmon and SR Basin steelhead are based on redd count data and LGD adult return data, respectively, and therefore lack the precision needed to monitor changes of the scale anticipated due to the proposed action; (2) population density of SR spring/summer Chinook salmon and SR Basin steelhead varies greatly from year to year; (3) fish harmed due to increased environmental stress caused by the reductions in flow would be difficult to distinguish from fish harmed due to environmental stress that normally occurs or that is caused by baseline actions; and (4) counting juvenile fishes entering or exiting water diversions is not practicable. Even if take that occurred within the action area could be adequately quantified, monitoring total take due to the proposed actions would still not be feasible because some mortality due to effects of the proposed action in the action area is likely to occur during the downstream migration or in the estuary. This is because fish growth is related to streamflow (Harvey et al. 2006; Davidson et al. 2010), so reducing streamflow in rearing habitat likely reduces the size of downstream migrating smolts. Smaller smolts have higher mortality outside of the natal tributaries (Zabel and Achord 2004), which results in lower smolt-to-adult return rates.

When take cannot be adequately quantified, NMFS describes the extent of take through the use of surrogate measures of take that would define the limits anticipated in this Opinion. As established above in Section 2.10.1.2, net reduction of streamflow due to water diversion, water use and management, and reduction in drainage area, will result in most of the take due to the proposed action; and a small amount of take will occur due to entrainment in the EFSFSR surface water diversion. Presence and condition of the EFSFSR diversion fish screen is relatively easy to ascertain and, as quantifiable habitat indicators, the amount of water flowing in Meadow Creek, Sugar Creek, and the EFSFSR; the net effect of the proposed action on flow (i.e., the amount diverted minus the amount released); and the reduction in drainage area can be accurately measured. In this case, the extent of take will be described as: (1) the amount of water flowing in the EFSFSR at the POD; (2) the amount of water flowing in Meadow Creek from the confluence of Meadow Creek and Blowout Creek to the confluence of Meadow Creek and the EFSFSR; (3) the amount of water flowing in the EFSFSR below the confluence of the EFSFSR and Sugar Creek; (4) the net effect of water diversion and water discharge on streamflow; and (5) the size of the WEP Lake drainage area. The extent of take exempted by this ITS would be exceeded if: (1) flow is less than 3.0 cfs in Meadow Creek from the confluence of Meadow Creek and Blowout Creek to the confluence of Meadow Creek and the EFSFSR when flow is diverted from any of the groundwater wells in Section 15, Township 18N, Range 9E; (2) water is diverted from the EFSFSR surface water diversion when flow at that POD is less than 7.25 cfs from June 30 to September 30, or is less than 5.0 cfs from October 1 to June 29; (3) the average monthly net flow reduction (diversion – discharge) exceeds the amounts in columns 2-5 of Table 60 for any month during MYs -1 through 20; (4) net flow reduction in the EFSFSR exceeds 20% of flow present below the confluence with Sugar Creek, when unimpaired flow in the EFSFSR below the confluence with Sugar Creek is less than 25 cfs; (5) the drainage area of the WEP Lake exceeds 185 acres; (6) the EFSFSR surface water diversion is operated without a fish screen and bypass system that meets criteria in NMFS 2022a.

## **2.15.3. Water Quality**

The new point and nonpoint sources of contaminants will impact water quality (contaminant concentrations and temperature) to a degree that is reasonably certain to result in incidental take of ESA-listed species. NMFS determined that incidental take is reasonably certain to occur as follows: (1) the proposed action will alter water quality within the mine site area in Meadow Creek, Sugar Creek, and the EFSFSR; and downstream from the mine site in the EFSFSR; (2) the affected habitat is or will be occupied by SR spring/summer Chinook salmon and SR Basin steelhead; (3) concentrations of copper, arsenic, mercury, and contaminant mixtures will be at levels associated with sublethal adverse effects for salmon and steelhead including, but not limited to: avoidance (adults and juveniles); reduced growth (juveniles); reduced ability to detect and avoid predators or capture prey; (4) mercury loads in West End Creek will substantially increase during operations, adding to the mercury load in Sugar Creek and the EFSFSR, which are already mercury-impaired; (5) stream temperatures are predicted to reach levels that could cause adult Chinook salmon to suffer pre-spawn mortality, reduced gamete viability, delayed or

blocked migration; reduced survival of incubating Chinook salmon embryos; and reduced growth of juvenile Chinook salmon and steelhead rearing in Meadow Creek and the EFSFSR.

NMFS is unable to quantify the amount of take that might occur as a result of these water quality changes because: (1) future conditions are predictions based upon multiple models, each of which has its own series of predictions; (2) the number of ESA-listed fish that spawn or rear in these areas is unknown and is expected to vary annually; (3) the actual exposure of ESA-listed fish to harmful concentrations of chemicals, and the duration of such exposure, is unpredictable; and (4) there is a large degree of variability in effects that could occur if fish were exposed to chemical concentrations, depending on the magnitude and duration of exposure, condition of exposed fish, life stage of fish, and presence of other stressors (e.g., increased temperature, reduced dissolved oxygen, pathogens). Instead of quantifying the amount of take, NMFS has elected to use surrogates for take that consider the extent of potentially occupied habitat where the proposed action may cause increases in chemical concentrations or stream temperatures to levels that are harmful to ESA-listed species or where biological responses indicate take is occurring that otherwise can't be detected by water column concentrations of selected, individual contaminants.

The best available indicators for the extent of take are the magnitude and extent of water quality impacts in the receiving water and associated biological changes (i.e., fish tissue concentrations and macroinvertebrate community composition) during construction, operations, closure, and post-closure. NMFS will use magnitude, frequency, and duration of chemical concentrations and instream temperatures at specific monitoring locations as the basis for our take surrogate related to water quality impacts. In addition, because the harmful effects of individual contaminants or contaminant mixtures are indirectly experienced through bioaccumulation or alterations in the macroinvertebrate communities, NMFS will also use fish tissue concentrations and macroinvertebrate community composition measures as available indicators for the extent of take. These measures are directly related to the amount of harm or harassment that the proposed action is likely to cause. The extent of take exempted by this ITS would be exceeded if:

- (1) Concentrations of arsenic exceed levels predicted at the monitoring locations specified in Table 65 where: (1) Measured, elevated concentrations are not attributable to background concentrations; (2) follow up monitoring indicates elevated concentrations are not outliers; and (3) annual trend analysis indicates increases concentrations. Follow up monitoring shall include the collection of weekly surface water samples at the location for a period of 4 weeks (i.e., 4 sample events). At a minimum, this shall be within 7 days of becoming aware of the exceedance.
- (2) Concentrations of total mercury exceed levels predicted at the monitoring locations specified in Table 66 where: Measured concentrations are not attributable to background concentrations; (2) follow up monitoring indicates elevated concentrations are not outliers; and (3) annual trend analysis indicates increased concentrations. Follow up monitoring shall include the collection of weekly surface water samples at the location for a period of 4 weeks (i.e., 4 sample events). At a minimum, this shall be within 7 days of becoming aware of the exceedance.
- (3) Concentrations of dissolved copper exceed the chronic criterion at the monitoring locations specified in Table 66 where: (1) the chronic criterion is calculated using site-

specific data for pH, temperature, DOC, and hardness collected on the same day and at the same time; (2) measured concentrations are not attributable to background concentrations; (3) follow up monitoring indicates elevated concentrations are not outliers; and (4) annual trend analysis indicates increased concentrations. Follow up monitoring shall include the collection of weekly surface water samples at the location for a period of 4 weeks (i.e., 4 sample events). At a minimum, this shall be within 7 days of becoming aware of the exceedance.

- (4) Stream temperatures exceed predictions (MWMT for July and September) made for reaches that contain the monitoring locations specified for the EFSFSR and Meadow Creek in Table 66.
- (5) Biological monitoring indicates the mine is having adverse effects on aquatic communities in Meadow Creek and the EFSFSR. At a minimum, aquatic community metrics that shall be used to evaluate potential adverse effects are listed below. In the future, alternate metrics may be developed that are more appropriate for evaluating potential impacts from mining activities. As such, NMFS recognizes that the suite of metrics may be adjusted to reflect the best available science, subject to verification by NMFS.
	- Total Taxa Richness
	- Ephemeroptera Taxa Richness
	- Plecoptera Taxa Richness
	- Percent Plecoptera
	- Percent Ephemeroptera
	- Trichoptera taxa Richness
	- Hilsenhoff Biotic Index
- Percent 5 Dominant Taxa
- Metals Tolerance Index
- Intolerant Taxa Richness
- Percent Tolerant Individuals
- $\bullet$  PIBO O/E

#### **Table 66. Monitoring Locations for Evaluating Extent of Incidental Take at The Mine Site.**



#### **2.16. Effect of the Take**

In the opinion, NMFS determined that the amount or extent of anticipated take, coupled with other effects of the proposed action, is not likely to result in jeopardy to the species or destruction or adverse modification of critical habitat.

## **2.17. Reasonable and Prudent Measures**

"Reasonable and prudent measures" refer to those actions the Director considers necessary or appropriate to minimize the impact of incidental take on the species (50 CFR 402.02). NMFS believes that full application of conservation measures included as part of the proposed action, together with use of the RPMs and terms and conditions described below, are necessary and appropriate to minimize the likelihood of incidental take of ESA-listed species due to completion of the proposed action.

The USFS and USACE shall:

- 1. Minimize the potential for incidental take from water quality impacts to streams.
- 2. Minimize the potential for incidental take from water quantity impacts to streams.
- 3. Minimize the potential for incidental take from fish handling and disturbance.
- 4. Ensure completion of a monitoring, and adaptive management reporting program to confirm that the terms and conditions in this ITS were effective in avoiding and minimizing incidental take and ensure incidental take is not exceeded.

#### **2.18. Terms and Conditions**

In order to be exempt from the prohibitions of section 9 of the ESA, the Federal action agency must comply (or must ensure that any applicant complies) with the following terms and conditions. The USFS, USACE, or any applicant has a continuing duty to monitor the impacts of incidental take and must report the progress of the action and its impact on the species as specified in this ITS (50 CFR 402.14). If the entity to whom a term and condition is directed does not comply with the following terms and conditions, protective coverage for the proposed action would likely lapse.

- 1. The following terms and conditions implement RPM 1:
	- a. Apply standard construction practices, including minimizing the amount of surface disturbance and clearly delineating all work zones before starting construction, to minimize the potential to deliver sediment to action area streams.
	- b. Monitor turbidity as proposed, but stop construction activities if turbidity levels 1,000 feet downstream of their source begin to approach 50 NTUs above background or are visible for more than 90 minutes or begin to approach 100 NTUs above background at any time. After stopping the activities, contact NMFS to determine when work can proceed and if additional BMPs need to be employed to further minimize the intensity of remaining plumes to ensure extent of take is not exceeded.
	- c. Initiate a visual turbidity monitoring program if drilling occurs in RCAs. Visual monitoring must occur at least two times during drilling activities at each location. If visible turbidity is present downstream of drilling activities, operations will cease until the source of turbidity can be identified and mitigated.
	- d. Perpetua will ensure water treatment plants are designed, operated, and maintained in a manner that ensures optimal removal of contaminants and adherence to effluent limits.
- e. Perpetua will request IDEQ derive effluent limits using  $5<sup>th</sup>$  percentile hardness levels representative of the season for which effluent limits are derived, even when those levels are lower than the hardness floor identified in the Idaho Water Quality Standards.<sup>14</sup>
- f. Perpetua will monitor the mine contact water treatment plant effluent, for the following contaminants: aluminum, arsenic, antimony, cadmium, copper, cyanide, WAD, dissolved organic carbon, TDS, hardness, iron, lead, mercury, selenium, silver, and zinc. Monitoring will be performed each month there is discharge. The effluent shall also be monitored daily for pH and continuously (15-minute intervals) for temperature and flow rate.<sup>14</sup>
- g. Perpetua will conduct whole effluent toxicity testing (acute and chronic) for the mine contact water treatment plant. WET testing must occur quarterly; however, monitoring frequency may be reduced, with review and verification from NMFS, if tests consistently demonstrate an absence of acute or chronic toxicity.<sup>14</sup>
- h. During operations, when West End Creek is diverted around the West End Pit, Perpetua will treat water in West End Creek prior to discharge to the existing channel below the West End Pit. The objective of treatment is to reduce total mercury concentrations to levels that currently exist as YP-T-6 and to not increase total mercury loading to Sugar Creek during operations.<sup>14</sup>
- i. Develop, in coordination with NMFS, and implement a stream temperature monitoring plan to ensure the assumptions forming the basis of this opinion remain valid and stream temperatures do not exceed the predicted temperatures described for each affected stream reach during construction, and post-closure mine phases. At a minimum, monitoring locations will include those identified in Table 66 as well as locations above and below the sanitary wastewater treatment plant discharge in the EFSFSR.
- j. Develop, in coordination with NMFS, and implement a stream water quality monitoring plan to ensure the assumptions forming the basis of this opinion remain valid and to ensure contaminants do not exceed the predicted concentrations during construction, operations, and post-closure periods at locations identified in Table 66. At a minimum, monitoring locations will include those identified in Table 66.
- k. Develop, in coordination with NMFS, and implement biological monitoring plan to ensure the assumptions forming the basis of this opinion remain valid and to ensure exposures to bioaccumulative contaminants and contaminant mixtures are not having unanticipated impacts. The biological monitoring plan shall include monitoring for mercury accumulation in fish tissue, arsenic accumulation in macroinvertebrate tissues, and macroinvertebrate community composition. At a minimum, monitoring locations will include those identified in Table 66, as well as a monitoring macroinvertebrate communication composition in the Tamarack Creek reference location (MWH-017).

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<sup>&</sup>lt;sup>14</sup> IDEQ is responsible for issuing IPDES discharge permits and may contain more stringent or additional requirements. Ultimately, what is identified in these terms and conditions are the minimum measures NMFS deems necessary for minimizing incidental take of SR spring/summer Chinook salmon and SR Basin steelhead.

- l. The monitoring required as part of 1.i, 1.j, and 1.k will continue through construction, operations, and closure and may cease only upon NMFS agreement that performance objectives are achieved and modeled predictions are not exceeded.
- 2. The following terms and conditions implement RPM 2:
	- a. Perpetua will measure and record the amount of water diverted from all surface and groundwater PODs at the SGP site.
	- b. Perpetua will measure and record the amount of water discharged from all discharge points at the SGP site.
	- c. Perpetua will use data from a and b to calculate the net effect on flow in Meadow Creek, the EFSFSR upstream from Meadow Creek, the EFSFSR between Meadow Creek and the YPP, and the EFSFSR between the YPP and Sugar Creek.
	- d. Perpetua will measure and record flows in Meadow Creek immediately upstream of the confluence of Meadow and Blowout Creeks, in Meadow Creek just upstream from the confluence of Meadow Creek and the EFSFSR, in the EFSFSR at the EFSFSR diversion, in the EFSFSR just upstream from the confluence of the EFSFSR and Sugar Creek, and in Sugar Creek just upstream from the EFSFSR.
	- e. Perpetua will manage water diversion, use, and discharge such that:
		- i. Flow in Meadow Creek will not be reduced below 3.0 cfs between the confluence of Meadow and Blowout Creeks and the confluence of Meadow Creek and the EFSFSR.
		- ii. Flow in the EFSFSR at the POD will not be reduced below 7.25 cfs from June 30 to September 30, or less than 5.0 cfs from October 1 to June 29.
		- iii. Flow in the EFSFSR below the confluence with Sugar Creek will not be reduced by more than 20% whenever unimpaired flows are less than 25 cfs.
		- iv. The diversion of water directly from the EFSFSR shall not exceed 4.5 cfs.
		- v. Recognizing that actual flow reduction may differ from that modeled and presented in Table 60, Perpetua and the USFS shall work with the IARB to review annual water quantity monitoring results to ensure that operations are adjusted, as needed, to be consistent with the effects analysis and extent of take provided in this opinion.
	- f. Perpetua will ensure that a fish screen(s) and bypass system(s) are present on the EFSFSR diversion and meet NMFS (2022a) criteria (or NMFS more recent criteria).
	- g. Perpetua will ensure that the WEP Lake drainage does not exceed 185 acres.
- 3. The following terms and conditions implement RPM 3:
	- a. If trap and haul procedures are used:
		- i. Ensure that personnel responsible for conducting trap and haul operations are trained in the event that if something breaks down, they have the

knowledge to know what to do in response to make quick decisions to prevent fish losses.

- ii. Ensure water temperatures and dissolved oxygen are within desired parameters (NMFS 2000) at all stages of fish capture and handling.
- b. To minimize impingement in block nets left in place for more than one day in Chinook salmon and steelhead spawning and rearing habitat, monitor nets every four hours for fish impingement.
- c. Ensure that the fish tunnel, and all bridges and culverts accessible to ESA-listed Chinook salmon and steelhead meet NMFS most recent design and fish passage guidelines (currently NMFS 2022a), including considerations for climate change.
- d. Ensure all charge limitations, controlled blasting techniques, and blasting setbacks proposed for use in the proposed action are followed for all blasting activities. Any modifications to blasting setback distances based on monitoring at the mine site shall be reviewed and verified by NMFS prior to implementation. Fish salvage will not be used in lieu of charge limitations, controlled blasting techniques, or blasting setbacks.
- 4. The following terms and conditions implement RPM 4:
	- a. Regarding adaptive management, ensure that NMFS is involved in all IARB communications, correspondence, and meetings. NMFS expects that at a minimum, the IARB will be engaged in the review of all documents referenced in section 1.11.1, and those listed below. Furthermore, Perpetua will obtain NMFS' verification prior to finalization and implementation of plans and designs that may affect ESA-listed species.<sup>15</sup> NMFS verification will be subject to its confirmation that effects remain consistent with the effects analysis and extent of take provided in this opinion.
		- i. Water quality and water quantity monitoring reports;
		- ii. Final SPPC and the Hazardous Materials Handling and Emergency Response Plan;
		- iii. SGP Road Maintenance Agreement with Valley County;
		- iv. Annual USFS summary of the implemented and planned road maintenance activities;
		- v. Road maintenance activities requiring more substantial efforts than typical maintenance (e.g., landslide or avalanche recovery), or outside the current road footprint (e.g., instream fill, LWD removal from streams, etc.);
		- vi. Explosives and Blasting Management Plan;

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- vii. Final Design Drawings for stream reaches at the mine site.
- viii. Final Design Drawings Lemhi River Restoration Project.

<sup>&</sup>lt;sup>15</sup> At a minimum, these include the FMP, WRMP, FOMP, Adaptive Management Plans, Closure and Reclamation Plans, atypical road maintenance/repair activities within RCAs, stream designs, surface water diversion intake design, and EFSFSR diversion tunnel design.

- b. Submit an annual project status/completion report to NMFS by March 15 following each year of the proposed action. At a minimum, reports shall identify:
	- i. Inwater work completed, including starting and ending dates for completed work. Site photos taken before and after work should be included and labeled.
	- ii. Records of all fish salvage completed, including the locations and dates of salvage, environmental conditions, equipment settings, number of fish captured, handled, moved, and killed by species.
	- iii. Results and summaries of the effluent quality (including WET testing results), surface water quality, surface water temperature, and biological monitoring performed that year. The annual report shall confirm the extent of incidental take exempted by this opinion was not exceeded.
	- iv. A summary of pollution and erosion control inspection results, including description of any erosion control failure, contaminant release, and efforts to correct such incidences.
	- v. Results of turbidity monitoring to demonstrate the authorized extent of take was not exceeded.
	- vi. Identification of blasting and exploratory drilling locations, including their distance from waters occupied by ESA-listed salmon or steelhead, and time needed to complete drilling at each location.
	- vii. Details of all closure, restoration, and reclamation work completed in previous year.
	- viii. Specific to revegetation efforts, annually submit post-construction revegetation reports documenting progress toward achieving the targeted goals for riparian vegetation and ground cover for other ground disturbing activities within three years of planting. Considering difficulties establishing vegetation in the project area in past rehabilitation efforts, ground cover monitoring and annual updates shall continue until desired targets are reached, and for 5 years thereafter.
	- ix. The report shall provide the above identified information and confirm the project's proposed BMPs and that this opinion's terms and conditions were successfully implemented.
- c. Exceedances of effluent limits or exceedance of predicted instream contaminant concentrations and temperature predictions will be reported to NMFS quarterly.
- d. Reports must be submitted electronically to [NMFSWCR.SRBO@noaa.gov.](mailto:NMFSWCR.SRBO@noaa.gov)

#### **2.19. Conservation Recommendations**

Section  $7(a)(1)$  of the ESA directs federal agencies to use their authorities to further the purposes of the ESA by carrying out conservation programs for the benefit of the threatened and endangered species. Specifically, "conservation recommendations" are suggestions regarding discretionary measures to minimize or avoid adverse effects of a proposed action on listed species or critical habitat or regarding the development of information (50 CFR 402.02).

- 1. The USFS and the applicant should consider developing redundancy as a safety factor for the liner (e.g., a thickening of the clay layer) used on the YPP backfill to better protect against bedload scour and potential failure, better ensuring that surface flows do not go subsurface. The safety factor should be based on a maximum scour depth of the 100-year flood event, factoring in necessary considerations for climate change.
- 2. Considering the proposed action is expected to contribute additional mercury loading to Sugar Creek and the EFSFSR, the USFS should find opportunities and implement actions on USFS land that can reduce mercury loadings from other sources of contamination (e.g., Cinnabar Mine Site or other historic sources of mercury in the EFSFSR watershed).
- 3. Considering the proposed action will increase traffic on roads, the USFS and applicant should consider implementing BMPs to ensure stormwater runoff from roads does not directly drain to stream channels, but rather is directed to vegetated ground where it can infiltrate.

### **2.20. Reinitiation of Consultation**

This concludes formal consultation for the SGP.

Under 50 CFR 402.16(a): "Reinitiation of consultation is required and shall be requested by the federal agency where discretionary federal involvement or control over the action has been retained or is authorized by law and: (1) If the amount or extent of incidental taking specified in the ITS is exceeded; (2) If new information reveals effects of the agency action that may affect listed species or critical habitat in a manner or to an extent not previously considered; (3) If the identified action is subsequently modified in a manner that causes an effect to the listed species or critical habitat that was not considered in the biological opinion or written concurrence; or (4) If a new species is listed or critical habitat designated that may be affected by the identified action."

#### **2.21. "Not Likely to Adversely Affect" Determinations**

The Southern Resident killer whale DPS, composed of J, K and L pods, was listed as endangered under the ESA on November 18, 2005 (70 FR 69903). A 5-year review under the ESA completed in 2021 concluded that Southern Residents continue to face a high risk of extinction and should remain listed as endangered and includes recent information on the population, threats, and new research results and publications (NMFS 2021a).

The limiting factors described in the final recovery plan included reduced prey availability and quality, high levels of contaminants from pollution, and disturbances from vessels and sound (NMFS 2008). This section summarizes the status of SRKWs throughout their range. This section summarizes information taken largely from the recovery plan (NMFS 2008), recent 5 year review (NMFS 2021a), as well as new data that became available more recently.

Critical habitat for the SRKW DPS was first designated on November 29, 2006 (71 FR 69054) in inland waters of Washington State. NMFS published a final rule to revise SRKW critical habitat in 2021 (86 FR 41668; August 2, 2021). This rule, which became effective on September 1, 2021, maintains the previously designated critical habitat in inland waters of Washington (Puget Sound, see 71 FR 69054; November 29, 2006) and expands it to include six additional coastal

critical habitat areas off the coast of Washington, Oregon, and California (about  $15,910 \text{ mi}^2$ ). Critical habitat includes approximately  $2,560$  mi<sup>2</sup> of inland waters of Washington in three specific areas: (1) the Summer Core Area in Haro Strait and waters around the San Juan Islands; (2) Puget Sound; and (3) the Strait of Juan de Fuca. It also includes  $15,910 \text{ mi}^2$  of marine waters along the U.S. west coast between the 20-ft. depth contour and the 656.2-ft. depth contour from the U.S. international border with Canada south to Point Sur, California.

Based on the natural history of SRKWs and their habitat needs, NMFS identified the following PBFs essential to conservation for critical habitat: (1) water quality to support growth and development; (2) prey species of sufficient quantity, quality and availability to support individual growth, reproduction and development, as well as overall population growth; and (3) passage conditions to allow for migration, resting, and foraging. A brief summary of the prey PBFs is presented below. Detailed information on all the PBFs essential to conservation can be found in the 2006 critical habitat final rule (71 FR 69054, November 29, 2006) and the recent 2021 critical habitat expansion final rule (86 FR 41668, August 2, 2021), and the Final Biological Report that supports the 2021 critical habitat rule (NMFS 2021b), which are incorporated here by reference.

Prey species of sufficient quantity, quality and availability are essential to conservation as SRKWs need to maintain their energy balance all year long to support daily activities (foraging, traveling, resting, socializing), as well as gestation, lactation, and growth. Reduced prey availability has been strongly associated with killer whale mortality and to a lesser degree with low fecundity (Ford et al. 2010; Nelson et al. 2024; Ward et al. 2009; Wasser et al. 2017). Most wild salmon stocks throughout the whales' geographic range are at fractions of their historic levels and 28 ESUs and DPSs of salmon and steelhead are listed as threatened or endangered under the ESA, with the Chinook salmon ESUs and DPSs being most important to the SRKW diet. Historically, overfishing, habitat losses, and hatchery practices were major causes of decline for salmonids. Poor ocean conditions over the past two decades have reduced salmon and salmonid populations already weakened by the degradation and loss of freshwater and estuary habitat, fishing, hydropower system management, and hatchery practices. In addition to sufficient quantity of prey, fish need to be accessible and of sufficient quality and size to support SRKW. Contaminants affect the quality of SRKW prey in Puget Sound and in coastal waters of Washington, Oregon, and California. The size of Chinook salmon is also an important aspect of prey quality (i.e., SRKWs primarily consume large Chinook), so changes in Chinook salmon size (for instance as shown by Ohlberger et al. [2018]) may affect the quality of this feature of critical habitat. This shift may be largely due to direct effects from size-selective removal by marine mammals and fisheries, followed by evolutionary changes toward these smaller sizes and early maturation (Ohlberger et al. 2018). Smaller fish have a lower total energy value than larger ones (O'Neill et al. 2014). Therefore, SRKWs need to consume more fish salmon in order to meet their caloric needs as a result of a decrease in average size of older Chinook salmon.

Human activities managed under a variety of legal mandates have the potential to affect SRKW critical habitat PBFs, including those that could increase water contamination and/or chemical exposure, decrease the quantity, quality, or availability of prey, or inhibit safe, unrestricted passage between important habitat areas to find prey and fulfill other life history requirements. Examples of these types of activities include (but are not limited to), in no particular order: (1)

salmon fisheries and bycatch; (2) salmon hatcheries; (3) offshore aquaculture/mariculture; (4) alternative energy development; (5) oil spills and response; (6) military activities; (7) vessel traffic; (8) dredging and dredge material disposal; (9) oil and gas exploration and production; (10) mineral mining (including sand and gravel mining); (11) geologic surveys (including seismic surveys); and (12) activities occurring adjacent to or upstream of critical habitat that may affect essential features, labeled "upstream activities" (including activities contributing to pointsource water pollution, power plant operations, liquefied natural gas terminals, desalinization plants) (NMFS 2021b).

SRKWs have been repeatedly observed feeding off the Columbia River plume in March and April during peak spring Chinook salmon runs (Krahn et al. 2004; Zamon et al. 2007; Hanson et al. 2008; and Hanson et al. 2010). For this reason, the eastern Pacific Ocean, where SRKW overlap with Chinook salmon from the Columbia River basin is included in the action area due to potential impacts on the whale's prey base.

Southern Resident killer whales consume a variety of fish species (22 species) and one species of squid (Ford et al. 1998; Ford and Ellis 2006; Hanson et al. 2010; Ford et al. 2016), but salmon are identified as their primary prey. Southern Residents are the subject of ongoing research, including direct observation, scale and tissue sampling of prey remains, and fecal sampling. Scale and tissue sampling from May to September indicate that their diet consists of a high percentage of Chinook salmon (monthly proportions as high as >90 percent) (Hanson et al. 2010; Ford et al. 2016). The diet data also indicate that the whales are consuming mostly larger (i.e., older) Chinook salmon. Deoxyribonucleic acid (DNA) quantification methods are also used to estimate the proportion of different prey species in the diet from fecal samples (Deagle et al. 2005). Ford et al. (2016) confirmed the importance of Chinook salmon to the Southern Residents in the summer months using DNA sequencing from whale feces. Salmon and steelhead made up to 98 percent of the inferred diet, of which almost 80 percent were Chinook salmon. Coho salmon and steelhead are also found in the diet in spring and fall months when Chinook salmon are less abundant. Specifically, coho salmon contribute to over 40 percent of the diet in late summer, which is evidence of prey shifting at the end of summer towards coho salmon (Ford et al. 1998; Ford and Ellis 2006; Hanson et al. 2010; Ford et al. 2016). Less than 3 percent each of chum salmon (*O. keta*), sockeye salmon (*O. nerka*), and steelhead were observed in fecal DNA samples collected in the summer months (May through September). Prey remains and fecal samples collected in inland waters during October through December indicate that Chinook and chum salmon are primarily contributors to the whales' diet (Hanson et al. 2021). Observations of whales overlapping with salmon runs (Wiles 2004; Zamon et al. 2007; Krahn et al. 2009), and collections of prey and fecal samples have occurred in the winter months. Preliminary analysis of prey remains and fecal samples sampled during the winter and spring in coastal waters indicated that the majority of prey samples were Chinook salmon (80 percent of prey remains and 67 percent of fecal samples were Chinook salmon), with a smaller number of steelhead, chum salmon, and halibut (Hanson et al. 2021). The occurrence of K and L pods off the Columbia River in March suggests the importance of Columbia River spring-run stocks of Chinook salmon in their diet (Hanson et al. 2013) at that time of year. Chinook salmon genetic stock identification from samples collected in winter and spring in coastal waters included 12 U.S. west coast stocks, and over half of the Chinook salmon consumed originated in the Columbia River (Hanson et al. 2021) for the K and L pods (primarily fall-run stocks). Based on genetic

analysis of feces and scale samples, Chinook salmon from Fraser River stocks dominate the diet of Southern Residents in the summer (Hanson 2010).

The proposed action will not have any direct effects on SRKW; however, it may indirectly affect the quantity of prey available to them. As described in the above Opinion and ITS, the proposed action may result in the loss of an average of four adult SR spring/summer Chinook salmon each year for up to 21 years. The ocean range of Snake River spring/summer Chinook salmon (Weitkamp 2010) overlaps with the known range and designated critical habitat of SRKW. The loss of four Chinook salmon each year would minimally reduce the SRKW's available prey base in the Pacific Ocean for up to 21 years.

Given the total quantity of prey available to SRKWs, the reduction in prey due to the proposed action will be extremely small. The above Opinion did not identify any potential for the proposed action to influence the quality (size) and/or quality (contaminant levels) of Chinook salmon to an extent that would cause sublethal effects should SRKW eat prey originating within the action area. NMFS finds that the proposed action will not have anything more than minimal effects on abundance, diversity, or distribution of ESA-listed Chinook salmon, and therefore the effects to the quantity of prey available to the whales in the long term across their vast range is expected to be very small. For these reasons, the proposed action will have an insignificant effect on SRKW, and therefore, NMFS concurs with the action agencies' NLAA determination for SRKW. Likewise, because so few of the SRKW prey will be affected by the action, the effect to the prey base PBF is insignificant, and NMFS concurs with the action agencies' NLAA determination for SRKW DCH.

#### **3. MAGNUSON–STEVENS FISHERY CONSERVATION AND MANAGEMENT ACT ESSENTIAL FISH HABITAT RESPONSE**

Section 305(b) of the MSA directs Federal agencies to consult with NMFS on all actions or proposed actions that may adversely affect EFH. Under the MSA, this consultation is intended to promote the conservation of EFH as necessary to support sustainable fisheries and the managed species' contribution to a healthy ecosystem. For the purposes of the MSA, EFH means "those waters and substrate necessary to fish for spawning, breeding, feeding, or growth to maturity," and includes the physical, biological, and chemical properties that are used by fish (50 CFR 600.10). Adverse effect means any impact that reduces quality or quantity of EFH, and may include direct or indirect physical, chemical, or biological alteration of the waters or substrate and loss of (or injury to) benthic organisms, prey species and their habitat, and other ecosystem components, if such modifications reduce the quality or quantity of EFH. Adverse effects may result from actions occurring within EFH or outside of it and may include direct, indirect, sitespecific habitat-wide impacts, including individual, cumulative, or synergistic consequences of actions (50 CFR 600.810). Section 305(b) of the MSA also requires NMFS to recommend measures that can be taken by the action agency to conserve EFH. Such recommendations may include measures to avoid, minimize, mitigate, or otherwise offset the adverse effects of the action on EFH [CFR 600.905(b)].

This analysis is based, in part, on the EFH assessment provided by the USFS and USACE, and descriptions of EFH for Pacific Coast salmon (PFMC 2014) contained in the fishery

management plans developed by the Pacific Fishery Management Council (PFMC) and approved by the Secretary of Commerce.

# **3.1. Essential Fish Habitat Affected by the Project**

The action area, as described in Section 2.3 of the above opinion, except for areas above natural barriers to fish passage, is also EFH for Pacific Coast Chinook salmon (PFMC 2014). The PFMC designated the following five habitat types as habitat areas of particular concern (HAPCs) for salmon: complex channel and floodplain habitat, spawning habitat, thermal refugia, estuaries, and submerged aquatic vegetation (PFMC 2014). The proposed action may adversely affect the following HAPCs: complex channel and floodplain habitat, spawning habitat, and thermal refugia.

Proper function of complex channels and floodplain habitat could be affected in the EFSFSR and Meadow Creek; spawning habitat could be affected in the EFSFSR, Meadow Creek, and Sugar Creek; and thermal refugia could be affected in the EFSFSR, Meadow Creek, Sugar Creek, Johnson Creek, Burntlog Creek, and East Fork Burntlog Creek.

# **3.2. Adverse Effects on Essential Fish Habitat**

Based on information provided in the BA and the analysis of effects presented in the ESA portion of this document, NMFS determined the proposed action would adversely affect EFH as follows:

- 1. The proposed action will result in ground-disturbing activities, which may adversely affect riparian and instream habitat in Meadow Creek, EFSFSR, and Johnson Creek and its tributaries.
- 2. The proposed action will affect water quality, which may adversely affect the temperature and chemical properties of instream habitat in Meadow Creek, EFSFSR, SFSR, and Johnson Creek and its tributaries.
- 3. The proposed action will reduce streamflow in portions of and tributaries to the EFSFSR; which will increase summer water temperature; reduce amount of habitat available for adult Chinook salmon; reduce food for juvenile Chinook salmon; reduce access to escape cover for juvenile Chinook salmon; reduce movement of sediment in the affected tributaries; and reduce cold water refugia habitat for juvenile rearing and adult holding Chinook salmon.

# **3.3. Essential Fish Habitat Conservation Recommendations**

NMFS determined that the following conservation recommendations are necessary to avoid, minimize, mitigate, or otherwise offset the adverse effects of the proposed action on EFH.

- 1. To address ground disturbing habitat effects to the spawning habitat, and complex channel and floodplain habitat HAPCs:
	- a. Standard construction practices, including minimizing the amount of surface disturbance and clearly delineating all work zones before starting construction,

should be applied to minimize the potential to deliver sediment to action area streams.

- b. Turbidity should be monitored as proposed, but construction activities should stop if turbidity levels 1,000 feet downstream of their source begin to approach 50 NTUs above background or are visible for more than 90 minutes or begin to approach 100 NTUs above background at any time. After stopping the activities, NMFS should be contacted to determine when work can proceed and if additional BMPs need to be employed to further minimize the intensity of remaining plumes to ensure extent of take is not exceeded.
- c. A visual turbidity monitoring program should be initiated if drilling occurs in RCAs. Visual monitoring should occur at least two times during drilling activities at each location. If visible turbidity is present downstream of drilling activities, operations should cease until the source of turbidity can be identified and mitigated.
- d. The USFS and the applicant should consider developing redundancy as a safety factor for the liner (e.g., a thickening of the clay layer) used on the YPP backfill to better protect against bedload scour and potential failure, better ensuring that surface flows do not go subsurface. The safety factor should be based on a maximum scour depth of the 100-year flood event, factoring in necessary considerations for climate change.
- e. The USFS should obtain NMFS review and verification of the following plans and design prior to their finalization and implementation: FMP, WRMP, FOMP, Adaptive Management Plans, Closure and Reclamation Plans, atypical road maintenance/repair activities within RCAs, stream designs, surface water diversion intake design, and EFSFSR diversion tunnel design.
- 2. To address water quality effects to the thermal refugia, spawning habitat, and complex channel and floodplain habitat HAPCs:
	- a. Perpetua should ensure water treatment plants are designed, operated, and maintained in a manner that ensures optimal removal of contaminants and adherence to effluent limits.
	- b. Perpetua should request IDEQ derive effluent limits using  $5<sup>th</sup>$  percentile hardness levels representative of the season for which effluent limits are derived, even when those levels are lower than the hardness floor identified in the Idaho Water Quality Standards. <sup>16</sup>
	- p. During operations, when West End Creek is diverted around the West End Pit, Perpetua should treat water in West End Creek prior to discharge to the existing channel below the West End Pit. The objective of treatment is to reduce total mercury concentrations to levels that currently exist as YP-T-6 and to not increase total mercury loading to Sugar Creek during operations.<sup>16</sup>
	- q. Considering the proposed action will increase traffic on roads, the USFS and applicant should consider implementing BMPs to ensure stormwater runoff from

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<sup>&</sup>lt;sup>16</sup> IDEQ is responsible for issuing IPDES discharge permits and may contain more stringent or additional requirements. Ultimately, what is identified in these terms and conditions are the minimum measures NMFS deems necessary for minimizing incidental take of SR spring/summer Chinook salmon and SR Basin steelhead.

roads does not directly drain to stream channels. Instead stormwater runoff from roads should be directed to vegetated ground where it can infiltrate.

- 3. To address streamflow effects to the thermal refugia, spawning habitat, and complex channel and floodplain habitat HAPCs:
	- a. The permittees should manage water diversion, use, and discharge such that:
		- i. Flow in Meadow Creek will not be reduced below 3.0 cfs between the confluence of Meadow and Blowout Creeks and the confluence of Meadow Creek and the EFSFSR.
		- ii. Flow in the EFSFSR at the POD will not be reduced below 7.25 cfs from June 30 to September 30, or less than 5.0 cfs from October 1 to June 29.
		- iii. Flow in the EFSFSR below the confluence with Sugar Creek will not be reduced by more than 20% whenever unimpaired flows are less than 25 cfs.
		- iv. The diversion of water directly from the EFSFSR should not exceed 4.5 cfs.
		- v. Recognizing that actual flow reduction may differ from that modeled and presented in Table 60, Perpetua and the USFS should work with the IARB to review annual water quantity monitoring results to ensure that operations are adjusted, as needed, to be consistent with the effects analysis in this opinion.
	- b. The permittees should ensure that a fish screen(s) and bypass system(s) are present on the EFSFSR diversion and meet NMFS (2022a) criteria.
	- c. The permittees will ensure that the WEP Lake drainage does not exceed 185 acres.

Fully implementing these EFH Conservation Recommendations would protect, by avoiding or minimizing the adverse effects described in Section 3.2, above, for Pacific Coast salmon.

#### **3.4. Statutory Response Requirement**

As required by section 305(b)(4)(B) of the MSA, the USFS or USACE must provide a detailed response in writing to NMFS within 30 days after receiving an EFH Conservation Recommendation. Such a response must be provided at least 10 days prior to final approval of the action if the response is inconsistent with any of NMFS' EFH Conservation Recommendations unless NMFS and the Federal agency have agreed to use alternative timeframes for the Federal agency response. The response must include a description of the measures proposed by the agency for avoiding, minimizing, mitigating, or otherwise offsetting the impact of the activity on EFH. In the case of a response that is inconsistent with the conservation recommendations, the Federal agency must explain its reasons for not following the recommendations, including the scientific justification for any disagreements with NMFS over the anticipated effects of the action and the measures needed to avoid, minimize, mitigate, or offset such effects [50 CFR 600.920(k)(1)].

### **3.5. Supplemental Consultation**

The USFS or USACE must reinitiate EFH consultation with NMFS if the proposed action is substantially revised in a way that may adversely affect EFH, or if new information becomes available that affects the basis for NMFS' EFH Conservation Recommendations [50 CFR 600.920(l)].

#### **4. DATA QUALITY ACT DOCUMENTATION AND PRE-DISSEMINATION REVIEW**

The DQA specifies three components contributing to the quality of a document. They are utility, integrity, and objectivity. This section of the opinion addresses these DQA components, documents compliance with the DQA, and certifies that this opinion has undergone predissemination review.

### **4.1. Utility**

Utility principally refers to ensuring that the information contained in this consultation is helpful, serviceable, and beneficial to the intended users. The intended users of this opinion are the USFS and USACE. Other interested users could include Perpetua and the agencies of the State of Idaho. Individual copies of this opinion were provided to the USFS and USACE. The document will be available within 2 weeks at the NOAA Library Institutional Repository [\(https://repository.library.noaa.gov/welcome\)](https://repository.library.noaa.gov/welcome). The format and naming adhere to conventional standards for style.

#### **4.2. Integrity**

This consultation was completed on a computer system managed by NMFS in accordance with relevant information technology security policies and standards set out in Appendix III, 'Security of Automated Information Resources,' Office of Management and Budget Circular A-130; the Computer Security Act; and the Government Information Security Reform Act.

## **4.3. Objectivity**

#### *Information Product Category:* Natural Resource Plan

*Standards:* This consultation and supporting documents are clear, concise, complete, and unbiased; and were developed using commonly accepted scientific research methods. They adhere to published standards including the NMFS ESA Consultation Handbook, ESA regulations, 50 CFR 402.01 et seq., and the MSA implementing regulations regarding EFH, 50 CFR 600.

*Best Available Information:* This consultation and supporting documents use the best available information, as referenced in the References section. The analyses in this opinion and EFH consultation contain more background on information sources and quality.

*Referencing:* All supporting materials, information, data and analyses are properly referenced, consistent with standard scientific referencing style.

*Review Process:* This consultation was drafted by NMFS staff with training in ESA and MSA implementation, and reviewed in accordance with West Coast Region ESA quality control and assurance processes.

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## **6. APPENDICES**

## **APPENDIX A**

## **Summary of Environmental Design Features and Protection Measures**

(Relevant Excerpts from BA Appendix B, Stantec 2024)

## **Table A-1. Summary of Environmental Design Features and Protection Measures: Fish - General**


























### **Table A-2. Summary of Environmental Design Features and Protection Measures: Fish – Sediment.**





## **Table A-3. Summary of Environmental Design Features and Protection Measures: Vegetation – General.**







### **Table A-4. Summary of Environmental Design Features and Protection Measures: Vegetation – Noxious Weeds.**



### **Table A-5. Summary of Environmental Design Features and Protection Measures: General Road Use and Maintenance.**





# **Table A-6. Summary of Environmental Design Features and Protection Measures: General – Reclamation and Restoration.**









graminoids, forbs, woody shrubs, and trees, provide a large variety of habitat features including food sources, LWD, and various rooting depths that provide streambank stability. In general, species composition and percent cover of wetland and riparian vegetation profoundly affect wetland ecosystem functions and habitat quality for wildlife. Stream restoration reaches and wetland restoration areas will be revegetated with a variety of native herbaceous and woody species. Seed mixes, live

stakes, and nursery-grown container plants and plugs of native graminoids, forbs, shrubs, and trees will be utilized for revegetation. Plant species for revegetation were chosen based on existing riparian and wetland vegetation observed during surveys of the Project area and reference sites.

Around streams, the width of riparian plantings will be 18 feet. The inner two feet of plantings will consist of riparian species typical to area streambanks and wetlands. These species include waterweed (*Elodea candesis* and *Elodea Nuttalii*), Bolander's quillwort (*Isoetes bolanderi*), alpine pondweed (*Potomogeton alpinus*), ribbonleaf pondweed (*Potomogeton epihydrus*), white water crowfoot (*Ranunculus aquatilis*), and common bladderwort (*Utricularia macrohiza*). These plantings will occur on approximately two-foot centers. The outer 16 feet of plantings will consist of forested wetland vegetation species due to their ability to generate shade for surface water. These species include thinleaf alder (*Alnus incana*), bluejoint reedgrass (*Calamagrostis canadensis*), redosier dogwood (*Cornus sericea*), largeleaf avens (*Geum macrophyllium*), twinberry honeysuckle (*Lonicera involucrate*), Engelmann's spruce (*Picea engelmannii*), prickly currant (*Ribes lacustre*), Drummond's willow (*Salix drummondiana*), Pacific willow (*Salix lasiandra*), slender hairgrass (*Deschampsia elongata*), slender wheatgrass (*Elymus trachycaulus*), and slender cinquefoil (*Potentialla gracilis*). Tree planting will occur on approximately six-foot centers.

The stream design has been developed to approximate a geomorphically appropriate quasi-equilibrium state, while enabling each stream reach to evolve naturally over time in response to changing environmental drivers and potential future disturbances (i.e., fire, climate change, etc.). Built into the SGP stream design are a diversity of treatments and channel prescriptions allowing a certain amount of variability and associated uncertainty in the channel response over time. For example, by using appropriate streambed, bank, and floodplain materials; allowing channels to migrate across appropriately sized floodplains; incorporating horizontal and vertical control at strategic locations; and incorporating bioengineered bank stabilization treatments and revegetation, the design mimics the stability and diversity observed in natural reference streams. This approach provides some amount of resilience to disturbances and sets the Project up for long-term resiliency and sustainability.

Design stream reach EF3 (parts A-D) is the portion of the restored EFSFSR that crosses the backfilled Yellow Pine Pit (YPP). The entirety of the pit backfill will be covered with a geosynthetic liner and the restored stream channel will be constructed in soil cover materials placed on top of another geosynthetic liner that underlies the stream corridor.

Objectives for the channel design will be restoration of fish passage, improvement of spawning and rearing habitat, and improvement of wetland function. The stream channel design parameters were based on four reference sites and were calculated to be:



Stream Design Report, Appendix

D.2











diversion sequencing, operational requirements, segment length, segment slope, flow conditions, depth, and fish salvage (see below). All isolation barriers will be monitored during installation and operation. Partial dewatering will generally be conducted during low-flow periods to facilitate stream segment isolation and fish salvage. Whenever possible, dewatering will not begin until fish have been captured and removed for relocation. However, depending on the location and water depth, it may be necessary to partially draw down the water first to perform fish removal. Partial dewatering before fish salvage operations begin may also improve fish capture efficiency by reducing the total volume of stream habitat that needs to be salvaged. In those cases, dewatering pumps will be screened to meet NOAA Fisheries and IDFG standards to avoid entrainment of juvenile fish. Fish capture from work area isolation will consist of: • slowly reducing flow in the work area to allow some fish to leave volitionally; • installation of block nets upstream and downstream of the isolation area with the nets secured to stream channel bed and banks until fish capture is complete and exclusion of fish from the work area is necessary; • hourly monitoring of block nets during instream disturbance in the work area; • if block nets are in place for more than one day, they will be monitored daily to ensure they are secured to banks and are free of organic accumulation plus monitored every four hours for fish impingement if located in bull trout spawning and rearing habitat (unless a variance is granted by the USFS); seining the isolated area to capture and relocate fish; if areas are isolated overnight, minnow traps will be placed overnight in conjunction with seining; • collecting any remaining fish by hand or dip nets as dewatering continues; and, • if all other techniques have been exhausted, electrofishing may be used to capture remaining fish under electrofishing conservation measures. Captured fish will be relocated as quickly as possible to pre-planned release areas using aerated and shaded transport buckets holding limited numbers of fish of comparable size to minimize predation. Dead fish will not be stored in transport buckets but will be left on the streambank to avoid mortality counting errors. Sediment controls will include the implementation and use of the following as needed in appropriate locations: • Instream work will conform with the work, turbidity, and dewatering procedures as specified in design conservation measures (Rio ASE 2023) and adhere to Bonneville Power Administration Habitat Improvement Program conservation measures; • Placement of fine mesh silt fences and straw waddles; • Minimization of equipment wet crossings with vehicles and machinery crossing at right angles to the main channel whenever possible; • No construction equipment stream crossings will occur within 300 feet upstream or 100 feet downstream of an existing redd or spawning fish; After construction, temporary stream crossings will be removed and banks restored while adhering to turbidity requirements; • Cofferdams and diversion structures will have one foot of freeboard; • Dewatering pump discharge will be released onto floodplain areas away from wetlands and construction activities where discharge will fully infiltrate prior to reaching wetlands and surface waters unless otherwise approved; • Any return flows from dewatering discharge will meet turbidity requirements; • Bag fill materials will be clean, washed, and rounded material meeting standard specifications for drain rock, streambed aggregate, streambed sediments, or streambed cobbles; and, • Work activities within the ordinary high-water channel will conform with the water quality standards established for the project. Upon completion of the instream work, flow diversions will be removed slowly to allow gradual rewatering of the isolated stream segment to minimize turbidity. Once the stream segment is rewatered, the upstream and downstream block nets will be removed. Erosion and sediment control for inwater work will be consistent with controls used for other aspects of the project. Turbidity monitoring and protocols will include:





### **Table A-7. Summary of Environmental Design Features and Protection Measures: General – Water Resources.**



















#### **Table A-8. Summary of Environmental Design Features and Protection Measures: General – Wastes and Hazardous Materials.**






# **Table A- 9. Summary of Environmental Design Features and Protection Measures: General – Other.**





### **APPENDIX B**

### **GOLDEN MEADOWS SUMMARY OF ENVIRONMENTAL DESIGN FEATURES AND PROTECTION MEASURES**

A variety of standard operating procedures (SOPs) and project design features (PDFs) were incorporated into the Golden Meadows proposed action to minimize and avoid the risk of adverse impacts to ESA-listed fish and designated critical habitat. These SOPs and PDFs are fully described in the Golden Meadows BA Amendment and its supporting documentation. The most notable SOPs and PDFs for protection of ESA-listed species and designated critical habitat are listed below.

- All petroleum products will be transported in accordance with state and Federal Department of Transportation regulations, and handled and stored as per applicable state and Federal petroleum product storage and handling laws and regulations.
- Fuel hauling will only occur during daylight hours and when weather conditions are acceptable. Acceptable weather conditions will be determined jointly by the MGI [Midas Gold Inc.], the PNF [Payette National Forest], and the Valley County Road Department on a case-by-case basis.
- Setup and confirmation of at least two caches for spill response equipment will occur on the fuel delivery route.
- The pilot and emergency response vehicles will carry appropriate containment and spill response equipment. All drivers will be required to have spill response, safety, and resource awareness training.
- A spill prevention control and countermeasure (SPCC) plan will be implemented.
- Staff handling fuel or petroleum products will be trained to successfully implement the SPCC plan. Inspections of fuel storage and handling areas will be conducted as specified in the SPCC plan.
- A spill prevention and cleanup kit consisting of absorbent pads, absorbent booms, shovels, and a fire extinguisher will be placed at the fuel storage site (private property), at the core shack (private property), and at drill sites or any other areas where fuel and/or petroleum products are present.
- Water intake pumps will be placed in containment capable of holding 120% of the pump engine's fuel, engine oil, and hydraulic fluid. The smallest practical pump and intake hose will be used.
- Any fuel, oil, or chemical discharges; or spills greater than 25 gallons on land, or any spill directly in a stream will be reported to NMFS immediately (or as soon as possible after onsite containment efforts are implemented as per the SPCC plan) and emergency

consultation will be initiated. Spill response will be in accordance with the SPCC plan, which includes a trained onsite emergency response team.

- For drill areas with a higher risk of drilling fluids emerging at the ground surface, the following SOPs will be implemented:
	- o Drillers will exercise a high degree of vigilance for signs of lost circulation at shallow depths.
	- o The casing will be advanced simultaneously with the drill string through the alluvial section of all drill holes.
	- o Regular monitoring of the adjacent slopes below the drill rig for daylighting of drilling fluids will be performed by environmental technicians. Stream channels in these areas will also be regularly monitored. At least one person will be stationed on the slope below the drill rig at all times until surface casing is set.
	- o For drill holes proposed to be sited within an RCA [riparian conservation area] in areas identified as having risk for daylighting of drilling fluid, an interdisciplinary team including the PNF resource specialists and MGI geologists, drillers, and environmental technicians will conduct an onsite review of all geologic target considerations and environmental risk and mitigation factors in order to identify the optimal hole locations that would present the least environmental risk.
	- o Silt fence, straw wattles, portable sumps, pumps, and hoses will be pre-staged for emergency use to contain drilling fluid should it daylight. A PNF representative will verify that such measures are in place on the ground at locations warranting such precautionary measures.
- Sediment and erosion control BMPs [best management practices] will be implemented to minimize the potential for sediment to reach steams. For example, to minimize sediment runoff from temporary roads and roadbeds, water bars, silt fencing, certified weed-free waddles, and/or weed-free straw bales will be installed in strategic downslope areas and in RCAs. Proper BMPs will be used to prevent sediment from escaping drill pad and sump locations.
- Road rutting from traffic will be minimized by requiring construction and maintenance of surface drainage structures (e.g., water bars), application of surfacing material, and by restricting road use when conditions are unacceptable due to moisture that is leading to the onset of rutting and concentrations of turbid flow.
- Road maintenance will be conducted along the Stibnite Road under a cooperative agreement with Valley County (this segment of road is operated under a Forest Roads and Trail Act easement between Valley County and the PNF). Road maintenance and improvement actions that will be performed include, but are not limited to, cleaning roadside ditches and improving drainage, reducing potential rock fall, improving and

regrading roadway surfaces, replacing soft roadway materials, and adding surface coat aggregate with appropriate gradation and durability characteristics followed by application of dust abatement and binding products in select areas. Maintenance activities will be performed in accordance with the PNF Road Maintenance Programmatic (NMFS Tracking #2008/04131), unless otherwise noted in this letter (i.e., application of dust abatement chemicals).

- Additional road maintenance on the Stibnite Road will focus on those areas where modeled sediment delivery is greater than 0.1 tons of sediment per year (using GRAIP). These areas will be inspected and prioritized for surface aggregate. Placement of surface aggregate will occur during the first summer of project implementation.
- A gate will be installed within 300 feet of the bridge just east of the Profile Gap Road (Forest Service [FS] Road 340) and Stibnite Road (FS Road 412) intersection. The gate will be closed during the snow plowing and spring-break up seasons in order to regulate full-sized vehicle access to the project area. This will help avoid damage caused by motorized vehicles that could lead to excessive erosion and deterioration of the overall road condition. Administrative access beyond the gate by landowners, law enforcement, or government personnel may be permitted by Valley County.
- No chemical deicers or sand/gravel less than 3/8-inch will be used on roads.
- Dust abatement chemicals will be applied in a manner that avoids runoff into the streams. Where the road surface is within 25 feet (slope distance) of surface water, dust abatement chemicals will only be applied to a 10-foot swath down the centerline of the road. The rate and quantity of application will be regulated to ensure all of the chemical is absorbed and does not leave the road surface.
- Unless a request is made to reauthorize road use, all temporary roads will be decommissioned immediately after use to a condition equivalent to, or better than, their condition prior to use.
- Drill sites and other reclaimed drill areas will be monitored during spring runoff to ensure that sediment and erosion control BMPs are in place and working so that soil erosion is minimized.
- Travel and drilling off designated routes on NFS lands will only occur when there is adequate snow depth or frozen soil to prevent rutting and puddling. Snow used to build snow bridges across non-fish bearing streams will not contain soil or other debris.
- When a drill pad is needed in an RCA, MGI will submit a written request to the PNF for approval of the location. The request must include an explanation as to why there is no reasonable alternative to siting the pad in an RCA. MGI must receive approval from the PNF prior to pad construction. Drill pads in RCAs will be sited to avoid removing any large trees to the extent possible. Any trees that are felled within the RCA will be left in the RCA.
- Drill pads will not be located within 100 feet of streams.
- Employees and staff will receive training and direction to avoid harassment of spawning adult Chinook salmon and steelhead.
- Visual turbidity monitoring will occur immediately upstream and downstream of active drilling operations for drills working within 300 feet of surface water. If monitoring of the proposed action identifies unanticipated effects to fish or fish habitat, activities will cease until corrections can be made. The Level 1 Team will be informed or consultation will be reinitiated.

#### **APPENDIX C**

### **WEED TREATMENT DESIGN FEATURES AND PROTECTION MEASURES (PNF 2020; NMFS 2020)**

### **Table C-1. Herbicides proposed for use, application rates, and buffers from water.**



OHWM = ordinary high-water mark.

<sup>1</sup> Aquatic formulations of 2,4-D amine, glyphosate, imazamox, imazapyr, and triclopyr TEA shall be used.

<sup>2</sup> Other product brands of identical or "substantially similar" formulation may be added or substituted in the future (as described in the Payette National Forest Programmatic Activities consultation (NMFS 2020; PNF 2020).

**Table C-2. Recommended adjuvant type by herbicide proposed for use on the Payette National Forest.**

Herbicide	<b>Recommended Adjuvant Types</b>
$2,4-D$	NIS, Nitrogen sources, MSO
Aminopyralid	<b>NIS</b>
Chlorsulfuron	NIS, MSO, organosilicone
Clopyralid	NIS. MSO
Dicamba	Any as allowed by label
Fluroxypyr	No specific adjuvants are recommended
Glyphosate	<b>NIS</b>
Imazamox	NIS, MSO, organosilicone
Imazapic	NIS, MSO, organosilicone
Imazapyr	NIS, MSO
Imazamox	NIS, Nitrogen sources, MSO, petroleum/crop oil concentrate
Metsulfuron methyl	NIS, MSO, organosilicone
Picloram	None needed; can add as per surfactant manufacturer's label
Sulfometuron methyl	Any allowed by label
Triclopyr triethylamine salt (TEA)	<b>NIS</b>

 $NIS = non-ionic surfactant$ ;  $MSO = methylated or ethylated seed oils$ 

**Table C-3. Required mitigations and best management practices for invasive weed management activities that will be implemented to avoid or minimize potential effects on the Payette National Forest.**

<b>Treatment</b> <b>Method</b>	Category	<b>Mitigations and Best Management Practices</b>
Chemical	General	Contracts and agreements will include all of these design criteria as a minimum. $\bullet$ Annual reports and plans will be presented to the Level 1 Team. $\bullet$ New end use products that meet the Adaptive Management Strategy criteria will $\bullet$ be presented to the Level 1 Team for herbicide use approval.
	Application	Herbicide application shall comply with applicable laws, policy, guidelines, and $\bullet$ product label directions. A state certified applicator will oversee all herbicide applications. $\bullet$ Applicators will read and follow label directions, including instructions for $\bullet$ herbicide use, application rates, equipment and techniques, personal protective equipment for applicators and mixers, and container disposal. Program managers will ensure proper permitting is in place prior to $\bullet$ implementation. Material safety data sheets, safety plans, spill prevention plans, and clean-up kits ٠ will be available to applicators and mixers. Wind speed and direction and equipment and spray parameters will be monitored throughout herbicide application. Herbicides will not be applied when sustained wind conditions exceeding 5 $\bullet$ miles per hour in riparian areas or 8 miles per hour in upland areas. Herbicides will not be applied if the weather forecast predicts precipitation $\bullet$ within the next 24 hours. Accurate and detailed application records will be kept. ٠ Practical measures to restrict access to herbicides, adjuvants, and spray $\bullet$ equipment by unauthorized personnel will be implemented.
		The 5-Year Herbicide Safety Plan will be updated with completion of this $\bullet$ consultation and implemented during herbicide use. Indicator dye will be used in the herbicide mix to visually ensure uniform $\bullet$ coverage and minimize overlapped or skipped areas and treatment of non- target areas. Low pressure and larger droplet sizes will be used to the extent possible, to $\bullet$ minimize herbicide drift during broadcast operations. Appropriate nozzles designed for herbicide application will be used. ٠





#### **REFERENCES**

- NMFS (National Marine Fisheries Service). 2020. Endangered Species Act Section 7(a)(2) Biological Opinion and Magnuson–Stevens Fishery Conservation and Management Act Essential Fish Habitat Consultation for the Payette National Forest Programmatic Activities (Seven Forest-wide Activities). December 15, 2020. 269 pp.
- PNF (Payette National Forest). 2020. Programmatic Biological assessment for the potential effects from road maintenance; trails, recreation, and administrative site operation and maintenance; fire management; invasive weed management, timber harvest and precommercial thinning; miscellaneous forest products; and fish habitat and riparian sampling to Snake River fall and spring/summer Chinook salmon, Snake River sockeye salmon, Snake River steelhead, Columbia River bull trout, Northern Idaho ground squirrel, Canada lynx, and wolverine on the Payette National Forest. May 11, 2020.

## **APPENDIX D**

### **LEMHI LITTLE SPRINGS HABITAT RESTORATION CONSTRUCTION DETAILS AND PRELIMINARY DESIGN DRAWINGS (RIO ASE 2023)**

- 1. Description of Work Elements
	- a. Channel excavation to create new perennial and non-perennial, multi-threaded, channel networks to include pool and riffle features.
	- b. Offsite haul of material generated from project excavations to an approved contractor provided spoil location.
	- c. Installation of numerous types of wood habitat structures (i.e., beaver dam analogue, apex jam, bleeder jam, individual tress and logs, and bank jams) and bank and floodplain roughening treatments (using natural materials such as woody debris, coir fabric, rock and soil, and vegetation).
	- d. Installation of constructed riffles and strategic filling of existing channels for conversion to floodplain habitats.
	- e. Installation of temporary construction access routes, staging areas, and stream crossings and/or bridges.
	- f. Revegetation through planting and seeding of native species within riparian, wetland, and upland zones.
- 2. General Construction Sequence
	- a. Mobilize to the site, install erosion and sediment control measures in accordance with the stormwater pollution prevention plan (SWPPP), install temporary construction entrances/exits, perform clearing for temporary access routes, perform staking and/or localization of survey and/or machine control.
	- b. Work area isolation (i.e., cofferdams, pumping, and water management) and fish salvage.
	- c. New channel construction (excavation, hauling, wood habitat structures, constructed riffles, and bank treatments). Channel construction outside of the ordinary high water mark (OHWM) may be performed outside of the in-water work window.
	- d. Filling or blocking existing channel segments.
	- e. Prewash and activate new channels while isolated from the existing Lemhi River or active water. Perform fish salvage within existing channels to be filled.
	- f. Finish grading of all floodplain areas.
	- g. Reclamation of temporary construction access routes and staging areas to pre-existing conditions.
	- h. Planting, seeding, final inspection, site cleanup, and demobilization.
- 3. Work Schedule
	- a. The approved in-water work window for this project is July 1 to August 23. All work requiring equipment to operate partly or wholly below the OHWM shall be completed

during the in-water work window. Work that is outside of OHWM may be completed prior to and/or after the in-water work window if approved by the contracting officer.

- b. The contractor may not leave the work site or suspend activity for more than five (5) consecutive days after mobilizing to the site and prior to reaching substantial completion unless otherwise approved by the contracting officer.
- 4. Contractors Use of Premises
	- a. Prior to performing work, contractor shall become thoroughly familiar with the project site, site conditions, and all portions of work.
	- b. The contractor shall only use designated temporary access routes and stream crossings.
	- c. The contractor shall take all measures necessary to minimize damage to existing vegetation during construction activities.
	- d. The contractor shall only remove trees and shrubs that are absolutely necessary for the execution of the work and shall make all efforts to minimize tree and shrub removal. Contractor shall obtain prior approval from contracting officer to remove any tree or shrub from outside disturbance limits. Any tree or shrub unnecessarily removed from the work site shall be replaced by a new tree or shrub of equal or greater value at the sole expense of the contractor as approved by the contracting officer.
- 5. Equipment and Refueling

 $\overline{a}$ 

- a. Contractor is required to pressure wash and remove all dirt, grease, oil, fuel, vegetation, and weed seeds before bringing equipment onto the site.
- b. Complete vehicle and equipment staging, cleaning, maintenance, refueling, and fuel storage 150 feet away from any natural waterbody.
- c. Inspect all vehicles and equipment operated within 150 feet of any natural waterbody daily for fluid leaks before leaving staging areas. Repair any leaks detected in designated temporary construction staging areas before resuming operation. Document inspections in a record to be made available for review on request by the contracting officer and regulatory agencies.
- d. Use of equipment in flowing water is limited by applicable permits. Equipment must be thoroughly cleaned before entering the water.
- e. Hydraulics fluids all equipment performing work in active stream channels, or permanent water bodies during project construction must use hydraulic oil that meets or exceeds environmentally acceptable lubricants by the U.S. EPA (2011) (e.g., mineral oil, polyglycol, vegetable oil, synthetic ester; Mobil® biodegradable hydraulic oils, Total® hydraulic fluid, Terresolve Technologies Ltd.® biobased biodegradable lubricants, Cougar Lubrication® 2xt bio engine oil, series 4300 synthetic bio-degradable hydraulic oil, 8060-2 synthetic bio-degradable grease no. 2, etc. Or meet stringent acute aquatic toxicity, which is inherently biodegradable<sup>17</sup>. All products shall be American Petroleum Institute (API) certified and the vendor shall

<sup>&</sup>lt;sup>17</sup> This does not include trucks, dozers, front end loaders, etc., that are operated on the floodplain or involved in the construction of new channels prior to adding water flow or filling abandoned channels after de-watering.

furnish documentation of the certification upon request. Products must meet manufacturer's performance and warranty requirements.

- 6. Special Procedures
	- a. Instream Work
		- i. Proposed earthwork activities will occur inside and outside of the OHWM. The contractor is required to perform the work in a manner that does not cause turbidity exceedances to include proper turbidity controls such as installation of cofferdams, pumping, or other facilities approved by the contracting officer.
		- ii. Streambank vegetation shall be preserved and protected to the extent practical. No tree or shrub shall be removed unless approved by the contracting officer. The contractor shall not disturb the roots of woody vegetation in this area during project excavations to the extent practical.
	- b. Turbidity Monitoring and Protocols
		- i. Turbidity monitoring is required and shall be completed by the sponsor in accordance with the established protocols.
		- ii. The contractor is required to perform the work in a manner that does not cause turbidity exceedances. If turbidity exceedances occur, the contractor shall stop work at the direction of the contracting officer until further notice. Any delays due to turbidity exceedances caused by the contractor will be at the contractor's sole expense.
	- c. Temporary Utilities
		- i. All generators shall be placed outside of the OHWM with appropriate spill prevention and containment measures.
- 7. Temporary Environmental Controls
	- a. The contractor shall:
		- i. Prepare, file, implement, and maintain a SWPPP for the project.
		- ii. Prepare a spill prevention, control, and countermeasure (SPCC) plan for this project to include requirements to prevent spills throughout construction.
		- iii. Provide all labor, equipment, and materials to control dust on all access roads and disturbed areas several times per day to prevent dust nuisance or damage to persons, property, or activities, included but not limited to crops, cultivated fields, wildlife habitats, residences, agricultural activities, recreational activities, traffic, and similar conditions.
		- iv. Provide specialty mufflers for continuously running generators, pumps, and other stationary equipment.
- v. Perform construction activities by methods that will prevent entrance, or accidental spillage, of solid matter, contaminants, debris, or other pollutants or wastes into streams, flowing or dry watercourses, lakes, wetlands, reservoirs, or underground water sources. Such pollutants and wastes include, but are not restricted to refuse, garbage, cement, sanitary waste, industrial waste, hazardous materials, radioactive substances, oil and other petroleum products, aggregate processing tailings, mineral salts, and thermal pollution.
- vi. Ensure absorbent pads to soak up leaks and a fuel spill response kit (including rag pads and booms) of appropriate size for the equipment are available on site at all times and readily available throughout the construction period.
- 8. General Conservation Measures Applicable to All Actions
	- a. Timing of In-water Work
	- b. Idaho Department of Fish and Game (IDFG) guidelines for timing of in-water work windows will be followed.
		- i. Changes to established work windows will be approved by regional state biologists.
		- ii. The in-water work window will be provided in the construction plans.
	- c. Site Layout and Flagging
		- i. Construction areas will be clearly flagged prior to construction.
		- ii. Areas to be flagged will include:
		- iii. Sensitive resource areas, such as areas below OHWM, spawning areas, springs, and wetlands;
		- iv. Equipment entry and exit points;
		- v. Road and stream crossing alignments;
		- vi. Staging, storage, and stockpile areas; and,
		- vii. No-spray areas and buffers.
	- d. Temporary Access Roads and Paths
		- i. Existing access roads and paths will be preferentially used whenever reasonable, and the number and length of temporary access roads and paths through riparian areas and floodplains will be minimized.
		- ii. Temporary access roads and paths will not be built on slopes where grade, soil, or other features suggest a likelihood of excessive erosion or failure. If slopes are steeper than 30%, then the road will be designed by a civil engineer with experience in steep road design.
		- iii. The removal of riparian vegetation during construction of temporary access roads will be minimized. When temporary vegetation removal is required, vegetation will be cut at ground level (not grubbed).
- iv. At project completion, all temporary access roads and paths will be obliterated, and the soil will be stabilized and revegetated. Road and path obliteration refer to the most comprehensive degree of decommissioning and involves decompacting the surface and ditch, pulling the fill material onto the running surface, and reshaping to match the original contour.
- e. Temporary Stream Crossings
	- i. Existing stream crossings will be preferentially used whenever reasonable, and the number of temporary stream crossings will be minimized.
	- ii. Temporary bridges and culverts will be installed to allow for equipment and vehicle crossing over perennial streams during construction. Treated wood shall not be used on temporary bridge crossings or in locations in contact with or directly over water.
	- iii. For projects that require equipment and vehicles to cross in the wet:
		- 1. The location and number of all wet crossings shall be documented in the construction plans;
		- 2. Vehicles and machinery shall cross streams at right angles to the main channel whenever possible;
		- 3. No stream crossings will occur 300 feet upstream or 100 feet downstream of an existing redd or spawning fish; and,
		- 4. After project completion, temporary stream crossings will be obliterated and banks restored.
- f. Staging, Storage, and Stockpile Areas
	- i. Staging areas (used for construction equipment storage, vehicle storage, fueling, servicing, and hazardous material storage) will be 150 feet or more from any natural waterbody or wetland. Staging areas closer than 150 feet will be pre-approved.
	- ii. Natural materials used for implementation of aquatic restoration, such as large wood, gravel, and boulders, may be staged within 150 feet if clearly indicated in the plans that area is for natural materials only.
	- iii. Any large wood, topsoil, and native channel material displaced by construction will be stockpiled for use during site restoration at a specifically identified and flagged area.
	- iv. Any material not used in restoration, and not native to the floodplain, will be disposed of outside the 100-year floodplain.
- g. Equipment
	- i. Mechanized equipment and vehicles will be selected, operated, and maintained in a manner that minimizes adverse effects on the environment (e.g., minimally-sized, low pressure tires; minimal hard-turn paths for tracked vehicles; temporary mats or plates within wet areas or on sensitive soils).
- ii. Equipment will be stored, fueled, and maintained in a clearly identified staging area that meets staging area conservation measures;
- iii. Equipment will be refueled in a vehicle staging area or in an isolated hard zone, such as a paved parking lot or adjacent, established road (this measure applies only to gas-powered equipment with tanks larger than 5 gallons);
- iv. Biodegradable lubricants and fluids will be used on equipment operating in and adjacent to the stream channel and live water.
- v. Equipment will be inspected daily for fluid leaks before leaving the vehicle staging area for operation within 150 feet of any natural water body or wetland; and,
- vi. Equipment will be thoroughly cleaned before operation below OHWM, and as often as necessary during operation, to remain grease free.
- h. Temporary Erosion Control
	- i. Temporary erosion controls will be in place before any significant alteration of the action site and appropriately installed downslope of project activity within the riparian buffer area until site rehabilitation is complete;
	- ii. If there is a potential for eroded sediment to enter the stream, sediment barriers will be installed and maintained for the duration of project implementation;
	- iii. Temporary erosion control measures may include sedge mats, fiber wattles, silt fences, jute matting, wood fiber mulch and soil binder, or geotextiles and geosynthetic fabric;
	- iv. Soil stabilization utilizing wood fiber mulch and tackifier (hydro-applied) may be used to reduce erosion of bare soil if the materials are noxious weed free and nontoxic to aquatic and terrestrial animals, soil microorganisms, and vegetation;
	- v. Sediment will be removed from erosion controls once it has reached one-third of the exposed height of the control; and,
	- vi. Once the site is stabilized after construction, temporary erosion control measures will be removed.
- i. Emergency Erosion Controls
	- i. The following materials for emergency erosion control will be available at the work site:
		- 1. A supply of sediment control materials; and,
		- 2. An oil-absorbing floating boom whenever surface water is present.
- j. Dust Abatement
	- i. The project sponsor will determine the appropriate dust control measures by considering soil type, equipment usage, prevailing wind direction, and the effects caused by other erosion and sediment control measures.
- ii. Work will be sequenced and scheduled to reduce exposed bare soil subject to wind erosion.
- iii. Dust-abatement additives and stabilization chemicals (typically magnesium chloride, calcium chloride salts, or lignin sulfonate will not be applied within 25 feet of water or a stream channel and will be applied so as to minimize the likelihood that they will enter streams. Applications of lignin sulfonate will be limited to a maximum rate of 0.5 gallons per square yard of road surface, assuming mixed 50:50 with water.
- iv. Application of dust abatement chemicals will be avoided during or just before wet weather, and at stream crossings or other areas that could result in unfiltered delivery of the dust abatement materials to a waterbody (typically these would be areas within 25 feet of a waterbody or stream channel; distances may be greater where vegetation is sparse or slopes are steep).
- v. Spill containment equipment will be available during application of dust abatement chemicals.
- vi. Petroleum-based products will not be used for dust abatement.
- k. Spill Prevention, Control, and Counter Measures
	- i. A description of hazardous materials that will be used, including inventory, storage, and handling procedures will be available on-site.
	- ii. Written procedures for notifying environmental response agencies will be posted at the work site.
	- iii. Spill containment kits (including instructions for cleanup and disposal) adequate for the types and quantity of hazardous materials used at the site will be available at the work site.
	- iv. Workers will be trained in spill containment procedures and will be informed of the location of spill containment kits.
	- v. Any waste liquids generated at the staging areas will be temporarily stored under an impervious cover, such as a tarpaulin, until they can be properly transported to and disposed of at a facility that is approved for receipt of hazardous materials.
	- vi. Pumps used adjacent to water shall use spill containment systems.
- l. Invasive Species Control
	- i. Prior to entering the site, all vehicles and equipment will be power washed, allowed to fully dry, and inspected to make sure no plants, soil, or other organic material adheres to the surface.
	- ii. Watercraft, waders, boots, and any other gear to be used in or near water will be inspected for aquatic invasive species.
	- iii. Wading boots with felt soles are not to be used due to their propensity for aiding in the transfer of invasive species unless decontamination procedures have been pre-approved.



D-8










































# **REFERENCES**

Rio ASE. 2023. Lemhi Little Springs Habitat Restoration Project, Draft Basis of Design Report (Preliminary Design). Prepared for Perpetua Resources Idaho, Inc. 28pp. June.

#### **APPENDIX E**

#### **SITE-WIDE WATER CHEMISTRY MODEL SUMMARY**

In order to predict instream concentrations, modeling was first performed to estimate contaminant contributions from the mine facilities (i.e., tailings storage facility (TSF) buttress and embankment, TSF, backfilled pits, and the West End pit lake). These source estimates were then coupled with the hydrologic model, and the wastewater treatment plant effluent in a sitewide water chemistry (SWWC) model to estimate instream concentrations. For context, a brief summary of the modeling is provided in this appendix. Additional detail can be found in the SWWC modeling report (SRK 2021a) and the SWWC sensitivity analysis report (SRK 2021b).

To minimize the volumes of contact water requiring treatment, upstream non-contact water will be diverted to prevent it from interacting with mine facilities during operations. Table E-1 identifies the non-contact diversion channels that are considered in the SWWC model. At closure, the diversion channels will be decommissioned, and non-contact water will follow its natural drainage pathways.

<b>Diversion Channel</b>	<b>Description</b>
North Diversion	Diverts non-contact runoff from the north of the tailings storage
	facility (TSF) and TSF buttress to Meadow Creek
South Diversion	Diverts Meadow Creek and its tributaries from the south and
	west of the TSF around the TSF
<b>Hennessy Diversion</b>	Diverts water from Hennessy Creek away from the Yellow Pine
	Pit (YPP) to Fiddle Creek
<b>Midnight Diversion</b>	Diverts Midnight Creek away from the YPP to the EFSFSR
<b>West End Diversion</b>	Diverts upper West End Creek around the West End pit
East Fork South Fork Salmon	Diverts the EFSFSR around the YPP downstream of assessment
River (EFSFSR) Tunnel	node YP-SR-6 to upstream of assessment node YP-SR-4

**Table E-1. Summary of diversion channels included in the site-wide surface water chemistry model.**

# **Mine Facility Source Terms**

Geochemical testing of mine material (e.g., development rock, ore, legacy facilities, and tailings) was performed (SRK 2021a). Based on humidity cell tested (SRK 2021a), the rock in the pit walls and the development rock deposited in the TSF buttress and pit backfills is expected to be largely non-acid generating. However, it will be capable of leaching aluminum, antimony, arsenic, cadmium, copper, manganese, mercury, zinc, sulfate, and total dissolved solids (TDS) into surface water and groundwater. Geochemical source terms were developed for various material groupings (e.g., development rock and ore by pit location, gold grade, and potentially acid generating [PAG] classification) and subsequently used to predict the chemical contributions of the mine facilities to the environment. In order to perform the modeling, a number of considerations and assumptions were made when assessing contaminant contribution over the life of the proposed action from these various sources. A detailed accounting of the considerations and assumptions made are included in the SWWC modeling report (SRK 2021a), with a select few summarized below.

## **TSF Buttress and Embankment/Ore Stockpiles.**

None of the development rock storage facilities or ore stockpiles will be lined at their base. As a result, a portion of the rainfall and snowmelt coming into contact with these facilities infiltrate into the groundwater, creating a nonpoint source of pollution to surface water. Long-term infiltration and associated groundwater contamination are expected to be reduced during closure and reclamation as a result of capping the TSF embankment, buttress, and any remaining ore stockpiles. The upstream face of the TSF embankment will be fully lined in order to minimize leakage.

Conceptual models illustrating water flow paths associated with the TSF Buttress and Embankment during operations and closure are shown in Figure E-1. The predicted runoff water quality, toe/pop-out seepage chemistry, and groundwater chemistry under the TSF buttress and embankment during operations and post-mining (before and after cover placement) are summarized in Tables 3.5-7, 3.5-8, and 3.5-9 in the biological assessment (BA) (Stantec 2024), respectively. Select considerations and assumptions employed for predicting water quality are listed below.

- Any groundwater recharge from the buttress and embankment will interact with alluvial groundwater within the upper 10 meters of the water table underneath the footprint of the facility prior to reporting to surface water (i.e., groundwater flow to Meadow Creek) or the Hangar Flats pit. The specific yield within the alluvium is 15 percent.
- Solutes will not attenuate along the flowpath.
- Only a small fraction of (4 percent) of the total mass of material within the TSF embankment and buttress will be effectively contacted by meteoric waters. This based on the following assumptions:
	- o Only 20 percent of the material will consist of fines and be available for chemical weathering reactions.
	- o Infiltration will flow along preferential flow paths and contact only 20 percent of rock volume.
- No ore stockpiles will remain on the TSF buttress.
- Following cover placement on the TSF buttress and embankment, residual solution from the buttress materials will continue to infiltrate into the groundwater, though to a lesser degree as modeled during operations. Any toe/pop-out seepage will occur under the liner and will recharge groundwater.
	- o Infiltration from rain and snowmelt will be reduced to 5 percent of the annual average precipitation.
	- o The flow paths are assumed to be reduced by approximately 86 percent to account for reduction in infiltration and resulting decrease in the proportion of material contacted by infiltrating water.
- Reaction rates and solute release in the field will vary from laboratory predictions due to differences in ambient temperatures. In the field, the annual average temperature is 2.6°C whereas the laboratory studies were conducted at 25°C. A scaling factor (ratio of temperature in the field to temperature in the lab) was applied to account for this.
- The TSF embankment and buttress were assumed to be fully oxygenated and no additional scaling factor was applied.
- It is assumed that chemistry in the TSF embankment and buttress will be proportional to the type of material housed in the facilities.
- Mercury source terms were set to 0 since Phase 2 leachate tests were generally below a detection limit of 6 ng/L.



**Figure E-1. Conceptual model for tailings storage facility buttress and embankment during operations and post closure (SRK 2021a).**

# **Tailing Storage Favility**

The TSF will be fully lined in order to minimize leakage. An underdrain groundwater collection and conveyance system will be installed beneath the TSF liner, TSF embankment, and TSF buttress. This conveyance system will facilitate detection of liner leakages as well as minimize the potential for groundwater to saturate the base of the embankment and buttress. During closure and reclamation, after the tailings have consolidated, the TSF will be capped in order to reduce long-term infiltration.

Figure E-2 provides conceptual models of water flow paths associated with the TSF during operations and post closure after placement of a cover. The predicted water quality of the TSF surface water and groundwater is summarized in Tables 6-7 and 6-8 of the SWWC modeling report (SRK 2021a), respectively. Select considerations and assumptions employed for predicting water quality include:

- Minor seepage from manufacturing defects and other larger holes in the liner or the seams developed during placement may occur, despite the best practice design. The model assumes one defect with an area of 99.9 mm<sup>2</sup> per acre. Seepage from these defects is assumed to interact with the uppermost 10 meters of the groundwater table below the TSF.
- Liner leakage will decrease from mine year 14 through mine year 41, when the tailings consolidation is expected to be complete (USFS 2023).
- Pore water within the TSF is primarily comprised of process water chemistry obtained from the metallurgical testwork program and weighted according to total tailings proportions over the life of the mine (SRK 2021a).
- Tailings reclamation will be completed within 9 years after ore processing operations cease. It is assumed that cover placement will begin 3 to 5 years after the end of tailings deposition and be completed by mine year 23.
- The low permeability geosynthetic cover will reduce infiltration to the TSF by 95% of the uncovered infiltration volume. Minor infiltration through the cover may contact the upper portion of the underlying tailings. It is assumed that this contact water will mix with clean runoff/run on water and consolidation water.
- Consolidation water from beneath the cover will be withdrawn using a combination of wells, wicks, and/or gravel drains and routed to water treatment.
- The quality of consolidation water is represented by decant solution chemistry from the metallurgical testwork program (Table 6-5 in SRK 2021a). This is conservative, as it does not account for dilution by precipitation that may infiltrate the TSF.
- Based on model results, treatment will no longer be required after about 25 years from the end of ore processing (mine year 40).



**Figure E-2.TSF Conceptual model during operations and post closure after cover placement. Although not shown, the conceptual model for mine years 15 through 22 is similar to that for Mine Years -2 to 12, with the only difference being the removal of the "thickened slurry to TSF" element (SRK 2021a).** 

### **Hangar Flats Pit, YPP, and Midnight Pit.**

During operations, groundwater (include water expressed along the pit walls) and any precipitation/snow melt will be pumped from the pits and used for ore processing. Upon closure, development rock placed in pit backfills will be inundated and contaminants are expected to leach from the backfilled material to alluvial and bedrock groundwater. Both the backfilled Yellow Pine Pit (YPP) and Hangar Flats pit will be covered with a low permeability geosynthetic liner, limiting infiltration of rainfall and snowmelt into the development rock and reducing potential groundwater impacts. The Midnight pit will not be covered with a liner.

Figures E-3 through E-5 illustrate the conceptual models of water flow paths associated with the YPP, Hangar Flats Pit, and Midnight Area Pit, respectively.



**Figure E-3.Yellow Pine backfilled pit conceptual model (SRK 2021a).**



**Figure E-4.Hangar Flats backfilled pit conceptual model (SRK 2021a)**



**Figure E-5. Midnight Area backfilled pit conceptual model (SRK 2021a).**

Contaminant loading within the backfilled pits will come from the backfill itself along with talus on the pit benches, groundwater, pit wall fractures, and precipitation that contacts exposed pit walls and backfill. The predicted water quality of backfilled Hangar Flats Pit and YPP, and Midnight Area pit is summarized in Tables 7-14, 7-16 and 7-17 of the SWWC modeling report (SRK 2021a), respectively. Select considerations and assumptions employed for predicting facility geochemistry include:

- The Midnight Area pit is above groundwater level; therefore, the backfill material and highwall will not be saturated.
- Limited water will pond within pit sumps in the YPP and Hangar Flats pit during operations.
- Pit backfill will be comprised of homogeneously mixed material types.
- The water table will rebound, partially flooding the backfill material within the YPP and Hangar Flats Pit.
- The low permeability geosynthetic cover on the YPP and Hanger Flats pit will reduce infiltration by 95% of the uncovered infiltration volume.
- Estimated proportional contribution of lithological units above or below water levels for the Hangars Flats pit and the YPP.
- Groundwater equilibrium occurs at Hangar Flats Pit by end of mine year 7; approximately 48 percent of the pit wall area will be submerged.
- Groundwater equilibrium occurs at the YPP by end of mine year 22; approximately 55 percent of the pit wall area will be submerged.
- Blast-induced fracturing of pit walls will occur, and it is assumed that fracturing will propagate to a depth of 3 feet and the density of fracturing will average 10 percent. Flushing os solutes from pit wall fractures by groundwater will only occur in the 'active' zone of groundwater inflow and will cease once pit walls become submerged.
- Only a small percentage of the talus remaining on pit benches will be contacted by runoff water entering the pit.
- The proportion of fines in the backfill is assumed to be 20 percent.
- The proportion of flow paths in unsaturated backfill is assumed to be 20 percent. This is reduced to 4 percent upon cover placement.
- Temperature scaling factor is 0.06.
- The overall contacted mass of backfill material contacted by water is 4 percent above the water table and 20 percent below the water table.
- Leaching of solutes below the water table is assumed to not occur due to minimal oxygen and flow.
- Groundwater quality entering the backfilled pits is influenced by the regional groundwater plus additional contaminants from upstream sources (e.g., TSF, TSF buttress and embankment, etc.).
- Water infiltrating the ore stockpile placed on the Hangar Flats backfill from mine year 7 to 12 is assumed to infiltrate into the backfill. It is assumed the stockpile will be removed by mine year 13.

## **West End Pit Lake**

During operations, the West End pit is expected to be relatively dry. Water from precipitation or groundwater infiltration will be pumped to a contact water pond. When mining of the pit is complete (mine year 12), dewatering will cease, West End Creek will be diverted to the pit, and the lake will begin to fill slowly. It is expected to take 57 years to fill. Figure E-6 illustrates the conceptual model developed for the West End pit lake.



**Figure E-6. West End pit lake conceptual model (SRK 2021a).**

The predicted water quality of the West End Pit Lake is summarized in Table 8-3 of the SWWC modeling report (SRK 2021a). Select considerations and assumptions employed for predicting pit lake geochemistry include:

- Contaminant loading will come from groundwater and pit wall runoff. Additional contaminant loading will come from fractures in the pit wall and talus on the pit benches.
- Predicted chemical concentrations are based on mixing of groundwater, pit wall runoff, precipitation, and surface water in ratios defined by the hydrologic model.
- Predicted chemical composition is based on the estimated proportional contribution of lithological units above or below water levels.
- The lake is likely to experience seasonal stratification, turning over each spring and fall; however, the model predicts the lake chemistry under well-mixed, annual average conditions.
- Precipitates are assumed to sink to the bottom of the pit lake and be removed from future chemical interactions. These precipitated mineral phases are assumed to be unavailable for re-dissolution.
- Trace metals removed from solution via sorption onto mineral precipitates such as iron oxides are permanently removed from the system.
- Water is predicted to exit the pit lake via groundwater losses (65 percent) and evaporation (35 percent).

# **Site-Wide Water Chemistry Model**

A SWWC model was developed in order to predict instream water quality as a number of assessment nodes in the project area (refer to Figure 29 in the opinion). The predicted chemical concentrations are a combination of natural and anthropogenic factors. Natural factors include the undisturbed mineralized orebody chemical contributions throughout the hydrologic system. Anthropogenic factors include legacy mining and the SGP. The natural and legacy mining constituent concentrations have been characterized by site-surface water monitoring data to establish constituent concentrations for the existing site condition. Contributions from the SGP are characterized by the individual facility geochemical models for the TSF, TSF buttress and embankment, the backfilled Hangar Flats, Yellow Pine and Midnight Pits, West End Pit lake, and water treatment plant (WTP) effluent quality. The SWWC model accounts for treated effluent as follows:

- Upstream of YP-T-27 from mine year -2 through mine year 12
- Upstream of YP-SR-10 from mine year 13 to mine year 23
- Upstream of YP-T-22 from mine year 23-40

The water chemistry models were coupled with surface and groundwater flow predictions from the site-wide water balance and hydrogeological model (Brown and Caldwell 2021; Perpetua 2021). Water quality was predicted on a monthly timestep for the following time periods:

- Existing conditions (Mine Year -37 to -3).
- Open pit mining (Mine Year -2 to 12), the last two years of operations where ore stockpiles are being processed is not included.
- Post-mining during water treatment (mine year 13 to 40)
- Post-mining when no water treatment is occurring (mine year 41 through 112).

The SWWC modeling report (SRK 2021a) and the Water Quality Specialist Report (Forest Service 2023, Section 7) provides details and references regarding the potential sources of chemical contaminants and predicted concentrations in surface waters during the construction, operations, and closure/post-closure periods.

Notable considerations and assumptions used in the SWWC model are listed below.

- There will be no runoff from the TSF discharged to Meadow Creek during operations.
- Following mine year 40, the model assumes there will be no contribution of consolidation water from the TSF to surface water.
- No surface outflow from the West End pit lake will occur.
- WTP effluent will be directed to Meadow Creek when flow augmentation is required through the end of mine year 12.
- WTP effluent will be directed to the EFSFSR (below Meadow Creek) the remainder of operations until the end of August of Mine Year 23.
- WTP effluent will be discharged to Meadow Creek upstream of YP-T-22 until Mine Year 40.
- Estimated surface and groundwater flow volumes contributing to each instream assessment node were based on information in the site-wide water balance and catchment wide hydrogeologic models. Catchment runoff volumes consisted of disturbed ground runoff (contact water) and undisturbed runoff and were proportional to estimated disturbed areas (refer to Table 10-1 of the SWWC model report (SRK 2021a)).
- Runoff from disturbed and undisturbed ground was assigned water chemistries associated with observed concentrations in water chemistry samples from the area.
- The model assumes existing impacts from the SODA/Bradley tailings and Hecla Heap will be completely removed and no residual flow from these facilities will remain following reclamation.
- The model accounted for losing stream reaches (i.e., reaches where surface water is lost to groundwater) by reducing inflows to surface water nodes by an equal volume. This was done to account for mass loss to groundwater prior to mixing at the next downstream location.
- The Bradley dumps are not removed and recharge estimates were assumed to remain the same as existing conditions during operations and post-closure. Infiltration from the Bradley dumps was predicted to comprise 44 and 16 percent of the groundwater flow reporting to instream assessment nodes YP-SR-4 (EFSFSR upstream of Sugar Creek confluence) and YP-SR-2 (EFSFSR below Sugar Creek confluence).
- Groundwater contributions from backfilled pits influence more than one instream assessment node:
	- o Hangar Flats pit backfill: YP-T-22 (66 percent), YP-SR-10 (14 percent), and YP-SR-8 (6 percent);
	- o YPP backfill: YP-SR-4 (68 percent), YP-T-1 (10 percent), and YP-SR-2 (2 percent); and
	- o West End pit lake: YP-T-6 (25 percent) and YP-T-1 (36 percent).
- West End pit lake outflow is predicted to take 16 years to reach West End Creek and Sugar Creek. The SWWC model accounts for this delay.
- The site facility geochemical predictions were assumed to be constant for each month of the corresponding year.
- A load reduction source term was developed to account for mining of the YPP and diversion of the EFSFSR and the associated removal of an existing load contributing to the EFSFSR at assessment node YP-SR-4.

The predicted water quality at each assessment node is calculated by mass-balance mixing of the various contributing water sources. These sources of flow and geochemical information used for each assessment node for the existing conditions and operation and post-operations periods are summarized in Tables 10-5 and 10-6 of the SWWC model report, respectively (SRK 2021a). Monthly variability in the SWWC model outputs are driven by variation in monthly surface and groundwater source terms as well as monthly variability in the hydrological model (SRK 2021a, Brown and Caldwell 2021). The average, minimum, and maximum values for each assessment node were calculated based on the modeled monthly values for each of the time periods.

# **Modeling Sensitivity Analysis**

The degree of potential predictive error from the geochemical model assumptions and SGP design features was evaluated through sensitivity analysis simulations (SRK 2021b). The sensitivity analysis addressed the model uncertainties associated with the potential for acidgeneration and leaching reactions as well as the scaling assumptions related to the proportion of preferential flow paths and finer particle gradation in the TSF Buttress and pit backfills and pit wall fracture thickness and density. Findings from the sensitivity analysis include the following:

- Varying model input parameters for the sensitivity analysis had little effect on the model results during mine operations.
- Varying model input parameters (percentage of development rock fines, the percentage of rock contacted due to preferential flow paths through the TSF Embankment and Buttress, and increasing the reaction temperature) had a substantial effect on closure and post closure model results for some parameters. – effectively doubling or tripling the predicted instream concentrations.
	- o When the bulk scaling factor of reactive rock is increased, concentrations of arsenic, antimony, sulfate, mercury, and aluminum are predicted to increase in contact water derived from the mined materials. The constituents exceeding surface water standards in contact water were the same as those predicted for the 2021 Modified Mine Plan, but the duration of contact water exceedances was affected in the model sensitivity runs and extended to the end of the mining period.
- In one of the model sensitivity runs, the neutralization potential ratio (NPR) cutoff for defining PAG material was increased to 2 (resulting in a greater percentage of pit wall rock and development rock lithology types being classified as PAG). The post-closure model results were not sensitive to increasing the NPR cutoff. The lack of model sensitivity to this parameter occurs because the mass loading rates for some constituents are lower in the PAG model source term input compared to some non-PAG units. Thus, increasing the percentage of PAG rock in the TSF Buttress and pit lake models does not lead to higher predicted post-closure concentrations.
- The model is not sensitive to varying the pit wall blast-damaged zone thickness.

Overall, the sensitivity analysis indicates the SWWC model is most sensitive to variations in the bulk scaling factor (e.g., reactive mass and temperature). Changing the NPR cutoff for defining PAG material or pit wall fracture depths do not substantially alter predicted instream concentrations during mine operational or post-closure periods. Incorporation of first-flush chemistry in the model predictions will slightly increase predicted analyte concentrations. Finally, effects of model uncertainty from simulating dissolved rather than total concentrations have not been evaluated, but total concentrations of analytes that appear in particulate form will be greater than the simulated dissolved concentrations.

### **Model Results**

### **Meadow Creek (YP-T-27 and YP-T-22).**

Meadow Creek assessment node YP-T-27 is located downstream of the TSF and TSF buttress/embankment and represents water quality conditions influenced by surface and groundwater inflow from upstream, TSF, TSF buttress and embankment, recharge from the SODA/Bradley tailings, and treated effluent. Meadow Creek assessment node YP-T-22 is located downstream of the Hangar Flats pit and represents water quality conditions influenced by upstream sources, Hecla Heap, Hangar Flats pit backfill, and EF Meadow Creek drainage. The modeled existing and predicted future surface water concentrations at nodes YP-T-27 and YP-T-22 are summarized in Tables E-2 and E-3, respectively. Predicted concentrations of pH, antimony, arsenic, and mercury for YP-T-22 are shown in Figure E-7.

Predicted concentrations of most metals are expected to increase during the open pit mining period; however, concentrations are expected to remain below applicable water quality criteria that have been deemed to be sufficiently protective. Arsenic, copper and antimony concentrations are predicted to decrease during open pit mining, though peak concentrations of arsenic and antimony will remain above applicable criteria. Mercury concentrations are predicted to increase, with predicted peak mercury concentrations exceeding the evaluation threshold of 2 ug/L.

During early closure, a number of metals concentrations will remain elevated relative to existing conditions; however, concentrations will remain below applicable water quality criteria. Antimony and arsenic concentrations are predicted to decrease, but will remain at levels higher than applicable criteria. Mercury concentrations are predicted to be similar to existing conditions, although peak concentrations will be slightly higher during early closure. By late closure, most metals concentrations are lower relative to existing conditions. Cadmium, chromium, and nickel are predicted to increase slightly, but concentrations are well below the evaluation thresholds. The predicted improvement in water quality of Meadow Creek is related to the removal of historical unlined mine waste disposal areas from the Meadow Creek drainage and the construction of lined and covered facilities as part of the SGP.

The exception to the reduced analyte concentrations are mercury concentrations which exhibit some variability during the operational and early closure periods attributable to predicted variations in effluent chemistry from the WTP. Predicted long-term surface water mercury concentrations are comparable to the existing conditions at the location.

<b>Parameter</b>	<b>Units</b>	<b>Evaluation</b> Threshold <sup>1</sup>		<b>Existing Condition Mine</b> Year -37 to -3			Open Pit Mining Mine Year Post-Mining during Water $-2$ to 12			<b>Treatment</b>		<b>Post-Mining no Water</b> <b>Treatment</b> Mine Year 41 to 112			
			$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	
pH	mg/L	$6.5 - 9$	7.1	6.8	7.3	7.1	6.8	7.3	7.0	6.5	7.2	6.9	6.4	7.2	
$Ag^{3,4}$	$\ensuremath{\mathrm{ug}}\xspace/L$	0.7	0.01	0.0099	0.011	0.012	0.0099	0.061	0.017	0.0100	0.046	0.021	0.0100	0.047	
${\rm Al}^5$	$\rm ug/L$	385	11	6.7	26	12	7.4	27	12	6.0	27	11	5.9	24	
$\mathbf{A}\mathbf{s}$	$\ensuremath{\mathrm{ug}}/\ensuremath{\mathrm{L}}$	10	35.0	3.1	83	2.9	0.80	23	1.9	0.77	17	1.4	0.76	2.7	
Cd <sup>4</sup>	$\rm ug/L$	0.3	0.01	0.01	0.01	0.01	0.01	0.04	0.01	0.01	0.03	0.02	0.01	0.03	
$Cr^4$	$\text{ug/L}$	10.6	0.18	0.12	0.28	0.20	0.10	1.1	0.14	0.070	0.29	0.13	0.062	0.28	
Cu <sup>6</sup>	$\ensuremath{\mathrm{ug}}\xspace/L$	$\mathfrak{2}$	0.32	0.16	1.6	0.27	0.16	1.3	0.22	0.11	0.30	0.21	0.10	0.27	
$\rm Hg^7$	ng/L	2	1.0	0.50	$2.2\,$	1.4	0.53	5.1	1.0	0.43	2.3	0.92	0.42	1.9	
Ni <sup>4</sup>	ug/L	24	0.18	0.11	0.23	0.27	0.10	3.0	0.71	0.10	3.3	1.10	0.14	3.4	
Pb <sup>4</sup>	$\rm ug/L$	0.9	0.02	0.01	0.09	0.034	0.010	0.35	0.013	0.0067	0.060	0.011	0.0061	0.018	
Sb	$\mu g/L$	5.2	7.3	1.0	18	0.94	0.24	7.3	0.61	0.20	7.6	0.40	0.20	1.0	
Se	$\mu g/L$	3.1	0.50	0.50	0.50	0.50	0.38	0.65	0.46	0.30	0.50	0.44	0.29	0.50	
$SO_4$	$\text{Im}g/L$	250	6.7	2.4	13	3.8	1.5	15	2.7	1.4	4.5	2.6	1.4	3.8	
$Zn^4$	ug/L	54	0.79	0.46	1.9	1.1	0.46	7.1	0.65	0.31	2.0	0.60	0.28	1.8	
TDS <sup>8</sup>	mg/L	500	57	26	80	38	20	89	33	19	44	32	18	41	
$NO2 + NO3$	mg/L as N		0.50	0.30	0.58	0.75	0.26	14	0.43	0.22	0.58	0.40	0.20	0.54	

**Table E-2. Summary of site-wide water chemistry model predictions in Meadow Creek at node YP-T-27.**

Notes: Yellow cells represent predicted increases in concentrations relative to the modeled existing conditions. Green cells represent predicted decreases in concentrations relative to the modeled existing conditions. Bold italic numerical values represent concentrations greater than surface water criteria/thresholds identified in the third column.

<sup>1</sup>Unless otherwise noted, the values presented here are the most stringent surface water quality criteria, reported as two significant digits.

<sup>2</sup>Average, minimum and maximum values are calculated based on the monthly predicted concentrations over the indicated time period.

<sup>3</sup>The evaluation threshold is based on the acute aquatic life criterion.

<sup>4</sup>Concentrations are based on a hardness value of 40 mg/L as calcium carbonate, which represents the 5<sup>th</sup> percentile hardness during the driest four months at EFSFSR node YP-SR-10.

<sup>5</sup>The aluminum threshold is based on EPA's nationally recommended criteria guidelines. Values for total hardness, pH, and dissolved organic carbon collected on the same day at USGS gaging stations EFSFSR above Sugar Creek (13311250) and EFSFSR at Stibnite (13311000) from 2014-2022 were used to calculate the chronic criterion. The 5<sup>th</sup> percentile value was selected as the evaluation threshold.

<sup>6</sup>The evaluation threshold was derived using the biotic ligand model per guidance contained in IDEQ 2017, and was estimated by applying the lowest  $10<sup>th</sup>$  percentile chronic aquatic life criterion based on regional classifications for the Salmon River Basin, Idaho Batholith, and third order streams.

 $7$ The evaluation threshold is based on recommendations by NMFS (2014).

Parameter pH	<b>Units</b>	Evaluation Threshold <sup>1</sup>		<b>Existing Conditions</b> Mine Year -37 to -3		<b>Open Pit Mining</b> Mine Year -2 to 12				<b>Post-Mining during Water</b> <b>Treatment</b> Mine Year 13 to 40		<b>Post-Mining no Water</b> <b>Treatment</b> Mine Year 41 to 112			
			$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	
	mg/L	$6.5 - 9$	7.2	6.9	7.3	7.2	6.9	7.3	7.1	6.6	7.3	7.0	6.6	7.3	
$\rm Ag^{3,4}$	ug/L	0.7	0.010	0.0099	0.011	0.012	0.0099	0.048	0.018	0.010	0.040	0.019	0.01	0.042	
Al <sup>5</sup>	$\text{ug/L}$	385	11	6.0	26	12	6.9	27	12	6.2	27	11	6.1	25	
As	ug/L	10	32	4.0	75	3.8	1.4	18	3.0	1.2	13	2.5	1.2	3.7	
Cd <sup>4</sup>	ug/L	0.3	0.01	0.01	0.01	0.01	0.01	0.03	0.01	0.01	0.02	0.014	0.01	0.024	
$Cr^4$	ug/L	10.6	0.20	0.13	0.28	0.20	0.10	1.0	0.25	0.09	0.82	0.14	0.074	0.29	
Cu <sup>6</sup>	ug/L	2	0.34	0.16	2.0	0.28	0.18	1.3	0.29	0.14	0.67	0.22	0.13	0.29	
$Hg^7$	ng/L	$\overline{2}$	1.0	0.48	2.3	1.5	0.68	4.9	1.2	0.59	2.3	1.1	0.57	$\overline{2}$	
Ni <sup>4</sup>	$\ensuremath{\mathrm{ug}}\xspace/L$	24	0.18	0.11	0.22	0.26	0.10	2.7	0.80	0.10	3.30	0.84	0.14	$\overline{3}$	
Pb <sup>4</sup>	$\mu g/L$	0.9	0.018	0.010	0.11	0.032	0.010	0.34	0.022	0.0093	0.071	0.012	0.0083	0.017	
Sb	$\mu g/L$	5.2	9.2	1.4	25	1.20	0.29	14	0.74	0.29	5.6	0.55	0.29	1.02	
Se	$\mu g/L$	3.1	0.50	0.50	0.51	0.50	0.38	0.61	0.50	0.32	0.67	0.46	0.32	0.5	
SO <sub>4</sub>	mg/L	250	6.3	2.1	12	3.6	1.4	12	3.2	1.4	7.0	2.6	1.4	3.7	
$Zn^4$	ug/L	54	0.81	0.46	1.8	1.0	0.48	6.7	1.15	0.38	3.9	0.63	0.34	1.8	
TDS <sup>8</sup>	mg/L	500	57	27	80	40	21	82	37	20	51	34	20	43	
$NO2 + NO3$	mg/L as N		0.50	0.32	0.58	0.70	0.30	11	0.44	0.26	0.58	0.42	0.25	0.56	

**Table E-3.Summary of site-wide water chemistry model predictions in Meadow Creek at node YP-T-22.**

Notes: Yellow cells represent predicted increases in concentrations relative to the modeled existing conditions. Green cells represent predicted decreases in concentrations relative to the modeled existing conditions. Bold italic numerical values represent concentrations greater than surface water criteria/thresholds identified in the third column.

<sup>1</sup>Unless otherwise noted, the values presented here are the most stringent surface water quality criteria, reported as two significant digits.

<sup>2</sup>Average, minimum and maximum values are calculated based on the monthly predicted concentrations over the indicated time period.

<sup>3</sup>The evaluation threshold is based on the acute aquatic life criterion.

<sup>4</sup>Concentrations are based on a hardness value of 40 mg/L as calcium carbonate, which represents the 5<sup>th</sup> percentile hardness during the driest four months at EFSFSR node YP-SR-10.

<sup>5</sup>The aluminum threshold is based on EPA's nationally recommended criteria guidelines. Values for total hardness, pH, and dissolved organic carbon collected on the same day at USGS gaging stations EFSFSR above Sugar Creek (13311250) and EFSFSR at Stibnite (13311000) from 2014-2022 were used to calculate the chronic criterion. The 5<sup>th</sup> percentile value was selected as the evaluation threshold.

<sup>6</sup>The evaluation threshold was derived using the biotic ligand model per guidance contained in DEQ 2017, and was estimated by applying the lowest  $10<sup>th</sup>$  percentile chronic aquatic life criterion based on regional classifications for the Salmon River Basin, Idaho Batholith, and third order streams.

 $7$ The evaluation threshold is based on recommendations by NMFS (2014).



**Figure E-7.Predicted surface water chemistry (i.e., pH, arsenic, antimony, and mercury) in Meadow Creek (YP-T-22) downstream of the tailing's storage facility, buttress, and Hangar Flats pit backfill (SRK 2021a).** 

### **West End Creek (YP-T-6).**

The West End Creek assessment node YP-T-6 is located just upstream of the confluence with Sugar Creek and represents conditions associated with West End pit area mining and subsequently the West End pit lake. The modeled existing and predicted future surface water concentrations at node YP-T-6 during operations are summarized in Table E-4. Predicted concentrations of pH, antimony, arsenic, and mercury for YP-T-6 are shown in Figure E-8.

During operations, water quality at YP-T-6 is assumed to be the same as upstream water quality since the creek will be routed around the mining facilities and will comprise the majority of the flow at YP-T-6. As such, antimony and arsenic concentrations are expected to decrease relative to modeled existing conditions whereas mercury concentrations are predicted to increase. The changes in concentration for these metals from baseline conditions during this period is fairly substantial. Existing mercury concentrations in West End Creek above the West End Pit area are approximately 50 ng/L and are approximately 4 ng/L below the pit area. This suggests that mechanisms that retain particulate mercury reduce mercury concentrations in the creek between the sample locations upstream and downstream of the pit area. Routing the creek around the mine facilities is assumed to result in higher mercury concentrations near the mouth during operations. Mercury concentrations are predicted to be upwards of ten times greater and well above levels associated with detrimental bioaccumulation, assuming conditions conducive to methylation are present. Chromium concentrations are also predicted to increase, though predicted concentrations are well below the chronic criterion for chromium IV (the most toxic form).

Upon completion of open pit mining, upper West End Creek will flow to the West End pit and predicted concentrations at YP-T-6 during this period are calculated as mixture of outflow from the pit lake. During early- and late-closure periods, average concentrations of arsenic, antimony, and mercury are predicted to return to levels similar to existing conditions, though maximum predicted concentrations will increase. Under existing conditions and predicted future conditions, both predicted average and maximum concentrations of arsenic, antimony, and mercury exceed the evaluation thresholds.

<b>Parameter</b>	<b>Units</b>	<b>Evaluation</b> Threshold <sup>1</sup>	<b>Existing Conditions</b> Mine Year -37 to -3				<b>Open Pit Mining</b> Mine Year -2 to 12			<b>Post-Mining during Water</b> <b>Treatment</b> Mine Year 13 to 40		<b>Post-Mining no Water</b> <b>Treatment</b> Mine Year 41 to 112		
			$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>
pH		$6.5 - 9$	8.2	8.0	8.4	7.7	7.4	7.9	8.2	8.0	8.4	8.2	8.0	8.4
$Ag^{3,4}$	ug/L	0.7	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.011	0.01	0.01	0.011
Al <sup>5</sup>	ug/L	385	2.2	1.3	5	3.3	1.4	$\overline{4}$	2.4	1.3	5	2.2	1.3	5
$\mathbf{A}\mathbf{s}$	ug/L	10	79	64	88	8.6	7.8	8.9	79	64	94	79	64	95
Cd <sup>4</sup>	ug/L	0.3	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.018	0.01	0.01	0.018
$Cr^4$	ug/L	10.6	0.22	0.1	0.37	0.33	0.29	0.48	0.22	0.1	0.37	0.22	0.092	0.37
Cu <sup>6</sup>	ug/L	2	0.25	0.15	0.5	0.13	0.11	0.17	0.26	0.15	0.57	0.26	0.15	0.57
$\rm{Hg^7}$	ng/L	2	4.3	3.7	5.6	53	37	63	4.4	3.7	9.7	4.3	3.7	9.5
Ni <sup>4</sup>	ug/L	24	0.35	0.25	0.52	0.16	0.1	0.27	0.33	0.24	0.52	0.35	0.2	0.52
Pb <sup>4</sup>	ug/L	0.9	0.013	0.01	0.028	0.01	0.01	0.013	0.017	0.01	0.24	0.019	0.01	0.25
Sb	ug/L	5.2	10	7.9	12	2.1	1.8	2.2	10	7.9	14	11	7.9	14
Se	ug/L	3.1	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.43	0.5	0.5	0.39	0.5
$SO_4$	mg/L	250	55	38	94	0.82	0.65	0.9	56	36	94	54	30	94
$Zn^4$	ug/L	54	0.59	0.48	0.85	0.56	0.44	0.8	0.63	0.48	$\overline{2}$	0.62	0.48	2.1
TDS <sup>8</sup>	mg/L	500	192	158	250	119	105	129	194	152	251	191	140	251
$NO2 + NO3$	$mg/L$ as N		0.69	0.53	0.89	1.0	0.28	6.0	0.69	0.5	0.89	0.69	0.43	0.89

**Table E-4.Summary of site-wide water chemistry model predictions in West End Creek at node YP-T-6.**

Notes: Yellow cells represent predicted increases in concentrations relative to the modeled existing conditions. Green cells represent predicted decreases in concentrations relative to the modeled existing conditions. Bold italic numerical values represent concentrations greater than surface water criteria/thresholds identified in the third column.

<sup>1</sup>Unless otherwise noted, the values presented here are the most stringent surface water quality criteria, reported as two significant digits.

<sup>2</sup>Average, minimum and maximum values are calculated based on the monthly predicted concentrations over the indicated time period.

<sup>3</sup>The evaluation threshold is based on the acute aquatic life criterion.

<sup>4</sup>Concentrations are based on a hardness value of 40 mg/L as calcium carbonate, which represents the 5<sup>th</sup> percentile hardness during the driest four months at EFSFSR node YP-SR-10. <sup>5</sup>The aluminum threshold is based on EPA's nationally recommended criteria guidelines. Values for total hardness, pH, and dissolved organic carbon collected on the same day at USGS gaging stations EFSFSR above Sugar Creek (13311250) and EFSFSR at Stibnite (13311000) from 2014-2022 were used to calculate the chronic criterion. The 5th percentile

value was selected as the evaluation threshold.

<sup>6</sup>The evaluation threshold was derived using the biotic ligand model per guidance contained in DEQ 2017, and was estimated by applying the lowest  $10<sup>th</sup>$  percentile chronic aquatic life criterion based on regional classifications for the Salmon River Basin, Idaho Batholith, and third order streams.

 $7$ The evaluation threshold is based on recommendations by NMFS (2014).



**Figure E-8.Predicted surface water chemistry (i.e., pH, arsenic, antimony, and mercury) in West End Creek (YP-T-6) (SRK 2021a).**

# **Sugar Creek (YP-T-1).**

The Sugar Creek assessment node YP-T-1 is located just upstream of its confluence with the EFSFSR and represents conditions associated with upstream surface and groundwater quality, Bailey Tunnel adit seepage, and activities within the West End Creek drainage. The modeled existing and predicted future surface water concentrations at node YP-T-6 during operations are summarized in Table E-5. Predicted concentrations of pH, antimony, arsenic, and mercury for YP-T-6 are shown in Figure E-9.

During operations, most metals concentrations are predicted to remain the same or decrease. Mercury concentrations are predicted to increase slightly, and will continue to exceed the evaluation threshold of 2 µg/L. Lead is also predicted to increase; though concentrations are expected to be an order of magnitude lower than applicable criteria. Arsenic concentrations are predicted to be slightly lower than existing conditions; however, concentrations are expected to remain elevated above applicable criteria.

During early- and late-closure periods, predicted concentrations of arsenic, cadmium, copper, mercury and lead are expected to increase. Cadmium, copper and lead concentrations are expected to be well below their respective evaluation thresholds. Arsenic and mercury concentrations will remain above their respective evaluation thresholds.

<b>Parameter</b> pH	<b>Units</b>	<b>Evaluation</b> Threshold <sup>1</sup>		<b>Existing Conditions</b> Mine Year -37 to -3			<b>Open Pit Mining</b> Mine Year -2 to 12		<b>Post-Mining during Water</b> <b>Treatment</b> Mine Year 13 to 40			<b>Post-Mining no Water</b> <b>Treatment</b> Mine Year 41 to 112		
			$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>
		$6.5 - 9$	7.4	7.2	7.6	7.4	7.2	7.6	7.4	7.2	7.6	7.4	7.3	7.6
$\rm Ag^{3,4}$	ug/L	0.7	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01
Al <sup>5</sup>	ug/L	385	6	2.8	13	6	2.7	13	6	2.7	13	6	2.6	13
As	$\mu g/L$	10	13	7.4	16	13	6.6	15	13	6.8	<b>17</b>	14	$\overline{\tau}$	17
Cd <sup>4</sup>	ug/L	0.3	0.0101	0.01	0.01	0.0101	0.01	0.01	0.0102	0.01	0.011	0.0103	0.01	0.011
$Cr^4$	$\mu g/L$	10.6	0.13	0.1	0.2	0.13	0.1	0.2	0.13	0.1	0.2	0.13	0.1	0.2
$\mathrm{Cu}^6$	ug/L	$\overline{2}$	0.27	0.2	0.6	0.27	0.21	0.6	0.27	0.2	0.6	0.28	0.2	0.6
$\rm Hg^7$	ng/L	$\overline{2}$	6.3	5.47	8.1	6.5	5.55	8.5	6.5	5.61	8.2	6.6	5.71	8.2
Ni <sup>4</sup>	$\mu g/L$	24	0.22	0.11	0.26	0.21	0.11	0.25	0.21	0.11	0.25	0.21	0.11	0.25
Pb <sup>4</sup>	ug/L	0.9	0.018	0.01	0.05	0.019	0.01	0.051	0.022	0.0117	0.057	0.026	0.0145	0.057
Sb	ug/L	5.2	3.6	1.42	6	2.4	1.26	$\overline{3}$	2.5	1.28	3	2.5	1.31	3
Se	$\mu g/L$	3.1	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.49	0.5
$SO_4$	mg/L	250	10.2	4.3	14	8.1	3.3	10	8.1	3.4	10	8.3	3.5	10
$Zn^4$	ug/L	54	0.67	0.31	0.8	0.6	0.29	0.7	0.62	0.28	0.8	0.64	0.29	0.8
TDS <sup>8</sup>	mg/L	500	69	40	81	66	38	75	66	38	75	67	39	76
$NO2 + NO3$	$mg/L$ as N		0.59	0.3	1.11	0.6	0.29	1.13	0.6	0.29	1.12	0.59	0.29	1.12

**Table E-5.Summary of site-wide water chemistry model predictions in Sugar Creek at node YP-T-1.**

Notes: Yellow cells represent predicted increases in concentrations relative to the modeled existing conditions. Green cells represent predicted decreases in concentrations relative to the modeled existing conditions. Bold italic numerical values represent concentrations greater than surface water criteria/thresholds identified in the third column.

<sup>1</sup>Unless otherwise noted, the values presented here are the most stringent surface water quality criteria, reported as two significant digits.

<sup>2</sup>Average, minimum and maximum values are calculated based on the monthly predicted concentrations over the indicated time period.

<sup>3</sup>The evaluation threshold is based on the acute aquatic life criterion.

<sup>4</sup>Concentrations are based on a hardness value of 40 mg/L as calcium carbonate, which represents the 5<sup>th</sup> percentile hardness during the driest four months at EFSFSR node YP-SR-10.  ${}^{5}$ The aluminum threshold is based on EPA's nationally recommended criteria guidelines. Values for total hardness, pH, and dissolved organic carbon collected on the same day at

USGS gaging stations EFSFSR above Sugar Creek (13311250) and EFSFSR at Stibnite (13311000) from 2014-2022 were used to calculate the chronic criterion. The 5<sup>th</sup> percentile value was selected as the evaluation threshold.

<sup>6</sup>The evaluation threshold was derived using the biotic ligand model per guidance contained in DEQ 2017, and was estimated by applying the lowest  $10<sup>th</sup>$  percentile chronic aquatic life criterion based on regional classifications for the Salmon River Basin, Idaho Batholith, and third order streams.

 $7$ The evaluation threshold is based on recommendations by NMFS (2014).



**Figure E-9.Predicted surface water chemistry (i.e., pH, arsenic, antimony, and mercury) in Sugar Creek (YP-T-1) (SRK 2021a)**

# **East Fork South Fork Salmon River Upstream of Sugar Creek**

The following four assessment nodes were used to evaluate water quality conditions in the EFSFSR upstream of Sugar Creek:

- YP-SR-10 is located just downstream of the confluence with Meadow Creek.
- YP-SR-8 is located upstream of the Fiddle Creek confluence.
- YP-SR-6 is located downstream of the Fiddle Creek confluence and represents conditions in the lined stream channel across the YPP (USFS 2023).
- YP-SR-4 is located downstream of the YPP backfill.

Predicted water quality conditions for assessment nodes YP-SR-10, 8, 6, and 4 are summarized in Tables E-6 through E-9, respectively. Predicted concentrations of pH, antimony, arsenic, and mercury for YP-SR-4 are shown in Figure E-10. Similar graphs for YP-SR-10, 8, and 6 can be found in the SWWC report (SRK 2021a; Figures G-5, G-7, and G-9, respectively).

Under existing conditions, water quality at assessment nodes YP-SR-10, 8, and 6 are influenced by the spent ore disposal area (SODA), Bradley tailings, and Hecla Heap. As a result, concentrations of antimony, arsenic, and mercury are elevated above the evaluation thresholds. Reclamation of these facilities during the operations period will is expected to result in decreased concentrations of arsenic and antimony, though concentrations will remain above the evaluation thresholds. Concentrations of mercury and other contaminants are predicted to increase during operations, though only mercury is predicted to exceed its evaluation threshold. At YP-SR-4, removal of unlined legacy mine wastes results in decreases in antimony and arsenic during operations.

During early-closure, concentrations of most constituents are predicted to be higher than existing conditions, which may be attributable to discharge of treated mine contact water. Maximum predicted concentrations of mercury are expected to exceed the evaluation thresholds, while average concentrations are predicted to remain below. Concentrations of antimony and arsenic are predicted to continue to decrease at locations above the YPP, though predicated average and maximum concentrations remain above the evaluation thresholds at most assessment nodes. At assessment node YP-SR-4, concentrations of arsenic increase relative to operations, because discharging groundwater chemistry is modified by interaction with the YPP backfill.

During late-closure, when water treatment is no longer needed, concentrations of many contaminants in the EFSFSR above the YPP are expected to generally be below existing conditions. Exceptions to this include slight increases in predicted average concentrations of silver, cadmium, and nickel at assessment nodes upstream of the YPP. Although quantifiable, concentrations are very similar to existing conditions and none of the predicted concentrations will exceed the evaluation thresholds used. Of particular note is the predicted reductions in arsenic and antimony concentrations relative to existing conditions at all EFSFSR assessment nodes; although concentrations will remain above evaluation thresholds for these two parameters. Similarly, predicted mercury concentrations will remain the same or decrease slightly upstream of the YPP backfill; however, predicted maximum concentrations will remain above the evaluation threshold. Downstream of the YPP backfill, mercury concentrations will be slightly higher than existing concentrations and will be greater than the evaluation threshold.

Parameter	<b>Units</b>	<b>Evaluation</b> Threshold <sup>1</sup>		<b>Existing Conditions</b> Mine Year -37 to -3			<b>Open Pit Mining</b> Mine Year -2 to 12			<b>Treatment</b> Mine Year 13 to 40	<b>Post-Mining during Water</b>		<b>Post-Mining no Water</b> <b>Treatment</b> Mine Year 41 to 112	
			$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>
pH		$6.5 - 9$	7.2	7.0	7.3	7.2	7.0	7.3	7.2	6.3	7.3	7.1	6.8	7.3
$\rm Ag^{3,4}$	ug/L	0.7	0.01	0.0099	0.011	0.011	0.0099	0.032	0.016	0.0100	0.090	0.015	0.0100	0.036
${\rm Al}^5$	ug/L	385	8.8	4.9	21	9	6	21	9	6	21	9	6	21
As	$\rm ug/L$	10	26	7.4	51	8.2	3.2	23.0	6.3	3.0	13.1	5.8	2.8	7.6
Cd <sup>4</sup>	ug/L	0.33	0.01	0.0095	0.011	0.0110	0.0097	0.0250	0.0120	0.0093	0.0410	0.0120	0.0100	0.0210
$\mathrm{Cr}^4$	ug/L	10.6	0.18	0.12	0.34	0.18	0.11	0.70	0.22	0.11	0.91	0.14	0.10	0.35
Cu <sup>6</sup>	ug/L	$\overline{2}$	0.3	0.14	1.5	0.3	0.14	1.0	0.3	0.14	0.7	0.2	0.1	0.5
$\rm Hg^7$	ng/L	$\overline{2}$	1.8	1.2	3.2	2.3	1.5	4.3	1.9	0.9	3.7	1.8	0.9	3.0
Ni <sup>4</sup>	ug/L	24	0.17	0.11	0.22	0.22	0.10	1.80	0.60	0.10	2.50	0.58	0.13	2.40
${\rm Pb^4}$	ug/L	0.9	0.016	0.0098	0.082	0.028	0.0100	0.250	0.022	0.0100	0.170	0.013	0.0097	0.030
${\rm Sb}$	ug/L	5.2	13	2.1	30	2.90	0.42	18.0	1.00	0.40	4.2	0.88	0.40	1.3
Se	$\rm ug/L$	3.1	0.51	0.5	0.52	0.50	0.42	0.58	0.50	0.36	0.60	0.47	0.36	0.50
$SO_4$	mg/L	250	4.8	1.7	8.1	3.0	1.2	9	3.3	1.2	32	2.4	1.2	3
$Zn^4$	ug/L	54	0.85	0.5	1.6	1.00	0.51	4.90	1.13	0.43	8.00	0.73	0.41	1.60
TDS <sup>8</sup>	mg/L	500	51	26	66	41	23	68	40	22	96	37	22	46
$NO2 + NO3$	mg/L as N		0.5	0.38	0.59	0.65	0.39	6.4	0.47	0.35	1.0	0.45	0.34	0.5

**Table E-6.Summary of site-wide water chemistry model predictions in the East Fork South Fork Salmon River at node YP-SR-10.**

Notes: Yellow cells represent predicted increases in concentrations relative to the modeled existing conditions. Green cells represent predicted decreases in concentrations relative to the modeled existing conditions. Bold italic numerical values represent concentrations greater than surface water criteria/thresholds identified in the third column.

<sup>1</sup>Unless otherwise noted, the values presented here are the most stringent surface water quality criteria, reported as two significant digits.

<sup>2</sup>Average, minimum and maximum values are calculated based on the monthly predicted concentrations over the indicated time period.

<sup>3</sup>The evaluation threshold is based on the acute aquatic life criterion.

<sup>4</sup>Concentrations are based on a hardness value of 40 mg/L as calcium carbonate, which represents the 5<sup>th</sup> percentile hardness during the driest four months at EFSFSR node YP-SR-10.

<sup>5</sup>The aluminum threshold is based on EPA's nationally recommended criteria guidelines. Values for total hardness, pH, and dissolved organic carbon collected on the same day at USGS gaging stations EFSFSR above Sugar Creek (13311250) and EFSFSR at Stibnite (13311000) from 2014-2022 were used to calculate the chronic criterion. The 5<sup>th</sup> percentile

value was selected as the evaluation threshold.

<sup>6</sup>The evaluation threshold was derived using the biotic ligand model per guidance contained in DEQ 2017, and was estimated by applying the lowest  $10<sup>th</sup>$  percentile chronic aquatic life criterion based on regional classifications for the Salmon River Basin, Idaho Batholith, and third order streams.

 $7$ The evaluation threshold is based on recommendations by NMFS (2014).

<b>Parameter</b>	<b>Units</b>	<b>Evaluation</b> Threshold <sup>1</sup>	<b>Existing Conditions</b> Mine Year -37 to -3			<b>Open Pit Mining</b> Mine Year -2 to 12			<b>Post-Mining during Water</b> <b>Treatment</b> Mine Year 13 to 40			Post-Mining no Water <b>Treatment</b> Mine Year 41 to 112		
			$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>
pH		$6.5 - 9$	7.2	7.0	7.3	7.2	7.0	7.3	7.2	6.5	7.3	7.2	6.9	7.3
$Ag^{3,4}$	ug/L	0.7	0.01	0.0099	0.011	0.011	0.010	0.027	0.015	0.010	0.072	0.014	0.010	0.031
Al <sup>5</sup>	ug/L	385	8.0	4.9	19	8.0	5.5	19	9.0	5.7	19	8.0	5.4	19
As	ug/L	10	33	18	52	19	12.4	32	17	12.2	25	17	12.1	20
Cd <sup>4</sup>	ug/L	0.33	0.01	0.01	0.01	0.01	0.01	0.02	0.01	0.01	0.03	0.01	0.01	0.019
$Cr^4$	ug/L	10.6	0.18	0.12	0.33	0.19	0.12	0.59	0.22	0.12	0.75	0.16	0.11	0.34
Cu <sup>6</sup>	ug/L	2	0.31	0.15	1.5	0.29	0.15	1.1	0.29	0.15	0.80	0.25	0.14	0.80
$\rm Hg^7$	ng/L	$\overline{2}$	1.7	1.2	3.1	2.1	1.4	3.6	1.8	0.92	3.2	1.7	0.92	3.0
Ni <sup>4</sup>	ug/L	24	0.17	0.11	0.21	0.21	0.10	1.5	0.51	0.10	2.03	0.50	0.14	1.9
Pb <sup>4</sup>	ug/L	0.9	0.016	0.0098	0.084	0.026	0.010	0.20	0.021	0.0098	0.14	0.014	0.0098	0.043
Sb	ug/L	5.2	16.6	6.2	31	8.4	4.5	21	6.9	4.4	<i>12</i>	6.7	4.3	9.0
Se	ug/L	3.1	0.51	0.50	0.52	0.50	0.43	0.56	0.50	0.39	0.58	0.48	0.39	0.50
$SO_4$	mg/L	250	5.0	2.0	8.0	3.6	1.6	8.0	3.8	1.6	26	3.1	1.6	4.0
$Zn^4$	ug/L	54	0.84	0.52	1.5	0.98	0.52	4.0	1.1	0.51	6.4	0.75	0.45	1.4
TDS <sup>8</sup>	$mg/L$	500	52	29	66	44	26	67	44	25	89	41	25	51
$NO2 + NO3$	$mg/L$ as N		0.49	0.41	0.61	0.62	0.42	5.0	0.47	0.38	0.85	0.46	0.38	0.56

**Table E-7.Summary of site-wide water chemistry model predictions in the East Fork South Fork Salmon River at nodeYP-SR-8.**

Notes: Yellow cells represent predicted increases in concentrations relative to the modeled existing conditions. Green cells represent predicted decreases in concentrations relative to the modeled existing conditions. Bold italic numerical values represent concentrations greater than surface water criteria/thresholds identified in the third column.

<sup>1</sup>Unless otherwise noted, the values presented here are the most stringent surface water quality criteria, reported as two significant digits.

<sup>2</sup>Average, minimum and maximum values are calculated based on the monthly predicted concentrations over the indicated time period.

<sup>3</sup>The evaluation threshold is based on the acute aquatic life criterion.

<sup>4</sup>Concentrations are based on a hardness value of 40 mg/L as calcium carbonate, which represents the 5<sup>th</sup> percentile hardness during the driest four months at EFSFSR node YP-SR-10.  ${}^{5}$ The aluminum threshold is based on EPA's nationally recommended criteria guidelines. Values for total hardness, pH, and dissolved organic carbon collected on the same day at

USGS gaging stations EFSFSR above Sugar Creek (13311250) and EFSFSR at Stibnite (13311000) from 2014-2022 were used to calculate the chronic criterion. The 5th percentile value was selected as the evaluation threshold.

<sup>6</sup>The evaluation threshold was derived using the biotic ligand model per guidance contained in DEQ 2017, and was estimated by applying the lowest  $10<sup>th</sup>$  percentile chronic aquatic life criterion based on regional classifications for the Salmon River Basin, Idaho Batholith, and third order streams.

 $7$ The evaluation threshold is based on recommendations by NMFS (2014).

<b>Parameter</b>	<b>Units</b>	<b>Existing Conditions</b> <b>Mine Year -37 to -3</b> <b>Evaluation</b> Threshold <sup>1</sup>					<b>Open Pit Mining</b> <b>Mine Year -2 to 12</b>			<b>Post-Mining during Water</b> <b>Treatment</b> Mine Year 13 to 40		<b>Post-Mining no Water</b> <b>Treatment</b> Mine Year 41 to 112			
			$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	Avg <sup>2</sup>	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	
pH		$6.5 - 9$	7.3	7.0	7.6	7.3	7.0	7.6	7.3	6.9	7.5	7.3	6.9	7.6	
$Ag^{3,4}$	ug/L	0.7	0.01	0.0098	0.011	0.011	0.01	0.021	0.014	0.01	0.048	0.013	0.0102	0.024	
Al <sup>5</sup>	$\mu g/L$	385	9	5.3	20	9	5.5	19	9	5.7	20	9	5.5	19	
As	$\mu g/L$	10	31	16.9	41	23	12.6	41	20	11.9	29	20	11.8	27	
Cd <sup>4</sup>	$\mu g/L$	0.33	0.0103	0.0098	0.011	0.0111	0.0099	0.018	0.0115	0.0094	0.025	0.0117	0.0103	0.016	
$Cr^4$	ug/L	10.6	0.22	0.15	0.31	0.22	0.14	0.48	0.24	0.14	0.57	0.2	0.13	0.32	
Cu <sup>6</sup>	ug/L	2	0.36	0.18	1.6	0.35	0.18	1.3	0.35	0.18	1.1	0.31	0.17	1.1	
$\rm Hg^7$	ng/L	2	1.8	1.22	3.1	2.1	1.26	3.4	1.8	1.01	3.2	1.8	1.02	$\mathfrak{Z}$	
Ni <sup>4</sup>	ug/L	24	0.3	0.11	0.45	0.33	0.11	1.2	0.54	0.11	1.57	0.53	0.21	1.51	
$\rm Pb^4$	$\mu g/L$	0.9	0.023	0.0106	0.093	0.035	0.0108	0.144	0.026	0.0104	0.099	0.021	0.0101	0.065	
Sb	$\mu g/L$	5.2	18.7	5.94	30	13.4	4.93	27	11.9	4.51	20	11.8	4.47	19	
Se	$\mu g/L$	3.1	0.5	0.49	0.51	0.49	0.45	0.53	0.49	0.42	0.55	0.48	0.42	0.5	
$SO_4$	mg/L	250	5.8	2.0	9	4.9	1.7	10	4.9	1.6	20	4.4	1.6	$\overline{7}$	
$Zn^4$	ug/L	54	1.01	0.58	1.6	1.14	0.59	3.2	1.16	0.58	4.5	0.93	0.54	1.4	
TDS <sup>8</sup>	mg/L	500	63	28	81	58	26	86	56	25	96	55	25	74	
$NO2 + NO3$	$mg/L$ as N		0.55	0.47	0.75	0.64	0.48	3.41	0.53	0.45	0.8	0.52	0.43	0.68	

**Table E-8.Summary of site-wide water chemistry model predictions in the East Fork South Fork Salmon River at node YP-SR-6.**

Notes: Yellow cells represent predicted increases in concentrations relative to the modeled existing conditions. Green cells represent predicted decreases in concentrations relative to the modeled existing conditions. Bold italic numerical values represent concentrations greater than surface water criteria/thresholds identified in the third column.

<sup>1</sup>Unless otherwise noted, the values presented here are the most stringent surface water quality criteria, reported as two significant digits.

<sup>2</sup>Average, minimum and maximum values are calculated based on the monthly predicted concentrations over the indicated time period.

<sup>3</sup>The evaluation threshold is based on the acute aquatic life criterion.

<sup>4</sup>Concentrations are based on a hardness value of 40 mg/L as calcium carbonate, which represents the 5<sup>th</sup> percentile hardness during the driest four months at EFSFSR node YP-SR-10.  ${}^{5}$ The aluminum threshold is based on EPA's nationally recommended criteria guidelines. Values for total hardness, pH, and dissolved organic carbon collected on the same day at

USGS gaging stations EFSFSR above Sugar Creek (13311250) and EFSFSR at Stibnite (13311000) from 2014-2022 were used to calculate the chronic criterion. The 5th percentile value was selected as the evaluation threshold.

<sup>6</sup>The evaluation threshold was derived using the biotic ligand model per guidance contained in DEQ 2017, and was estimated by applying the lowest  $10<sup>th</sup>$  percentile chronic aquatic life criterion based on regional classifications for the Salmon River Basin, Idaho Batholith, and third order streams.

 $7$ The evaluation threshold is based on recommendations by NMFS (2014).

<b>Parameter</b>	<b>Units</b>	<b>Evaluation</b> Threshold <sup>1</sup>	<b>Existing Conditions</b> Mine Year -37 to -3				<b>Open Pit Mining</b> Mine Year -2 to 12			<b>Post-Mining during Water</b> <b>Treatment</b> Mine Year 13 to 40		<b>Post-Mining no Water</b> <b>Treatment</b> Mine Year 41 to 112			
			$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	
pH		$6.5 - 9$	7.3	7.0	7.5	7.3	7.0	7.5	7.3	7.0	7.5	7.3	6.9	7.4	
$Ag^{3,4}$	ug/L	0.7	0.010	0.0098	0.011	0.010	0.0063	0.012	0.011	0.0063	0.031	0.010	0.0057	0.017	
Al <sup>5</sup>	$\mu g/L$	385	9.0	6.6	19	9.0	4.0	19	7.0	4.3	<b>20</b>	7.0	3.9	19	
As	$\mu g/L$	10	64	18.7	117	25	12.6	97	35	13	63	34	14.1	62	
Cd <sup>4</sup>	$\mu g/L$	0.3	0.01	0.01	0.01	0.01	0.006	0.02	0.01	0.008	0.02	0.01	0.007	0.02	
$Cr^4$	$\mu g/L$	10.6	0.20	0.15	0.29	0.22	0.11	0.48	0.20	0.13	0.45	0.16	0.12	0.31	
Cu <sup>6</sup>	$\mu g/L$	$\overline{2}$	0.34	0.21	1.4	0.34	0.17	1.2	0.30	0.17	0.90	0.27	0.17	0.80	
$\rm Hg^7$	$\text{ng/L}$	$\overline{2}$	1.2	0.17	3.0	2.0	0.46	3.4	1.9	0.96	3.4	1.6	0.91	3.3	
Ni <sup>4</sup>	$\mu g/L$	24	0.27	0.11	0.40	0.32	0.11	1.2	0.42	0.11	1.1	0.40	0.21	1.0	
Pb <sup>4</sup>	$\mu g/L$	0.9	0.018	0.011	0.080	0.034	0.011	0.14	0.055	0.013	0.20	0.040	0.011	0.18	
Sb	$\mu g/L$	5.2	33.4	7.7	56	14.5	4.9	63	13.3	5.0	23	13	5.4	23	
Se	$\mu g/L$	3.1	0.50	0.49	0.51	0.49	0.25	0.51	0.39	0.29	0.51	0.38	0.26	0.49	
SO <sub>4</sub>	mg/L	250	16.1	2.5	32	5.7	1.7	26	6.8	1.7	17	6.5	1.8	10	
$Zn^4$	$\mu g/L$	54	1.4	0.65	1.8	1.1	0.54	3.1	1.2	0.64	4.2	0.97	0.61	2.0	
TDS <sup>8</sup>	mg/L	500	67	29	97	57	26	98	52	26	88	50	26	71	
$NO2 + NO3$	$mg/L$ as $N$		0.42	0.27	0.62	0.60	0.32	2.3	0.44	0.36	0.65	0.43	0.35	0.56	

**Table E-9.Summary of site-wide water chemistry model predictions in the East Fork South Fork Salmon River at node YP-SR-4.**

Notes: Yellow cells represent predicted increases in concentrations relative to the modeled existing conditions. Green cells represent predicted decreases in concentrations relative to the modeled existing conditions. Bold italic numerical values represent concentrations greater than surface water criteria/thresholds identified in the third column.

<sup>1</sup>Unless otherwise noted, the values presented here are the most stringent surface water quality criteria, reported as two significant digits.

<sup>2</sup>Average, minimum and maximum values are calculated based on the monthly predicted concentrations over the indicated time period.

<sup>3</sup>The evaluation threshold is based on the acute aquatic life criterion.

<sup>4</sup>Concentrations are based on a hardness value of 40 mg/L as calcium carbonate, which represents the 5<sup>th</sup> percentile hardness during the driest four months at EFSFSR node YP-SR-10.  ${}^{5}$ The aluminum threshold is based on EPA's nationally recommended criteria guidelines. Values for total hardness, pH, and dissolved organic carbon collected on the same day at

USGS gaging stations EFSFSR above Sugar Creek (13311250) and EFSFSR at Stibnite (13311000) from 2014-2022 were used to calculate the chronic criterion. The 5th percentile value was selected as the evaluation threshold.

<sup>6</sup>The evaluation threshold was derived using the biotic ligand model per guidance contained in DEQ 2017, and was estimated by applying the lowest  $10<sup>th</sup>$  percentile chronic aquatic life criterion based on regional classifications for the Salmon River Basin, Idaho Batholith, and third order streams.

 $7$ The evaluation threshold is based on recommendations by NMFS (2014).



**Figure E-10.Predicted surface water chemistry (i.e., pH, arsenic, antimony, and mercury) in the East Fork South Fork Salmon River (YP-SR-4), below the Yellow Pine Pit (SRK 2021a).**

### **East Fork South Fork Salmon River downstream of Sugar Creek**

The East Fork South Fork Salmon River assessment node (YP-SR-2) is downstream of Sugar Creek represents conditions associated with all proposed mine operation and closure activities. The modeled existing and predicted future surface water concentrations at node YP-SR-2 during operation, early-closure, and late-closure periods are summarized in Table E-10. Predicted concentrations of pH, antimony, arsenic, and mercury for YP-SR-2 are shown in Figure E-11.

During operations, predicted surface water chemistry is similar to existing conditions, with some variability in predicted antimony, arsenic, and mercury concentrations. Antimony and arsenic concentrations are expected to be lower during the operating period due to the removal of unlined legacy mine wastes. Mercury concentrations are slightly higher than existing conditions during the operating period, due to the predicted discharge from the mine contact water treatment plant.

During early closure, concentrations of antimony and arsenic are expected to increase (Figure E-11), but will remain above their respective evaluation thresholds. However, concentrations will be less than existing conditions. Mercury concentrations are expected to remain elevated above 2 ng/L; however, concentrations are expected to slightly decrease relative to the operations period.

During late closure, most contaminant concentrations will be lower than existing conditions, with the exception of mercury, nickel, and lead. These constituents are predicted to be slightly above existing conditions. Concentrations of nickel and lead will be well below their respective evaluation thresholds, whereas mercury is expected to be above 2 ng/L. Most of the mercury loading at this assessment node is due to legacy mining activities at the Cinnabar Mine Site in the Sugar Creek drainage.

<b>Parameter</b>	<b>Units</b>	<b>Evaluation</b> Threshold <sup>1</sup>		<b>Existing Conditions</b> Mine Year -37 to -3				<b>Open Pit Mining</b> Mine Year -2 to 12		<b>Post-Mining during Water</b> <b>Treatment</b> Mine Year 13 to 40		<b>Post-Mining no Water</b> <b>Treatment</b> Mine Year 41 to 112		
			$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>	$Avg^2$	Min <sup>2</sup>	Max <sup>2</sup>
pH		$6.5 - 9$	7.3	7.1	7.5	7.3	7.1	7.5	7.3	7.0	7.5	7.3	7.0	7.5
$Ag^{3,4}$	$\mu g/L$	0.7	0.01	0.0099	0.011	0.01	0.0077	0.012	0.011	0.0077	0.023	0.01	0.0074	0.015
Al <sup>5</sup>	ug/L	385	8	5.1	17	$\overline{7}$	4.8	17	$\overline{7}$	$\overline{4}$	17	$\overline{7}$	3.8	17
As	ug/L	10	45	14.1	76	20	10.2	66	28	10.4	45	27	11.3	47
Cd <sup>4</sup>	ug/L	0.3	0.0102	0.0099	0.01	0.0106	0.0075	0.016	0.0105	0.0086	0.018	0.0101	0.0082	0.013
Cr <sup>4</sup>	$\mu g/L$	10.6	0.17	0.14	0.23	0.18	0.12	0.37	0.17	0.13	0.33	0.15	0.12	0.25
Cu <sup>6</sup>	ug/L	2	0.32	0.23		0.31	0.2	0.8	0.29	0.2	0.6	0.28	0.2	0.6
$\rm Hg^7$	ng/L	2	4.8	3.2	9.6	5.7	3.66	10	5.2	3.23	9.3	5.2	2.96	9.4
$\mathrm{Ni}^4$	ug/L	24	0.25	0.11	0.35	0.28	0.11	0.89	0.35	0.11	0.87	0.33	0.18	0.84
Pb <sup>4</sup>	ug/L	0.9	0.018	0.011	0.056	0.028	0.0107	0.103	0.043	0.0211	0.139	0.035	0.0204	0.127
Sb	ug/L	5.2	22.2	5.2	37	9.7	3.51	41	9.5	3.45	16	9.3	3.82	16
Se	$\mu g/L$	3.1	0.5	0.49	0.51	0.49	0.35	0.5	0.43	0.36	0.51	0.42	0.35	0.49
SO <sub>4</sub>	mg/L	250	13.8	3.2	23	6.7	2.3	20	7.3	2.4	15	7.1	2.5	10
$Zn^4$	ug/L	54	1.11	0.64	1.4	0.93	0.57	2.3	1.02	0.63	2.9	0.86	0.6	1.5
TDS <sup>8</sup>	mg/L	500	67	33	87	61	31	88	57	30	82	56	30	71
$NO2 + NO3$	$mg/L$ as N		0.49	0.32	0.71	0.6	0.31	.68	0.5	0.34	0.72	0.49	0.33	0.72

**Table E-10.Summary of site-wide water chemistry model predictions in the East Fork South Fork Salmon River at node YP-SR-2**

Notes: Yellow cells represent predicted increases in concentrations relative to the modeled existing conditions. Green cells represent predicted decreases in concentrations relative to the modeled existing conditions. Bold italic numerical values represent concentrations greater than surface water criteria/thresholds identified in the third column.

<sup>1</sup>Unless otherwise noted, the values presented here are the most stringent surface water quality criteria, reported as two significant digits.

<sup>2</sup>Average, minimum and maximum values are calculated based on the monthly predicted concentrations over the indicated time period.

<sup>3</sup>The evaluation threshold is based on the acute aquatic life criterion.

<sup>4</sup>Concentrations are based on a hardness value of 40 mg/L as calcium carbonate, which represents the 5<sup>th</sup> percentile hardness during the driest four months at EFSFSR node YP-SR-10.

<sup>5</sup>The aluminum threshold is based on EPA's nationally recommended criteria guidelines. Values for total hardness, pH, and dissolved organic carbon collected on the same day at USGS gaging stations EFSFSR above Sugar Creek (13311250) and EFSFSR at Stibnite (13311000) from 2014-2022 were used to calculate the chronic criterion. The 5th percentile value was selected as the evaluation threshold.

<sup>6</sup>The evaluation threshold was derived using the biotic ligand model per guidance contained in DEQ 2017, and was estimated by applying the lowest  $10<sup>th</sup>$  percentile chronic aquatic life criterion based on regional classifications for the Salmon River Basin, Idaho Batholith, and third order streams.

 $7$ The evaluation threshold is based on recommendations by NMFS (2014).



**Figure E-11.Predicted surface water chemistry (i.e., pH, arsenic, antimony, and mercury) in the East Fork South Fork Salmon River (YP-SR-2), below the Sugar Creek confluence (SRK 2021a).**
#### REFERENCES

- Brown and Caldwell. 2021. Final Stibnite Gold Project, Stibnite Hydrologic Site Model Refined Modified Proposed Action (ModPRO2) Report. Prepared for Perpetua Resources Idaho, Inc. August 2021.
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- Perpetua. 2021. Stibnite Gold Project Site-Wide Water Balance Model Refined Modified Proposed Action (ModPRO2) Report. October 2021.
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### **APPENDIX F**

#### **SUMMARY OF TOXICOLOGICAL EFFECTS FOR SELECT CONSITUENTS**

#### **INTRODUCTION**

This appendix summarizes toxicity information for constituents of potential concern associated with the Stibnite Gold Project. For each contaminant, a general discussion about its toxicity is followed by sections specific to fish and their prey. A discussion about prey items is included because aquatic macroinvertebrates serve as significant food sources for early life stages of Endangered Species Act (ESA)-listed fish, as well as for other aquatic organisms that are in turn prey items for salmon and steelhead. The toxicity for many constituents were examined as part of the consultation for Idaho water quality standards for toxics (NMFS 2014; WCRO-2000-1484), and much of that information has been incorporated here. We also incorporated best available science that has been published since 2014.

For ease of use, a general description of common toxicological terminology used in this section is provided for quick reference below. The terminology is organized alphabetically, by its respective acronym.

**Bioaccumulation Factor (BAF).** A measure of how readily a contaminant concentrates in biotic tissues relative to all potential exposure routes (e.g., waterborne and dietary exposures).

**Bioconcentration Factor (BCF).** A measure of how readily a contaminant concentrates in biotic tissues from waterborne exposures only. For aquatic toxicology, the BCF is the ratio of a contaminant concentration in biota (i.e., tissue concentrations) to its concentration in water.

**Criterion Continuous Concentration (CCC).** More commonly referred to as the "chronic criterion." Idaho water quality standards (WQS) (Idaho Administrative Procedures Act 58.01.02) contain chronic criteria for many of the23000 parameters of concern included in this biological opinion (Opinion). The CCC is established at a level (magnitude) that protects aquatic communities against long-term effects and that is not to be exceeded on average for longer than 4 days (duration) and more than once every 3 years (frequency).

**Criterion Maximum Concentration (CMC).** More commonly referred to as the "acute criterion," this is an enforceable criterion that is included in Idaho WQS for the protection of aquatic life. The concentration is established at a level that protects aquatic communities against short-term effects and that is not to be exceeded on average for longer than 1-hour and more than once every 3 years.

**Effect Concentration (ECx).** The estimated concentration at which a specified proportion of the test organisms exhibit the endpoint (e.g., reduced growth, immobility, loss of equilibrium, etc.) of interest. For example, immobility was observed in 20 percent of the test organisms. Or, the estimated concentration at which an organism demonstrated a specified proportional reduction in a given endpoint (e.g., growth, reproduction, etc.). For example, test organisms experienced a

20 percent reduction in growth compared to controls. The later definition is synonymous with the inhibition concentration (IC) (listed below).

**Genus Mean Acute Value (GMAV).** The geometric mean of the species means acute values for an identified toxicity test statistic (most commonly the  $LC_{50}$ ).

Inhibition Concentration  $(IC_x)$ . The concentration estimated to cause a specified percent reduction in a biological test endpoint such as reproduction or growth.

**Lethal Concentration (LCx).** The estimated concentration where a specified percentage of the test organisms die. The  $LC_x$  is estimated using a statistical distribution or regression model of the dose-response relationship based on toxicity testing. The most common test statistic used in acute studies is the  $LC_{50}$  (the lethal concentration where 50 percent of the organisms die).

**Lowest Observed Effect Concentration (LOEC).** The lowest tested concentration in which the measured endpoints are statistically significantly different from the controls. This is the first treatment that is greater than the No Observed Effect Concentration (NOEC) (defined below).

**Maximum Acceptable Threshold Concentration (MATC).** The geometric mean of the NOEC and the LOEC for the most sensitive endpoint (such as mortality, reproduction, or growth) for which data is available.

**No Observed Effect Concentration.** The highest tested concentration in which the measured endpoints are not statistically significantly different from the controls.

**Species Mean Acute Value (SMAV).** The geometric mean of individual toxicity test values  $(e.g., LC_{50}s)$  for a species.

# **ANTIMONY**

Antimony (Sb) naturally occurs in soil and water and is generally present in low concentrations. Antimony is present in the environment in numerous oxidative states in both organic and inorganic forms (Bolan et al. 2022). The two oxidation states of environmental and biological importance include antimonite (i.e.,  $Sb^{+3}$ , or  $Sb(III)$ ) and antimonate (i.e.,  $Sb^{+5}$  or  $Sb(V)$ ) (Filella et al. 2009, 2002). Typical concentrations of Sb in uncontaminated freshwater are generally below 1 ug/L (Filella et al. 2009); observed background concentrations in Meadow Creek (upstream of historic mining impacts) have typically been well below 1 ug/L (Etheridge 2015).

Antimony is not an essential element, meaning it does not have any known biological function. The toxicity of Sb is not well studied, though it is thought to be less toxic than arsenic (Filella et al. 2009). Most toxicity studies involve Sb(III), which is not as prevalent as Sb(V) in oxic systems such as rivers and streams. Antimony mobility and bioavailability (and hence its toxicity), is dependent upon pH, redox potential, and organic matter (He et al. 2019, Bolan et al. 2022, Obiakor et al. 2017a). The mode of toxicity is not yet understood.

Most of the antimony toxicity research has focused on Sb(III). There is some disagreement in the literature regarding which oxidized state is more toxic, and many claims of Sb(III) being more toxic appear to be unfounded (Filella et al. 2009). Some authors suggest Sb(III) is more toxic than Sb(V); Kennedy (2023, as cited in B.C. Ministry of Water, Land and Resource Stewardship 2023) did not observe a significant difference in toxicity of Sb(III) and Sb(V), suggesting nearly equivalent toxicity. Sb(V) is the dominant oxidized state in Meadow Creek (Dovik et al. 2016).

# **Toxicity to Fish**

Antimony toxicity is understudied relative to metals such as arsenic, copper, and mercury; and available information examines waterborne exposures. No studies regarding dietborne exposure were found. Recently, Canada published draft recommended chronic and acute water quality guidelines of 74 ug/L and 250 ug/L, respectively, to protect aquatic life (B.C. Ministry of Water, Land and Resource Stewardship 2023). These recommendations were derived from a species sensitivity distribution of toxicity data for chronic tests on aquatic organisms. The HC<sub>5</sub> estimation was 445 ug/L (upper and lower  $95<sup>th</sup>$  percentile confidence intervals of 175 and 1,400 ug/L respectively), and divided by a selected adjustment factor of 6 (range from 29 to 233). In 1988, EPA reported a final acute value of 175 µg/L and after application of an adjustment factor of 2, recommended an acute criterion of 88 µg/L to protect aquatic life. An acute-to-chronic ratio from one test was applied to the final acute value (FAV) to derive a chronic criterion recommendation of 30  $\mu$ g/L (EPA 1988). More recently, Obiakor et al. (2017a) examined acute toxicity of antimony to aquatic organisms and reported a predicted hazardous concentration (the concentration protective of 95 percent of species within the species sensitivity distribution) of 781 µg/L (lower- and upper-95 percent confidence intervals of 580 and 1,203 µg/L, respectively). In examined studies, insects were more sensitive to antimony than crustaceans and fish. Studies underlying one, or more, of the above recommendations are presented below.

MacPhee and Ruelle (1969) identify similarity in the toxicity (LC50 >10,000 ug/L) of Sb(III) and Sb(V) on steelhead, Chinook salmon, and coho salmon; although the authors used older juvenile fish (which are generally less sensitive) in their experiments. Brooke et al. (1986) exposed rainbow trout fry to two concentrations of antimony for 96 hours. While an  $LC_{50}$  could not be determined, deaths were observed after 24 hours and 45 percent of the fish died after a 96 hour exposure to 25,700  $\mu$ g/L of antimony. No deaths were reported for the 11,400  $\mu$ g/L exposure. The criteria recommendations previously mentioned are largely driven by nonsalmonid species.

Doe et al. (1987) reported an LC<sub>50</sub> of 16,000  $\mu$ g/L for juvenile rainbow trout exposed to antimony for 30 days. A more recent study found similar results for sockeye salmon. Kennedy (2023, as cited in B.C. Ministry of Water, Land and Resource Stewardship 2023) exposed juvenile sockeye salmon to antimony for 30 days and reported a NOEC of 12,800 µg/L and a LOEC of 25,600  $\mu$ g/L. Unfortunately, the details of this study were not available for critical review. Kennedy (2020, as cited in B.C. Ministry of Water, Land and Resource Stewardship 2023) also exposed larval rainbow trout to antimony for 30 days and reported a NOEC of 3,200 µg/L for both growth and survival endpoints.

Research suggests that early life stages (embryos and larvae) may be more susceptible to chronic antimony exposures than older juvenile stages. Birge (1978) exposed rainbow trout eggs from fertilization through 4 days post-hatch (28-day test) and reported an  $LC_{50}$  of 580  $\mu$ g/L (upperand lower-95 percent confidence intervals of 340  $\mu$ g/L and 920  $\mu$ g/L, respectively) and an LC<sub>1</sub> of 28.6 µg/L (upper- and lower-95 percent confidence intervals of 4.66 and 72.2 µg/L, respectively). LeBlanc and Dean (1984) reported a NOEC for mortality of fathead minnow (*Pimephales promelas*) embryos exposed to antimony at concentrations up to 7.5 µg/L in water. Similarly, when exposed to concentrations of up to 7.5 ug/L, larval fathead minnow survival and growth did not exhibit a consistent dose-response and the authors concluded the test concentrations were not toxic to fathead minnows. The authors did not test higher concentrations. In other aquatic toxicity tests with fathead minnow eggs, growth was the most sensitive endpoint (Kimball 1978). The 28-d LOEC for effects on growth (length) was 2,310  $\mu$ g/L. There were no significant effects on growth at 1,130  $\mu$ g/L. Xia et al (2021) examined developmental toxicity of antimony in zebrafish embryos. The authors exposed embryos (5 hours post fertilization) to high concentrations of antimony and evaluated mortality, development, and growth. Embryos exposed to all tested concentrations (73 to 292 mg/L) of antimony exhibited pericardial edema, cerebral hemorrhaging, uninflated swimming bladders, and/or a curved spine. There was no significant difference in survival 24 hours post fertilization. If these fish were to survive through hatching, such deformities would impact their fitness.

Data is limited on bioaccumulation and food web transfer of antimony; however, the current best available information indicates antimony does not biomagnify and bioaccumulation is limited in higher trophic organisms, such as fish (Obiakor et al. 2017b). Culioli et al (2009) found that trophic transfer of antimony from primary producers to trout was low; and there was not substantial accumulation of antimony in fish. Similarly, Dovick et al. (2016) studied antimony concentrations in biota at the Stibnite Mine site and found biodimunition of antimony with higher trophic levels. No studies relating bioaccumulation and adverse effects were found.

# **Toxicity to Aquatic Invertebrates**

There is a large variation in toxicity of antimony to invertebrates, with  $LC_{50}$  values ranging from 687 to 29,600 ug/L (ECOTOX database; Olker et al. 2022). Brooke et al. (1986) investigated caddisfly susceptibility to antimony toxicity and was not able to establish a dose response because only 5 percent of the test organism died in the highest exposure concentration (25.7 mg/L). Estimated LC50 values for a midge (*Chironomus tentans*) were reported as 4.1 and 5.3 mg/L (ECOTOX database). In a life-cycle test with the cladoceran *Daphnia magna*, a 28-d LC<sub>50</sub> of 4,510 µg/L was calculated for exposure to antimony trichloride in water of hardness 220 mg/L as CaCO3 (Kimball 1978). There were no effects on reproduction at 3,900 µg/L but there was a significant decrease in the number of progeny at 7,050  $\mu$ g/L. Brooke et al. (1986) also investigated sublethal effects (clubbed tentacles or reduced growth) associated with antimony exposure. The authors reported an EC<sub>50</sub> of 500 ug/L for adult *Hydra sp*.

Regarding bioaccumulation of antimony, Culioli et al (2009) observed differences among invertebrate functional feeding groups, with shredders accumulating the most antimony and predators accumulating the least. This indicates the dietary update route is a driver of antimony accumulation rather than water column concentrations.

### **Summary**

Species response to antimony exposures vary drastically and the science remains scarce relative to other metals. For the purposes of this assessment, we have elected to utilize the EPA recommendations of 88 and 30 ug/L for characterizing the risk of adverse effects to ESA-listed species and their prey resulting from acute and chronic exposures, respectively. Considering Sb(V) is more prevalent in the aquatic environment and the recent work evaluating Sb(III) and Sb(V) toxicities, comparing concentrations of total dissolved Sb to toxicity thresholds developed from Sb(III) toxicity tests is deemed appropriate for our current assessment.

#### **ARSENIC**

Arsenic is ubiquitous; it is present in air, water, and soil. Usually, arsenic concentrations in surface waters are low, ranging from 0.1 to 2.0  $\mu$ g/L. In a probabilistic study of arsenic in 55 Idaho rivers, the median total arsenic concentration was 2.0 µg/L, ranging from 0.06 to 17 µg/L, from unfiltered samples (Essig 2010).

The toxicity of arsenic is influenced by a number of factors including water temperature, pH, redox potential, organic matter, phosphate content, suspended solids, presence of other toxicants, and chemical speciation (Dabrowski 1976; Eisler 1988a; McGeachy and Dixon 1989; Sorensen 1991; Rankin and Dixon 1994; McIntyre and Linton 2011). Toxicity of arsenic does not appear to vary with hardness (Borgmann et al. 2005). Trivalent arsenic tends to be more toxic than other forms, and inorganic forms of arsenic are typically more toxic than organic forms (EPA 1985a; Eisler 1988a; Sorensen 1991).

# **Toxicity to Fish**

Arsenic can enter fish through their diet or through absorption through the skin and gills. When ingested, arsenic can be converted to other forms such as the less toxic arsenocholine or arsenobetaine (Malik et al. 2023). Arsenic is a suspected carcinogen in fish and is known to bioaccumulate in fish tissue. It is associated with necrotic and fibrous tissues and cell damage, especially in the liver. At high enough concentrations, arsenic can result in immediate death through increased mucus production and suffocation. Other effects include anemia and gallbladder inflammation. Juvenile salmonids have been found to be more sensitive to arsenic toxicity than alevins (Buhl and Hamilton 1990; 1991). Arsenic does not readily bioconcentrate in aquatic species. Robinson et al. (1995, in EPA 1999) found no evidence of arsenic uptake or accumulation from water in both rainbow and brown trout (*Salmo trutta*). As described below, research suggests that adverse effects in fish from arsenic are most likely from dietary rather than waterborne exposures.

Arsenic is not very toxic in classic toxicity tests with exposures through water (McIntyre and Linton 2011). Acute toxicity appears to occur at concentrations that are significantly higher than the CMC of 340  $\mu$ g/L (Buhl and Hamilton 1990). Erickson et al. (2011) observed very little difference in survival of juvenile rainbow trout exposed to 8 mg/L compared to controls. When testing higher concentrations of arsenic (i.e., 16 and 32 mg/L), most mortalities occurred within

2 days (32 mg/L exposure) and 6 days (16 mg/L exposure). A 96-hour LC $_{50}$  of 10.8 mg/L was reported for rainbow trout (Hale 1977).

Birge et al. (1978; 1981) suggest that chronic arsenic toxicity from waterborne exposures occurs to developing embryos of listed salmonids at concentrations below Idaho's CCC of 150 µg/L. Rainbow trout embryos were exposed to arsenic for 28 days (4-days post-hatching) at 12°C (53.6°F) to 13°C (55.4°F) and a hardness of 93 mg/L to 105 mg/L CaCO<sub>3</sub> in static tests. Concentrations of 42 to 134 μg/L were estimated to be associated with the onset of mortality, as LC<sup>1</sup> and LC<sup>10</sup> respectively (Birge *et al*. 1980). Studies reviewed in Eisler (1988a) and EPA (1985a) indicate that chronic effects do not occur in other life stages until concentrations are at least about an order of magnitude higher than the levels determined by Birge et al. (1978; 1981) to be detrimental to developing embryos. Nichols et al. (1984) described significant chronic effects to coho salmon (*O. kisutch*) from a 6-month exposure to 33 µg/L of arsenic trioxide, where the normal increase in plasma thyroxine was delayed, causing a transitory reduction in gill sodium-potassium ATPase activity. Although treated fish showed no direct effects in growth or survival, they were less successful in seaward migration than control fish. Rankin and Dixon (1994) observed significant reductions in growth (believed to be caused by reduced appetite and direct metabolic impacts) at chronic exposures to waterborne arsenite concentrations of 9.64 mg/L. In addition, trout exposed to this concentration also suffered 10 percent mortality and all fish exposed to this concentration showed inflammation of the gallbladder wall.

Studies have shown that inorganic arsenic in the diet of rainbow trout are associated with reduced growth, organ damage, and other physiological effects at concentrations in food items of about 20 milligrams/kilogram (mg/kg) dry weight (dw) and above (Cockell et al. 1991; Hansen et al. 2004; Erickson et al. 2010, Erickson et al. 2011). Hansen et al. (2004) collected metalscontaminated sediments (170 mg/kg arsenic) from the Clark Fork River, reared aquatic earthworms (*Lumbriculus*) in them, and fed the *Lumbriculus* (average arsenic concentrations of 129 micrograms per gram [µg/g] dw) to rainbow trout. Fish fed the *Lumbriculus* diet had reduced growth and physiological effects, and the presence of effects was strongly correlated with arsenic but not to other elevated metals in the sediments. Erickson et al. (2010) experimentally mixed arsenic into clean sediments, reared *Lumbriculus* in them, and fed the *Lumbriculus* to rainbow trout. The rainbow trout fed the worms contaminated with arsenic (26 to 77  $\mu$ g/g dw) had reduced growth and disrupted digestion. These studies involved the trivalent form of arsenic.

Kiser et al. (2010) collected bull trout (*Salvelinus confluentus*) from Gold Creek (a mineinfluenced stream in northern Idaho) and observed inflammation, necrosis, and cellular damage of the liver. The authors found fish tissue concentrations were correlated with elevated concentrations of arsenic in the sediments and macroinvertebrates. Damage to livers and gall bladders occurred in lake whitefish (*Coregonus clupeaformis*) fed arsenic contaminated diets as low as 1 mg/kg food dw (Pedlar et al. 2002). Cutthroat trout exhibited reduced growth and liver damage when fed contaminated invertebrates with arsenic concentrations between 14–51 mg/kg dw (Farag et al. 1999). While many studies focused mostly on the effects of arsenic on organs and growth, at least one study has shown that arsenic in zebrafish (*Danio rerio*) diets can reduce reproduction. The single dietary exposure tested for reproductive effects was higher (~135 mg/kg dw) than other dietary toxicity studies with salmonids (Boyle et al. 2008).

Recent attention has been given to assessing whether adverse effects are influenced by the various forms of arsenic that can occur in dietborne exposures, including inorganic arsenic (which has been the focus of most literature to date), monomethylarsonate, dimethylarsinate, and asrsenobetaine (Erickson et al. 2019). The authors documented growth rate reductions on the order of 41% and 18% when fish were fed inorganic arsenic feed containing 134 ug/g and 72 ug/g total arsenic. Of the 70 percent of measured species in the inorganic arsenic feed, 88 percent was inorganic As(III) and 12 percent was inorganic As(V). The authors did not observe growth rate reductions in juvenile rainbow trout fed the dimethylarsinate diet (total arsenic concentrations of 556 ug/g), indicating dimethylarsinate is far less toxic than inorganic As(III). Some reductions in growth rates occurred as a result of monomethylarsonate, though not to the extent as inorganic arsenic (III).

Bioaccumulation of arsenic in prey organisms to concentrations higher than 30 mg/kg dw has been documented in streams where concentrations of arsenic in the water were greater than background ( $\leq$  5 µg/L) but significantly less than the chronic water quality criterion of 150 μg/L dissolved arsenic (NMFS 2014). These studies demonstrate that bioaccumulation of arsenic in invertebrate prey organisms harmful to salmonids can occur in streams with dissolved arsenic concentrations on the order of 10 µg/L or less. However, there has also been documentation of much lower levels of arsenic in invertebrate tissues (0.5 to 2 mg/kg dw) in waters with dissolved arsenic concentrations between 0.5 and 7  $\mu$ g/L. Not surprisingly, a major difference between these streams was the concentration of arsenic in the sediments. Invertebrates from streams with higher arsenic concentrations in sediments had greater tissue concentrations. These studies did not examine arsenic speciation in invertebrate tissues.

Recently, Erikson et al. (2019) concluded that an understanding of the speciation of arsenic in the water and diet is critical for assessing risk to fish, and it is prudent to pay particular attention to inorganic arsenic. As part of their study, the authors examined the variability of arsenic species in macroinvertebrates collected from Panther Creek, Idaho. The macroinvertebrate species examined were a detritivorous stonefly (*Pteronarcys californica*), a filter-feeding caddisfly (*Arctopsyche grandis*) and a predatory stonefly (*Hesperoperla pacifica*). They found that the percentages of total extractable arsenic, inorganic arsenic, and organoarsenical species varied considerably among the species. They also found arsenic concentrations accumulated the most in the detritivore and the least in the predator.

# **Toxicity to Aquatic Invertebrates**

The limited data available suggests that the risk of toxicity to salmonid food organisms is lower than the risk of toxicity to salmonids from eating arsenic exposed organisms. However, we did not locate any studies that tested invertebrates using environmentally relevant exposures through arsenic enriched periphyton or sediments, and conducted through full life exposures or obviously sensitive life stages. Although likely less relevant than exposures through diet, there are some data available regarding arsenic toxicity to aquatic invertebrates from water only exposures.

Irving et al. (2008) exposed mayfly (Ephemeroptera) nymphs to tri- and pentavalent arsenic in water-only exposures for 12 days. For trivalent arsenic, the threshold of growth effects was about 100 μg/L. However, arsenic levels accumulated by the mayfly nymphs in their study (1.2 to 4.6

 $\mu$ g dw) were far lower than those reported from stream locations with lower water concentrations of arsenic. In these stream locations, arsenic was elevated in the diet or sediments, suggesting that the water-only exposures may have underrepresented likely environmental exposures. Results reported in Eisler (1988a) suggest that gammarid amphipods may experience acute toxicity at concentrations of trivalent arsenic that are below the chronic criterion of 150 µg/L. Canivet et al. (2001) similarly found increased mortality of gammarid amphipods and heptageniid mayflies at about 100 μg/L.

# **Summary**

Adverse effects to anadromous fish and aquatic invertebrates could occur from chronic exposure to arsenic when water column concentrations are lower than Idaho's CCC of 150 μg /L. The dietary exposure route is of most concern because bioaccumulation of arsenic in the food chain to levels that could adversely affect fish has been documented in streams with dissolved arsenic concentrations between 5 and 10 µg/L (NMFS 2014). These concentrations are far lower than concentrations found to cause adverse effects from waterborne exposure only (e.g., 100 μg/L for mayflies and about 40 μg/L for rainbow trout embryos).

# **COPPER**

Copper is a naturally occurring trace element. Ambient monitoring studies by the U.S. Geological Survey at 811 sites across the United States have revealed background dissolved copper concentrations ranging from 1 to 51  $\mu$ g/L, with a median of 1.2  $\mu$ g/L (Hecht et al. 2007). In the Salmon River basin, dissolved copper concentrations are generally within the  $< 0.5$  to 4 µg/L range (NMFS 2014). At low concentrations, copper is an essential micronutrient to plants and animals; however, it becomes toxic at slightly higher concentrations. Copper is relatively insoluble in water; however, it becomes more soluble with decreasing pH (Nelson et al. 1991). It can dissolve or bind to organic and inorganic materials in suspension or within the sediments.

Copper toxicity is influenced by chemical speciation, hardness, pH, alkalinity, total and dissolved organic carbon (DOC) in the water, previous exposure and acclimation, fish species and life stage, water temperature, and presence of other metals and organic compounds. Idaho's current copper criteria are based on the biotic ligand model (BLM). This is a computer-based model that predicts copper toxicity based on its expected bioavailability to organisms, which is influenced by various water chemistry parameters (e.g.*,* temperature, pH, dissolved organic carbon, etc.). In its evaluation of the BLM, NMFS identified that the BLM performed relatively well in predicting toxicity; however, it may underpredict acute toxicity in very soft waters. In addition, the BLM performed reasonably well in adequately protecting against sublethal effects such as reduced growth, neurological damage, or behavioral impairments.

# **Toxicity to Fish**

Hecht et al. (2007), NMFS (2014), and EPA (2007) conducted literature reviews regarding copper and its effects on aquatic life. Table B-1 contains information excerpted from those publications, summarizing both acute and chronic effect concentrations reported in the literature. Mortality to fish occurs when insoluble copper-protein compounds form on gill surfaces, causing the sloughing of gill epithelia and eventual suffocation (Nelson et al. 1991). Acute mortality of various life stages of salmon and steelhead exposed to dissolved copper occur at low concentrations in soft waters (hardness between 9 and 42 mg/L), with 96-hour  $LC_{50}$  values ranging from 2.4 to 57 µg/L (Hecht et al. 2007; NMFS 2014).

In addition to direct mortality, adverse effects of copper to salmonids include reduced growth and reproductive impairment. Furthermore, exposure to low levels of dissolved copper can cause other sublethal physiological and behavioral effects such as interference with immune response and reduced disease resistance, reduced swimming stamina, damage to olfactory cellular tissue, and impaired olfactory function. These sublethal effects can lead to death through impairment of the ability of salmonids to avoid predators, find prey, and migrate from and to their natal streams (McIntyre et al. 2012, Puglis et al. 2019, Sommers et al. 2016, Thomas et al. 2016, Calfee et al. 2014, Lorz and McPherson 1976, 1977, Saucier et al. 1991, Saunders and Sprague 1967).

A common chronic effect observed with copper exposure has been reduced growth in laboratory toxicity tests with salmonids. In tests using soft water, copper concentrations of 3.6 ug/L were associated with a 4 to 7.5 percent reduction in the lengths of Chinook salmon and rainbow trout, depending on the statistical model used to analyze the toxicity data. Similarly, exposures to 2.1 µg/L corresponded to length reductions ranging from zero to 4.5 percent. To evaluate the relevance of reduced growth in a laboratory setting to natural-origin populations, Mebane and Arthaud (2010) used population modeling. The authors found that a size reduction of four percent as length was associated with survival reductions ranging from 12 to 38 percent for different migrant groups during their migration downstream to the Lower Granite Dam.

Reproductive impairment (measured as reduced fecundity) was a sensitive endpoint in some chronic tests with copper and fish (Mount 1968; Mount and Stephan 1969; McKim and Benoit 1971; Suter et al. 1987). However, with anadromous steelhead and salmon, presumably longterm exposure of adults to copper in freshwater would be unlikely, since adults are either only passing through migratory areas or are exposed on their spawning grounds for a few weeks or less. Thus, the risk of chronic effects from copper is higher for juvenile fish.

Reduced immune response and disease resistance is an effect of copper that appears to be understudied, considering its potential implications. Stevens (1977) reported that pre-exposure to sublethal levels of copper interfered with the immune response and reduced the disease resistance in yearling coho salmon.

<b>Species</b> (life stage)	<b>Effect</b>	<b>Effect Conc.</b> $(\mu g/L)^b$	<b>Effect Statistic</b>	<b>Exposure</b> <b>Duration</b>	<b>Source</b>
Chinook salmon (fry)	Death	19	$LC_{50}$	96 hr	Chapman 1978a
Coho salmon (fry)	Death	$28 - 38$	$LC_{50}$	96 hr	Lorz and McPherson 1976
Coho salmon (adult)	Death	46	$LC_{50}$	96 hr	Chapman and Stevens 1978
Steelhead/rainbow trout (fry)	Death	$9 - 17$	$\rm LC_{50}$	96 hr	Chapman 1978a, Marr et al. 1999
Steelhead/rainbow trout (fry)	Death at pH 7	2.8	$LC_{50}$	96 hr	Cusimano et al. 1986
Steelhead (adult)	Death	57	$LC_{50}$	96 hr	Chapman and Stevens 1978
Coho salmon (juvenile)	Death	$21 - 22$	<b>NOEC</b>	60 d	Mudge et al. 1993
Steelhead (juvenile)	Death	$24 - 28$	<b>NOEC</b>	60 d	Mudge et al. 1993
Steelhead (egg-to-fry)	Death	11.9	$\mathrm{EC}_{10}$	120 d	Chapman 1982
Rainbow trout (1 days post-hatch [dph]	Death	47.8	$LC_{50}$	96 hr	Ingersoll and Mebane 2014 <sup>d</sup>
Rainbow trout (18 dph)	Death	43.4	$LC_{50}$	96 hr	Ingersoll and Mebane 2014 <sup>d</sup>
Rainbow trout (32 dph)	Death	42.4	$LC_{50}$	96 hr	Ingersoll and Mebane 2014 <sup>d</sup>
Rainbow trout (46 dph)	Death	37.8	$LC_{50}$	96 hr	Ingersoll and Mebane 2014 <sup>d</sup>
Rainbow trout (60 dph)	Death	32.2	$LC_{50}$	96 hr	Ingersoll and Mebane 2014 <sup>d</sup>
Rainbow trout (74 dph)	Death	44.3	$LC_{50}$	96 hr	Ingersoll and Mebane 2014 <sup>d</sup>
Rainbow trout (95 dph)	Death	15.2	$LC_{50}$	96 hr	Ingersoll and Mebane 2014 <sup>d</sup>
Coho salmon (juvenile)	Reduced olfaction and compromised alarm response	$0.18 - 2.1$	$EC_{10}$ to $EC_{50}$	3 <sub>hr</sub>	Sandahl et al. 2007
Chinook salmon (juvenile)	Avoidance in laboratory exposures	0.75	<b>LOEC</b>	$20 \text{ min}$	Hansen et al. 1999
Chinook salmon (juvenile)	Avoidance (soft water)	0.91	$EC_{20}$	$20 \text{ min}$	Hansen et al. 1999
Rainbow trout (juvenile)	Avoidance in laboratory exposures	1.6	<b>LOEC</b>	$20 \text{ min}$	Hansen et al. 1999
Rainbow trout (juvenile)1	Avoidance (soft water)	0.84	$EC_{20}$	30 min	Hansen et al. 1999; Meyer and Adams 2010
Rainbow trout (juvenile)1	Avoidance (soft water)	2.4	$EC_{20}$	96 hr	Morris et al. 2019
Rainbow trout (juvenile)1	Avoidance (soft water)	1.6	$EC_{20}$		Giattina et al. 1982
Chinook salmon (juvenile)	Loss of avoidance ability	2.2	<b>LOEC</b>	21d	Hansen et al. 1999
Atlantic salmon (juvenile)	Avoidance in laboratory exposures	2.4	<b>LOEC</b>	$20 \text{ min}$	Sprague et al. 1965
Atlantic salmon (adult)	Spawning migrations in the wild interrupted	20	<b>LOEC</b>	Indefinite	Sprague et al. 1965
Chinook salmon (adult)	Spawning migrations in the wild apparently interrupted	$10 - 25$	<b>LOEC</b>	Indefinite	Mebane 2000

**Table F-67. Selected examples of adverse effects with copper to salmonids or their prey (Hecht 2007; NMFS 2014; EPA 2007; Mebane 2023) a .**



<sup>a</sup>Abbreviations: LOEC = Lowest observed adverse effect concentration (and most LOEC values given are not thresholds, but were simply the lowest concentration tested); NOEC = No observed adverse effect concentration; LC<sub>50</sub> = the concentration that kills 50 percent of the test population; EC<sub>p</sub> = effective concentration adversely affecting (p) percent of the test population or percent of measured response, e.g., 10 percent for an EC<sub>10</sub>, etc.; and Indefinite = field exposures without defined starting and ending times. NA = not applicable;  $NR = not$  reported;  $d = days$ ;  $hr = hours$ ;  $min = minutes$ ;  $yr = years$ ;  $wk = weeks$ .

bEffects and exposure durations stem from laboratory and field experiments; therefore, in some experiments multiple routes of exposure may be present (i.e., aqueous and dietary) and water chemistry conditions will likely differ (see reference for details).

<sup>c</sup>Acute sensitivity of salmonids to copper varies by life stage, and the swim-up fry stage is probably more sensitive than older juvenile life stages such as parr and smolts or adults.

<sup>d</sup>Toxicity data from the test normalized for EPA's BLM standard water.

 $e^{\theta}$  The EC<sub>20</sub> values were calculated by EPA (2007) using either a probability distribution analysis or a logistic regression analysis.

<sup>f</sup>This study examined ecosystems consisting of a number of species or unidentified species.

Reduction of sensory capabilities is another sublethal impact that can occur when fish are exposed to copper. Dissolved copper is a neurotoxicant that directly damages the sensory capabilities of salmonids at low concentrations (Hecht et al. 2007). These effects can manifest over a period of minutes to hours and can persist for weeks. To estimate toxicological effect thresholds for dissolved copper in surface waters, Hecht et al. (2007) calculated benchmark concentrations (BMCs) for juvenile salmonid olfactory function based on recent, available data. The BMCs ranged from increases of 0.18 to 2.1 μg/L above background copper concentrations (defined as being  $\langle 3 \mu g/L \rangle$  corresponding to reductions in predator avoidance behavior of approximately 8 to 57 percent. These BMCs for juvenile salmonid sensory and behavioral responses fall within the range of copper concentrations (0.75 to 2.5 μg/L) correlated with other low sublethal endpoints (Hecht et al. 2007).

More recently, McIntyre et al. (2012) conducted predation experiments that showed diminished predator avoidance behaviors of juvenile salmon in the presence of elevated copper. A short-term (30-minute) copper exposure made prey easier for predators to detect and capture. Predatory cutthroat trout captured and ate juvenile coho salmon that had been exposed to 4.5 μg/L copper in only about one-third of the time needed to capture and eat coho that had not been exposed to copper. The primary impact of copper on predator-prey dynamics in the McIntyre study (2012) was faster prey detection, manifested as faster time to attack and time to capture. This effect was similar when predators and prey were co-exposed to copper. Morris et al. (2019) examined rainbow trout olfactory inhabitation in low hardness waters and reported a 20 percent reduction in alarm cue response after being exposed to copper concentrations of 2.7 and 2.5 µg/L for 24 or 96 hours, respectively. Further, the fish did not recover (regain their ability to respond to an alarm cue) within 24 hours of exposure in clean water.

Salmonids are known to avoid elevated copper concentrations. Adult Atlantic salmon migrating upstream avoided areas contaminated with a mixture of zinc and copper. The threshold for copper avoidance was about 17 to 21 µg/L (Sprague et al. 1965; Saunders and Sprague 1967). Hansen et al. (1999) found that Chinook salmon avoided copper concentrations of about 0.8 µg /L, and 2.8 to 22.5 µg/L; however, avoidance was not observed in water containing 1.6 µg/L. Rainbow trout were found to avoid concentrations in the range of 1.6 to 88 µg/L. Fish may be more sensitive to (i.e., more readily avoid) lower concentrations of copper if the gradient from clean to contaminated water is sharp rather than gradual (Black and Birge 1980 *in* Atchison et al. 1987).

Many of the studies to date have examined waterborne copper exposures to fish, and less attention has been paid to toxic effects that may manifest through dietary exposures. A few authors (DeForest & Meyer 2015, Erickson et al. 2010, Mebane et al. 2015 and Saiki et al. 1995) have reported reduce growth in fish from dietary exposures to copper; however, fairly high levels of copper were necessary to elicit observable effects (Mebane 2023). Of the studies reviewed, DeForest and Meyer (2015) concluded fish were relatively insensitive to copper exposures through their diet, with the lowest LOEC being 760 ug/g dry weight.

#### **Toxicity to Aquatic Invertebrates**

Copper is highly toxic to many freshwater invertebrates (NMFS 2014). Benthic macroinvertebrate communities that form the food base of salmonids in freshwater streams appear particularly sensitive to copper in both the water column and stream sediments (NMFS 2014). NMFS (2014) concluded that while acute toxicity values ( $LC_{50}$ s) noted by EPA (1985b) were relatively high, compilations of short-term LC<sub>50</sub>s derived from laboratory tests tend to do a poor job of reflecting the sensitivities of invertebrates to metals in field conditions.

Clements et al. (1988) observed a 24 to 36 percent reduction in the number of taxa and a 35 to 52 percent reduction in number of individuals when a stream was contaminated with low doses of copper (15 to 32 µg/L) for 96 hours compared to controls. Beltman et al. (1999) found that copper concentrations in Panther Creek as low as about 10 µg/L resulted in significant changes in community structure, with reductions in stonefly, mayfly, caddisfly and beetle taxa compared to uncontaminated sites. Mebane et al. (2015) found that recovery of *Rhithrogena* sp. in Big Dreek Creek was "apparently prevented" by about  $\lt 5$  ug/L copper, which is far less than the LC50 of 137 ug/L copper for *Rhithrogena hageni* in a 96-hour laboratory toxicity test (Brinkman and Johnston 2008).

Extensive field surveys have been conducted on Panther Creek, near the Blackbird Mine in central Idaho (Beltman et al. 1999, NMFS 2014, Mebane et al. 2015). Prior to the mid-1990s, measured copper concentrations in Panther Creek downstream of Blackbird Creek were always elevated (ranging from 12 to 140 ug/L), and have since declined by more than a factor of 10 as a result of restoration efforts. In 2013, copper concentrations ranged from <0.1 to 2.9 ug/L (Mebane et al. 2015) and were below the BLM-based copper criterion. Farther downstream, below Big Deer Creek, copper concentrations in Panther Creek ranged from 12-97 ug/L in the mid-1990s. By 2004, copper concentrations in Panther Creek at this location were generally below the BLM-based chronic criterion due to restoration actions. Mayflies and stoneflies were effectively eliminated from Panther Creek as a result of the water quality impacts; however, within a few years of restoration actions, stoneflies reappeared and by 2002 stonefly abundance was similar to upstream reference sites. Mayflies took longer to reestablish, becoming more abundant after in 2008.

Sediments with elevated copper that were collected from Chinook salmon and steelhead habitat in Panther Creek, Idaho, and tested in a laboratory setting with clean overlying water caused high mortality to *Hyalella azteca*, a freshwater benthic crustacean (Mebane 2002). The resident benthic invertebrates collected from the same locations as the copper-contaminated sediments had reduced diversity compared to reference collections. Unlike the sediment toxicity tests, adverse effects to the instream invertebrates could not be attributed solely to either copper in the sediments or in water, because copper was elevated in both (Mebane 2002). Elevated copper in sediments is also associated with elevated copper in benthic invertebrate tissues in field studies conducted in metals-contaminated streams (e.g., Ingersoll et al. 1994; Woodward et al. 1994; Beltman et al. 1999; Besser et al. 2001). Uptake and toxicity of copper by invertebrates is strongly influenced by the amount of acid-volatile sulfide in the sediments or by the amount of organic carbon in the sediments (Besser et al. 1995; Mebane 2002).

Elevated copper levels can reduce the availability of preferred invertebrate prey organisms for juvenile salmonids. These reductions have been observed even with relatively low copper concentrations. Reductions or changes in prey availability could translate to adverse effects on juvenile salmonid populations. In the Panther Creek field studies that NMFS (2014) reviewed in some detail, no obvious extinctions of macroinvertebrate effects to salmonids were observed. This suggests either or both that juvenile salmonids are able to switch prey when preferred prey are diminished, or that the food web effects were too subtle to tease out of the natural variability inherent in field monitoring studies using available information.

### **Summary**

As described above, copper is highly toxic to aquatic life. Toxicity to fish includes, but is not limited to, direct mortality, reduced growth, and reduced olfaction function. The estimated 96 hour LC<sub>50</sub> for Chinook salmon is 7.4  $\mu$ g/L in test water at a pH of 7.7 and a hardness of 35 mg/L. Sandahl et al. (2007) reported reduced olfaction function after short-term (i.e., 3 hours) exposures to copper concentrations as low as increases in copper concentrations by as little as 0.18  $\mu$ g/L over background (where background was identified as 3  $\mu$ g/L). Reported EC<sub>10</sub> values for reduced growth in Chinook salmon (Chapman 1982) and rainbow trout (Marr et al. 1996) chronically exposed (i.e., 120 days) to elevated levels of copper were 1.9 µg /L and 2.8 µg /L, respectively. Significant changes in aquatic macroinvertebrate community structure and abundance was observed when copper concentrations were as low as 10 µg /L.

### **CYANIDE**

Cyanide occurs naturally in the environment via production by a variety of plant species. The most likely sources of cyanide in waters are probably forest fires, gold mining operations that use cyanide leaching, and perhaps road salting. Barber et al. (2003) examined releases of cyanides from biomass burning and their effect on surface runoff water. In laboratory test burns, available cyanide concentrations in leachate from residual ash were much higher than in leachate from partially burned and unburned fuel and were similar to or higher than a 96-hour median  $LC_{50}$ value of 45 µg/L for rainbow trout. Free cyanide concentrations in stormwater runoff collected after a wildfire in North Carolina averaged 49 µg/L, which is again similar to the rainbow trout LC<sup>50</sup> and an order of magnitude higher than in samples from an adjacent unburned area (Barber et al. 2003). Eisler (1991) reported average background levels of cyanide in freshwater systems to be  $0.9 \mu g/L$ .

Cyanide is toxic to most living organisms and primarily occurs in aquatic environments as free cyanide (i.e., hydrocyanide and cyanide ion), simple cyanide salts, metal-cyanide complexes, and in some organic compounds. The most bioavailable and toxic forms are the free cyanide (Gensemer et al. 2007). The Idaho criteria for cyanide are expressed as weak acid dissociable (WAD) cyanide, which includes free cyanide as well as metal-cyanide complexes that readily dissociate under weakly acidic conditions (pH 5–6). Analyzing samples for WAD cyanide produces a higher value than analyzing for free cyanide (NMFS 2014).

# **Toxicity to Fish**

Free cyanide is extremely toxic and fast acting. The mechanism of cyanide toxicity involves inhibiting cytochrome oxidase, the terminal oxidative enzyme of the mitochondrial electron transport chain, thus blocking aerobic adenosine triphosphate synthesis. The result of this mechanism of toxicity is that cyanide is a rapid and potent asphyxiant (Eisler 1991). Its toxicity is strongly influenced by temperature in that toxicity increases with decreasing temperature. Kovacs and Leduc (1982a), reported 96-hour rainbow trout  $LC_{50}$  values of 27, 40, and 65 µg/L for exposures in waters with temperatures of 42.8, 53.6, and 64.4 $\rm{°F}$  (6, 12, and 18 $\rm{°C}$ ), respectively. The authors Kovacs and Leduc (1982b) also examined sublethal effects associated with chronic exposure (i.e., 20-day exposure) to elevated cyanide concentrations at varying temperatures. The authors observed chronic toxicity effects on growth in terms of average fat gain and dry weight when juvenile rainbow trout were exposed to 5  $\mu$ g/L around 6.1°C. At about 12.8°C, toxicity effects were evident at concentrations  $\geq$ 10 µg/L. Significant growth reductions were observed when juvenile rainbow trout were exposed to 15  $\mu$ g/L at all temperatures tested. No statistically significant difference between controls and test specimens were recorded at exposures to concentrations less than 4.8  $\mu$ g/L in water temperatures of 6°C. The swimming abilities of juvenile rainbow trout were also found to be reduced at all cyanide concentrations tested in the range of 5 to 45 µg/L, with the effect increasing at lower temperatures.

Bioconcentration of cyanide is considered to be negligible in fish because the compound is easily metabolized. As reported by EPA (1985c) the existing literature does not provide evidence for cyanide biomagnification. This is likely due to the fact that vertebrate species, such as fish, may readily metabolize cyanide, thus removing the cyanide from the food chain at that level. Accumulation of metallocyanide complexes in sediment is not likely because dissociation occurs easily at pH values lower than 8.

# **Toxicity to Aquatic Macroinvertebrates**

Available toxicity data for the types of aquatic insects and crustaceans that juvenile salmonids feed on indicate that they are similarly or less sensitive to cyanide compared with listed salmon and steelhead (EPA 1985c; Eisler 1991). As an example, Call and Brook (1982; as cited in ECOTOX) reported a 96-hour LC<sub>50</sub> of 436 µg/L for the American salmonfly (*Pteronarcys dorsata*).

# **Summary**

Short-term  $LC_{50}$  values for rainbow trout were reported as low as 27  $\mu$ g/L for exposures in water temperatures of around 4.4°C. Sublethal effects (e.g., reduced growth and reduced swimming abilities) were documented when juvenile rainbow trout were exposed to cyanide concentrations as low as 5 µg/L for 20 days. Based on available information, aquatic invertebrates do not appear to be adversely affected by concentrations that are protective of fish.

#### **MERCURY**

Mercury is a naturally occurring heavy metal that is found in trace amounts in air, water, and soil. Mercury can enter the aquatic environment from aerial deposition, surface runoff and spills, and via contaminated groundwater. In surface water, mercury can be present in both inorganic and organic forms, with organic mercury (more commonly referred to as methylmercury) being the more highly toxic form. Mercury speciation is influenced by physical, chemical, and biological reactions. Elemental and inorganic mercury complexes are not bioavailable; however reactive forms of inorganic mercury and methylmercury are bioavailable.

### **Bioavailability**

Mercury transformation in the environment influences how it moves through a watershed, and the biogeochemical cycling of mercury is complex and not fully understood (Bravo and Cosio 2020; Rodrigues et al. 2019). Inorganic mercury is often the dominant form in aquatic environments. In the water column, inorganic mercury can be reduced to elemental mercury and reemitted back into the atmosphere, methylated to the organic form of mercury, or bind to organic matter as well as inorganic particles and deposit to stream sediments. Mercury has a high affinity to sorb to sediments as well as suspended dissolved and particulate matter. Thus, hydraulic transport of mercury in streams is controlled by dissolved organic carbon and suspended particulate matter. Additionally, stream sediments may serve as a source and a sink for mercury, facilitating sequestration and reduction through burial in the aquatic ecosystem (Ullrich et al. 2001, Branfireun et al. 2020). The risk of bioaccumulation in an aquatic ecosystem depends on whether mercury transformation processes favor forms of mercury that are bioavailable for methylation (Bravo 2020).

Rates of bioaccumulation of methylmercury in the food web are thought to be affected by water temperature, pH, water hardness, organism age, organic carbon availability, dissolved oxygen, number of trophic levels, concentration of sulfates, and in the presence of zinc, cadmium, or selenium in solution (Porcella et al. 1995). Methylmercury levels in stream biota are determined by the supply of methylmercury to the base of the foodweb (Wentz 2014), which is determined by the amount of bioavailable inorganic mercury. While it was once believed that only sulfate reducing bacteria could methylate bioavailable inorganic mercury in anoxic conditions, it has been recently discovered that a variety of microorganisms carrying specific genes can methylate inorganic mercury. Furthermore, methylation of mercury can occur in oxygen poor aquatic environments such as lakes, ponds, wetlands, sediments, and flooded soils (Eckley et al. 2005, Jonsson et al. 2012; Windham-Myers et al. 2014, Bravo et al. 2020).

Methylation of bioavailable, inorganic mercury is primarily accomplished by bacteria (e.g., sulfate reducing bacteria, iron reducing bacteria, and methanogens), and methylation is greatest in environments with low pH, low dissolved oxygen, and high organic matter (Peterson et al. 2023; Eckley 2021a). Abundant sources of organic material as well as terminal electron accepting compounds such as sulfate or ferric iron can increase the activity of methylating bacteria. Contributing to the complexity of mercury cycling is the fact that methylmercury can break down in the environment through decomposition (demethylation) via abiotic processes involving chemical and photo-chemical reactions as well as microbial processes (Eckley et al.

2021b; Black et al. 2012; Kim and Zoh 2012). Demethylation yields methane and inorganic mercury species that will continue to cycle in the environment (EPA 2024).

Total mercury and methylmercury are not always strongly correlated and there can be substantial variation. For example, in Idaho waters, the percentage of methymercury ranged from 0.2 to 58 (Essig 2010; Poulin et al. 2020). In the project area, Holloway et al. (2017) reported methylmercury concentrations were up to two percent of mercury concentrations in samples from Sugar Creek and the EFSFSR. Eckley et al. (2021a) reported decreasing total mercury concentrations in Sugar Creek with increasing distance from the Cinnabar mine site, whereas methylmercury concentrations increased. The authors attributed the downstream increase in methylmercury to the more favorable lower gradient, forested ecosystem. Both organic and inorganic mercury bioaccumulate; however, methylmercury accumulates at greater rates because it is more efficiently absorbed and preferentially retained (Scheuhammer 1987; Wiener 1995). Methylmercury is biomagnified between trophic levels in aquatic systems and in general proportion to its supply in water (Wattras and Bloom 1992). Accumulated mercury in fish tissue is almost entirely methylmercury (Bloom 1992; Hammerschmidt et al. 1999; Harris et al. 2003). As such, the toxicity of methylmercury is particularly important with respect to effects to higher trophic level fish (Sorensen 1991; Nichols et al. 1999).

Baseline concentrations of total inorganic mercury in Idaho waters ranged from <0.2 to 6.8 nanograms/liter (ng/L) (NMFS 2014). In April 2024, EPA proposed to promulgate revised mercury aquatic life criteria for Idaho. The proposed criteria include total mercury concentrations in fish tissue and the water column. More specifically, EPA has recommended a water column mercury criterion of 2.1 ng/L, expressed as a 30-day average concentration that is not to be exceeded more than once every three years.

# **Toxicity to Fish**

Mercury is a potent neurotoxin that causes neurological damage, which in turn leads to behavioral effects (e.g., loss of coordination and reduced swimming activity) that can lead to reduced growth and reproduction (Berntssen et ak, 2003, Wiener et al. 2003, Weis 2009; Sandheinrich and Wiener 2010; Kidd and Batchelar 2011). Methylmercury readily penetrates the blood brain barrier, produces brain lesions, spinal cord degeneration, and central nervous system dysfunctions.

Most available data suggest that salmonid species are not susceptible to acute toxicity from direct exposure to mercury in water at concentrations approaching 2.1  $\mu$ g/L (Kidd and Batchelor 2011). The EPA (1985d) reported  $LC_{50}$  values for salmonids exposed to inorganic mercury that ranged between 155  $\mu$ g/L and 420  $\mu$ g/L. For organic mercury, reported LC<sub>50</sub>s ranged from 5  $\mu$ g/L to 84 µg/L, depending on the chemical form, with a phenylmercuric compound being the most toxic (EPA 1985c). Buhl and Hamilton (1991) exposed coho salmon and rainbow trout alevins and parr to mercuric chloride, and determined average  $LC_{50}$ s ranging between 193  $\mu$ g/L and 292  $\mu$ g/L. Devlin and Mottet (1992) determined a methylmercury LC<sub>50</sub> equal to 54  $\mu$ g/L for coho salmon embryos exposed for 48 days. Niimi and Kissoon (1994) exposed rainbow trout subadults to 64 μg/L of mercuric chloride until the fish died. The average time to death was 58 days. All of these concentrations are orders of magnitude higher than a water column concentration

that was previously deemed protective of anadromous salmonids  $(0.002 \mu g/L)$ . In another experiment, the authors determined rainbow trout lived more than 100 days when exposed to 4 μg/L of methylmercury chloride. The lowest effect level noted from an "acute" type study was an LC<sub>10</sub> of 0.9 μg/L following a 28-day exposure of rainbow trout embryos to mercury, with a no-effect (LC<sub>1</sub>) estimate of 0.2 μg/L (Birge et al. 1980).

Effects due to chronic exposures to low concentrations of mercury in the water column are generally manifested through dietary exposures. Methylmercury both bioconcentrates and biomagnifies across trophic levels, and corresponding, field-measured BAFs can be in the millions for top trophic level fish (Nichols et al. 1999) and have been reported to be in the hundreds of thousands for rainbow trout (EPA 2024). Because methylmercury can readily bioaccumulate and biomagnify, food chain transfer is by far the most important mercury exposure pathway in aquatic ecosystems (Hall et al. 1997, Wiener et al. 2003, NMFS 2014, EPA 2024). To a lesser extent, fish obtain mercury from water passed over the gills, and fish also methylate inorganic mercury in their gut (Wiener and Spry 1996).

Methylmercury concentrations are generally greatest in higher trophic levels (e.g., piscivorous fish). However, exceptions to this general rule of thumb have been documented. MacRury et al. (2002) documented higher mercury burdens in fish that preyed on benthic invertebrates compared to fish that were strictly piscivorous. Juvenile Chinook salmon and steelhead in freshwater ecosystems feed predominantly on macroinvertebrates and would generally be at lesser risk of methylmercury accumulation, particularly if the recommended water quality criteria are adopted and implemented.

A number of researchers have attempted to quantify fish tissue burdens associated with lethal and sublethal effects (Table B-2). Fish tissue concentrations associated with acute mortality are quite high, ranging from 6 to 20 mg/kg ww in muscle (Sandheinrich and Wiener 2010; Wiener and Spry 1996). Devlin and Mottet (1992) exposed coho salmon eggs (after hardening) to methylmercury concentrations ranging from 6 to 139  $\mu$ g/L for 50 days. Calculated prehatch LC<sub>50</sub> values ranged from 54.1 to 70.8  $\mu$ g/L. Some mortality was observed in the 13  $\mu$ g/L treatment, and it is unclear whether any observed mortality occurred in the 6 µg/L treatment. The authors reported embryo mercury concentrations up to 20 times greater than water concentrations, illustrating that incubating embryos can rapidly uptake mercury from their environment. We did not find any studies examining embryo mortality from environmentally relevant exposures concentrations. (i.e., concentrations similar to those predicted to occur under the proposed action). Studies relating fish tissue concentrations to other apical endpoints such as reduced growth or reduced reproduction are limited. Available information suggests that sublethal effects are far more sensitive (Table F-2).

<b>Species</b> (life stage)	<b>Exposure</b> <b>Duration</b>	<b>Dietary Conc.</b> (mg/kg dw)	<b>Tissue</b> Conc. (mg/kg) ww)	<b>Tissue</b>	<b>Observed Effects</b>	<b>Source</b>
Rainbow trout (fingerling)	84 days	23.9	10	Whole Body	Reduced growth; Survival not affected	Rodgers and Beamish 1982
Rainbow trout (fingerling)	84 days	46.9	20	Whole Body	Reduced growth; Survival not affected	Rodgers and Beamish 1982
Rainbow trout (fingerling)	84 days	94.8	30	Whole Body	Reduced growth; Survival not affected	Rodgers and Beamish 1982
Atlantic salmon (parr)	120 days	5	$0.61(0.4 -$ 0.8)	Muscle <sup>1</sup>	Increased antioxidant enzyme activity Severe vacuolation in brain tissue	Berntssen et al. 2003
Atlantic salmon (parr)	120 days	10			Decreased antioxidant enzyme activity Severe vacuolation in brain tissues; and diffuse necrosis Decreased neural enzyme activity Decreased feeding behavior	Berntssen et al. 2003
Fathead minnow (adult)	250 days	0.87	$0.86$ male 0.92 female	Whole body (minus gonads)	Decreased number of spawning fish; Decreased sex hormones Increased time to first spawn Decreased female gonadal somatic index Increased ovarian follicular apoptosis	Drevnick and Sandheinrich (2003); Drevnick et al. 2006
Fathead minnow	195 days	0.88	$0.39$ male <sup>2</sup> $0.52$ female <sup>2</sup>	Whole body (minus gonads)	Decreased number of fish spawning Increased time required to first spawn Decreased egg production Decreased female gonadosomatic index	Hammerschmidt et al. 2002 <sup>2</sup>
Fathead minnow (adult)	$>250$ days	0.87	0.71	Whole body	Decreased reproductive behavior	Sandheinrich and Miller 2006
Walleye (1 year)	180 days	1 ww	2.37	Carcass (minus viscera)	Decreased growth Increased gonadal atrophy Decreased gonadal somatic index	Friedmann et al. 1996
Walleye (1 year)	180 days	$0.1$ ww	0.25	Carcass (minus viscera)	Decreased plasma cortisol (impaired immune function) Increased gonadal atrophy	Friedmann et al. 1996

**Table 68. Selected examples of adverse effects of mercury to salmonids or their prey (NMFS 2014; EPA 2024).**

Notes:

1 NMFS (2014) estimated muscle tissue concentrations from the reported brain tissue residues using organ tissue ratios derived from McKim et al. 1976

2. Dry weight converted to wet weight using 80% moisture.

Histological changes in the spleen, kidney, liver, and gonads have been reported for multiple species of freshwater fish at tissue concentrations of methylmercury well below 1.0 mg/kg ww (Sandheinrich and Wiener 2010). Sandheinrich and Wiener (2010) concluded that effects on biochemical processes, damage to cells and tissues, and reduced reproduction in fish have been documented at methylmercury concentrations of about 0.3 to 0.7 mg/kg ww (whole body) and 0.5 to 1.2 mg/kg ww (muscle tissue). In addition to organ tissue damage, exposure to sublethal concentrations of MeHg can cause neurotoxic effects resulting in impairment of the ability of fish to locate, capture, and ingest prey and to avoid predators (Wiener et al. 2003). In Atlantic salmon parr, brain tissue concentrations of 0.69 mg/kg ww were associated with brain lesions and behavior alterations (Berntssen et al. 2003). No reductions in growth or survival were noted. Using organ tissue ratios derived from McKim et al. 1976, NMFS (2014) extrapolated brain residues to muscle tissue concentrations of 0.61 mg/kg wet weight (range of  $0.4 - 0.8$  mg/kg ww).

Fjeld et al. (1998) showed that the feeding efficiency and competitive ability of grayling (*Thymallus thymallus*) exposed as eggs to waterborne methylmercury chloride at concentrations of  $> 0.8 \mu$ g/L for 10 days and having yolk-fry with mercury concentrations of 0.27 mg/kg ww or greater, were impaired when fish were tested 3 years later. Based on studies by McKim et al. (1976), NMFS (2014) translated the NOEC (0.09 mg/kg ww in yolk-fry, associated with a 0.16 µg/L exposure) from Fjeld et al. (1998) to a maternal mother tissue concentration of 0.7 (range 0.15 to 1) mg/kg ww. Similar extrapolations of the 0.09 embryo NOEC by the U.S. Fish and Wildlife Service (2003) and Idaho Department of Environmental Quality (2005) using other data yielded muscle tissue concentrations ranging from 0.45 to 1.8 mg/kg ww. Bridges et al (2016) found that females fed diets of 5 mg/kg dw MeHg over a period of 30 days transferred MeHg to eggs. Survival of embryos was significantly reduced due to maternal transfer of MeHg, and a large number of the surviving offspring had spinal deformities and exhibited circular swimming patterns. The authors also noted earlier hatch times for embryos whose mothers were fed a 0.87 mg/kg dry weight MeHg diet, though there was no statistically significant reduction in survival.

NMFS (2014) found the most sensitive effects of long-term exposures of a variety of fish species to methylmercury to generally have been reproductive or behavioral effects, with concentrations greater than about 0.3 mg/kg ww in whole bodies or axial muscle tissues likely to be harmful to fish. However, NMFS noted that adverse effects (e.g., steroidogenesis, and changes in metabolic, endocrine, and immune-related genes) occurred at concentrations lower than this. Whether these sub-clinical effects would translate into harmful organism-level effects (e.g., reduced growth or reproduction) is unknown. Bevkar et al. (2005) recommended a 0.2 mg/kg ww (whole body) threshold as being protective of fish from sublethal effects. Most recently, EPA recommends a whole-body fish tissue concentration of 0.162 mg/kg ww whole body tissue, applied to adult trophic level 4 fish.

*Evaluation of Water Column Threshold Concentrations.* NMFS (2014) evaluated a variety of matched water and tissue mercury samples to assess concentrations in water that might result in mercury accumulation to harmful levels. Matched samples repeatedly show that mercury concentrations in fish commonly approach or exceed the lowest adverse effect threshold of about 0.2 to 0.3 mg/kg ww, even when mercury concentrations in the water were commonly an order of magnitude lower than 12 ng/L. NMFS then used two approaches to calculate potential water

column concentrations of mercury that would present a low risk of bioaccumulation of mercury to harmful levels in fish. The first approach involved linear regression between paired water column and fish tissue data; the second approach used BAFs to estimate potential mercury tissue residues from mercury concentrations measured in the field. Using data reported by Essig (2010), a linear regression ( $r^2 = 0.22$ ; p < 0.00001) suggested that a water concentration of 0.9 ng/L would, on the average, result in a fish tissue concentration of about 0.3 mg/kg. Using data from DeForest et al. (2007) and Essig (2010), the BAF approach suggested that if total mercury concentrations in the water were less than 2 ng/L, then fish tissue concentrations would be expected to be less than 0.3 mg/kg ww.

*Evaluation of Dietary Concentrations.* DePew et al. (2012) used data from twenty experimental studies to derive thresholds for dietary methylmercury concentrations. The authors concluded that adverse effects on behavior usually occurred when dietary concentrations exceeded 0.5 mg/kg ww (2 mg/kg dw, assuming a 75 percent water content), and adverse effects on reproduction occurred when dietary concentrations were at 0.2 mg/kg ww (0.8 mg/kg dw, assuming a 75 percent water content). Because these thresholds were derived from laboratory experiments where fish are held in ideal conditions (temperature, food availability, etc.), the thresholds may still underestimate adverse effects. This is because fish in the wild experience additional stressors related to foraging, predation, temperature fluctuation, and other contaminants that are not present in laboratory settings.

Bernssten et al 2003 fed Atlantic salmon parr diets with 4.35 mg/kg dw and 8.48 mg/kg dw MeHg and 10 mg/kg and 100 mg/kg inorganic mercury over a period of 4 months. No reduced growth or mortality was observed. While these levels did not result in outright mortality, the researchers observed increased brain lesions and behavior alterations (reduced swimming activity) at 8.48 mg/kg dw MeHg diets and increase oxidative stress at 4.35 mg/kg dw MeHg diets.

# **Toxicity to Aquatic Invertebrates**

Nagpal (1989) summarized both acute and chronic toxicity data for macroinvertebrates. Acute toxicity for invertebrates was found to be dependent upon species, developmental stage, and overall environmental conditions. For inorganic mercury, *Daphnia sp*. were found to be the most sensitive invertebrate, with LC50s of 1.4 to 4.4 μg/L for *D. magna* and 2.2 μg/L for *D. pulex*. In chronic toxicity tests of inorganic mercury and *D. magna,* adverse effects occurred at concentrations ranging from 0.72 to 1.82 μg/L. In chronic tests for methylmercury, adverse effects were observed in *D. magna* at concentrations less than 0.04 μg/L. In a field study, Kraus et al. (2022) documented reduced insect diversity in a stream with elevated levels of mercury from historic mining. The authors found that total mercury in the biota up to 4.7 miles downstream were up to 11 times greater than background levels and MeHg propagated through the downstream food web. Invertebrate communities sampled approximately 2.5 miles downstream of the mine site were substantially less diverse than reference conditions. The farthest downstream site did not have an apparent reduction in benthic invertebrate diversity; however, densities of mayfly and stonefly taxa were depressed.

Based on reviewed literature, toxicity of mercury to aquatic invertebrates and subsequent loss of abundance or taxonomic shifts is not the most important pathway of effect to ESA-listed species. Instead, bioaccumulation of mercury in invertebrate tissue is of greater concern due to the propensity of mercury to biomagnify.

# **Summary**

Bioaccumulation of mercury through the food chain is the most important pathway of toxicity for fish. There is relatively little information from which to derive risk evaluation thresholds for anadromous species. Relying on previous work, the best available information suggests fish tissue concentrations of less than 0.2 mg/kg of total mercury should be adequately reduce the risk of sublethal effects in anadromous salmonids. While site specific factors will influence the extent to which mercury becomes bioavailable and subsequently bioaccumulates, in general, water column concentrations less than 2 ng/L are expected to have a low risk of contributing to sublethal effects in anadromous salmonids.

# **TOTAL DISSOLVED SOLIDS**

Total dissolved solids (TDS) is an integrated measure of all the ions in water that can pass through a filter. Common freshwater ions include calcium, magnesium, bicarbonate, chlorides, nitrate, phosphorus, iron, and sulfate. The concentration of TDS affects the water balance in aquatic organism cells. In low TDS environments, water will move into the cell, causing it to swell. In high TDS environments, organisms may shrink as water moves out of their cells. The ionic composition of TDS affects its toxicity and is not predictable from TDS concentrations (Chapman et al. 2000; Weber-Scannell and Duffy 2007). Mining activities can increase major ion concentrations in streams by enhancing weathering of geological material (Brent et al. 2022; Pond et al. 2014).

TDS toxicity is exerted through changes in salinity, changes in the ionic composition of the water, and individual ions. In addition, TDS can influence the toxicity of other contaminants, such as metals. Because the toxicity of TDS depends on the specific ion composition comprising the measure, ecological thresholds to protect aquatic life are not readily available.

# **Toxicity to Fish**

Toxicity of TDS to fish is dependent upon: (1) the ion composition; (2) concentrations of individual ions; and (3) species and life stage exposed (Weber-Scannell et al. 2007, Stekoll et al. 2003, Stekoll et al. 2009). As a result, it is not surprising that toxicity results vary in the literature. Chapman et al. (2000) did not observe any mortality or reduced growth of rainbow trout fry exposed to up to  $\sim$ 2,000 mg/L TDS for 7 days. Similarly, no adverse effects were observed for rainbow trout eggs fertilized and incubated in up to 2,000 mg/L TDS solutions.

When beginning exposure to elevated TDS during fertilization, Stekoll et al. (2009) found a dose-response relationship associated with continuous TDS exposure (fertilization through early development). Reported  $EC_{20}$  and LOEC values for steelhead were 1,100 and 750 mg/L, respectively. The reported  $EC_{20}$  and LOEC values for Chinook salmon were 85 and 250 mg/L,

respectively (Stekoll et al. 2009). The LOECs were substantially higher when TDS exposure occurred after fertilization. The authors reported a LOEC of 1,875 mg/L for steelhead, and no LOEC was reported for Chinook since no mortality difference was observed between the test concentrations and controls. The authors also investigated which ion was likely responsible for the toxicity and concluded that calcium was the most likely culprit. In their specific ion tests, the fertilization process was most sensitive to calcium, and calcium was the most prevalent ion in the TDS mixture. The calcium ion has also been shown to be responsible for toxicity by a Ketola et al. 1988 and Brannock et al. 2002 (as cited in Weber-Scannell and Duffy 2007). Other authors (Mount et al. 1997 and Peterson et al. 1988) have found potassium as being the source of toxicity in their experiments, providing further evidence for the importance of site-specific considerations.

The majority of available toxicity studies show relatively high concentrations of TDS  $(>1,000)$ mg/L) are needed to elicit effects. However, few studies examined exposures at fertilization, which based on available information appears to be the most sensitive life stage (Stekoll et al. 2009).

# **Toxicity to Invertebrates**

Similar to fish, invertebrate toxicity is dependent upon the concentration of TDS, ionic composition, and species being tested. A wide range of toxicity has been reported for aquatic invertebrates, with the vast majority of studies reviewed by Scannell et al. (2007) reporting  $LC_{50}$ or EC<sub>50</sub> values greater than 1,000 mg/L (Weber-Scannell et al. 2007). Brent et al. 2022 reported LC<sup>50</sup> values >3,000 mg/L for *C. dubia* and *H. azteca*. Chapman et al. (2000) documented reduced growth and survival of chironomid larvae exposed to 2,000 mg/L TDS. No observed effects were observed in test solutions of up to 1,200 mg/L TDS.

A number of authors have found macroinvertebrate assemblages to be influenced by TDS (Olson and Hawkins 2017, Pond et al. 2008, Pond et al. 2014, Lind 2018, and Cormier et al. 2018). These authors noted loss of macroinvertebrate genera with increasing TDS. Cormier et al. 2018 developed a specific conductance (which is closely correlated with TDS) chronic benchmark using a field-based methodology for the Central Appalachians ecoregion. They proposed a specific conductance benchmark < 301 microseimens per centimeter as being protective of 95 percent of the macroinvertebrate assemblage documented in that ecoregion.

Mayfly genera appear to be among the most sensitive species to TDS. Kennedy et al. (2004) go so far as to suggest the standard test species, such as *D. magna* or *C. dubia* were not "ecologically relevant gauges" of toxicity TDS. They found the mayfly *Isonychia bicolor* was at least 2.5 times more sensitive to specific conductance than *C. dubia*. Similarly, Brent et al. (2022) suggest site-specific indigenous macroinvertebrate species be used in toxicity tests to more accurately evaluate toxicity risks from TDS exposures. NMFS did not find any studies of toxicity in low hardness and low TDS waters; however, once study, Cormier et al. (2018), suggested that macroinvertebrate assemblages are less vulnerable to low specific conductance (and by extension TDS).

#### **Summary**

Available science indicates that TDS toxicity is heavily dependent upon the ion composition and concentrations of ions. Furthermore, TDS tolerance varies widely among genera. The most sensitive fish life stage appears to be fertilization through early emergence, with the lowest  $EC_{20}$ value being 85 mg/L for Chinook salmon. Benthic community richness metrics decrease with increasing TDS concentrations, with mayflies being among the most sensitive genera. Given that juvenile salmonids are opportunistic feeders, as long as a diverse group of macroinvertebrates are protected, some loss of prey items would not be expected to reduce individual fitness.

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#### **APPENDIX G**

## **FLOW RELATED EFFECTS ON SNAKE RIVER SPRING/SUMMER CHINOOK SALMON AND SNAKE RIVER BASIN STEELHEAD**

#### **Calculation of The Flow Rate (cfs) Effects of the Proposed Action**

The biological assessment (BA) (Stantec 2024) included estimates of the net effects of the proposed action on stream flow (flow effects). The effects are expressed as the percent change in flow compared to "baseline." The flow effects for mine years (MYs) -2 through Post-Closure, for the three affected reaches of the East Fork South Fork Salmon River (EFSFSR) and for the affected reach of Meadow Creek, are in the BA Table 4.1-35 on pages 379-398. Table 4.1-35 included "baseline" flows, which enabled a mathematical conversion from percent change from baseline, to a flow rate (i.e., cubic feet per second [cfs]). For Sugar Creek, page 167 of the BA, flow effects are described as an approximate 3% reduction from baseline during the mine operation period and for approximately 50 years post-closure. However, the BA Table 3.5-33 describes flow effects in Sugar Creek as 1% for May and June, 2% for April, and 3% for all other months. We applied these estimated effects to average flows measured in Sugar Creek at U.S. Geological Survey (USGS) gage 13311450, and the result was 0.42 cfs. We also estimated the effects on flow in Sugar Creek by estimating the amount that flow would be reduced by removing 185 acres (i.e., the catchment area of the West End Pit Lake) from the drainage, and the result was 0.39 cfs. Because the mechanism for flow reduction in Sugar Creek will be the effective reduction in drainage area, we presumed that the proposed action would reduce flow in Sugar Creek by approximately 0.39 cfs from MY 1 through 50 years post closure. We also presumed that the effect of the proposed action on flow in downstream reaches (i.e., downstream from Sugar Creek) will be the sum of the effect on flow in Sugar Creek and the effect on flow in the EFSFSR reach upstream from Sugar Creek (i.e., between the Yellow Pine Pit [YPP] and Sugar Creek). The effects of the proposed action on flow expressed as cfs, calculated as described above, are in Table G-1.

These estimated effects are less than the maximum allowable diversion rates stipulated in the water rights, suggesting that flow effects could be greater. We assumed that the estimates in the BA are based on project water needs and discharge of treated water, and that the information in Table 4.1-35 accurately describes effects on mean monthly flow. Under this assumption, the effects could be greater on a temporary basis, but the effects on mean monthly flow would be as described in Table G-1. We therefore used the flow effects described in Table G-1 to estimate the flow related effects of the proposed action on Snake River (SR) spring/summer Chinook salmon and SR Basin steelhead.

<b>Mine</b>		<b>Net Flow Reduction (Cfs) Due to Water</b> <b>Diversion and Discharge</b>		Flow <b>Reduction</b> (Cfs) Due to <b>Drainage Area</b> <b>Reduction</b>	<b>Total Net Flow</b> <b>Reduction in the</b> <b>EFSFSR</b> <b>Downstream from</b>		
Year	<b>Meadow</b> <b>Creek</b>	<b>EFSFSR</b> <b>Above</b> <b>Meadow</b> <b>Creek</b>	<b>EFSFSR</b> <b>Between</b> <b>Meadow</b> <b>Creek</b> and the <b>YPP</b>	<b>EFSFSR</b> <b>Between</b> the YPP And <b>Sugar</b> <b>Creek</b>	<b>Sugar Creek</b>	<b>Sugar Creek and</b> the SFSR <b>Downstream from</b> the EFSFSR	
$-2$	$-0.20$	0.00	$-0.20$	$-0.21$	$\mathbf{0}$	$-0.21$	
$-1$	0.20	0.00	0.20	0.82	$\mathbf{0}$	0.82	
$\mathbf{1}$	0.42	0.00	0.44	1.50	0.39	1.90	
$\overline{2}$	0.77	0.00	0.76	3.42	0.39	3.82	
$\overline{3}$	0.84	0.00	0.86	2.40	0.39	2.80	
$\overline{4}$	1.40	0.00	1.35	2.61	0.39	3.01	
5	$-0.21$	0.01	$-0.15$	0.97	0.39	1.37	
6	1.12	0.02	1.34	2.48	0.39	2.88	
$\overline{7}$	2.23	0.02	2.38	3.69	0.39	4.09	
8	1.75	0.01	1.70	3.54	0.39	3.94	
9	0.55	0.00	0.56	2.10	0.39	2.50	
10	0.55	0.00	0.54	2.20	0.39	2.60	
11	0.54	0.00	0.56	2.40	0.39	2.80	
12	0.62	0.00	0.63	1.97	0.39	2.37	
13	0.67	0.08	0.75	2.14	0.39	2.54	
14	0.47	0.14	0.61	1.45	0.39	1.85	
15	0.53	0.10	0.48	1.02	0.39	1.42	
16	0.20	0.05	0.13	0.45	0.39	0.85	
17	0.21	0.14	0.30	0.54	0.39	0.94	
18	0.26	0.11	0.34	0.57	0.39	0.97	
19	0.18	0.08	0.22	0.34	0.39	0.74	
20	0.11	0.07	0.15	0.33	0.39	0.73	
$20+$	0.00	0.00	0.00	0.00	0.24	0.24	

**Table G-11. Flow reduction (cfs) in the affected reaches of the EFSFSR, Meadow Creek, Sugar Creek, and the SFSR.**

# **1.1 Flow Data for the Action Area**

Flow in the affected stream reaches was characterized as described in the following bullets:

- EFSFSR above Meadow Creek Data from the 13310800 EFSFSR above Meadow Creek gage.
- EFSFSR between Meadow Creek and the YPP Data from the 13311000 EFSFSR at Stibnite, Idaho gage.
- Meadow Creek We subtracted flows measured in the EFSFSR above Meadow Creek from those measured in the EFSFSR between Meadow Creek and the YPP.
- EFSFSR between the YPP and Sugar Creek Data from the 13311250 EFSFSR above Sugar Creek hear Stibnite, Idaho gage.
- Sugar Creek Data from the 13311450 Sugar Creek near Stibnite, Idaho gage.
- EFSFSR between Sugar Creek and Johnson Creek We added flows measured in Sugar Creek and the EFSFSR above Sugar Creek (gages 13311450 and 13311250).
- EFSFSR between Johnson Creek and the South Fork Salmon River (SFSR) We added flows measured in Sugar Creek, the EFSFSR above Sugar Creek, and Johnson Creek (gages 13311450, 13311250, and 13313000).
- SFSR between the EFSFSR and the Salmon River We added flows measured in Sugar Creek, the EFSFSR above Sugar Creek, Johnson Creek, and the SFSR near the Krassel Ranger Station (gages 13311450, 13311250, 13313000, and 13310700).

We used average July-September flow for all analyses and we normalized flows by dividing the average July-September flow, for all years, by the average July-September flow during the period of record, and expressed the result as a percentage. We calculated the percent reduction in July-September flow, due to the proposed action, by dividing the estimated reduction in flow, expressed as cfs (Table G-1), by the average July-September flow for the period of record. The estimated percent reductions July-September flow for each affected stream reach, for each MY are in Table G-2.

	Percent Reduction in Average July - September Flow									
<b>Mine</b> Year	<b>Meadow</b> <b>Creek</b>	<b>EFSFSR</b> <b>Above</b> <b>Meadow</b> <b>Creek</b>	<b>EFSFSR</b> <b>Betwee</b> $\mathbf n$ <b>Meadow</b> <b>Creek</b> and the <b>YPP</b>	<b>EFSFSR</b> <b>Betwee</b> n the <b>YPP And</b> <b>Sugar</b> <b>Creek</b>	<b>Sugar</b> <b>Creek</b>	<b>EFSFSR</b> <b>Between</b> <b>Sugar</b> <b>Creek and</b> Johnson <b>Creek</b>	<b>EFSFSR</b> <b>Betwee</b> $\mathbf n$ Johnson <b>Creek</b> and the <b>SFSR</b>	<b>SFSR</b> <b>Betwee</b> $\mathbf n$ <b>EFSFSR</b> and the <b>Salmon</b> <b>River</b>		
$-2$	$-2.2$	0.0	$-1.1$	$-1.0$	0.0	$-0.6$	$-0.1$	$-0.1$		
$-1$	2.2	0.0	1.1	3.9	0.0	2.3	0.5	0.2		
$\mathbf{1}$	4.6	0.0	2.4	7.2	2.7	5.3	1.0	0.5		
$\overline{2}$	8.4	0.0	4.1	16.3	2.7	10.6	2.1	1.0		
$\overline{3}$	9.2	0.0	4.6	11.4	2.7	7.8	1.5	0.7		
$\overline{4}$	15.4	0.0	7.3	12.4	2.7	8.4	1.7	0.8		
$\overline{5}$	$-2.4$	0.0	$-0.8$	4.6	2.7	3.8	0.8	0.4		
6	12.3	0.0	7.2	11.8	2.7	8.0	1.6	0.7		
$\overline{7}$	24.5	0.0	12.8	17.6	2.7	11.4	2.3	1.1		
8	19.2	0.0	9.1	16.9	2.7	11.0	2.2	1.0		
9	6.0	0.0	3.0	10.0	2.7	7.0	1.4	0.6		
10	6.1	0.0	2.9	10.5	2.7	7.2	1.4	0.7		
11	5.9	0.0	3.0	11.4	2.7	7.8	1.5	0.7		
12	6.8	0.0	3.4	9.4	2.7	6.6	1.3	0.6		
13	7.4	0.1	4.0	10.2	2.7	7.1	1.4	0.7		
14	5.2	0.1	3.3	6.9	2.7	5.1	1.0	0.5		
15	5.9	0.1	2.6	4.8	2.7	3.9	0.8	0.4		
16	2.1	0.1	0.7	2.2	2.7	2.4	0.5	0.2		
17	2.3	0.1	1.6	2.6	2.7	2.6	0.5	0.2		
18	2.9	0.1	1.8	2.7	2.7	2.7	0.5	0.3		
19	1.9	0.1	1.2	1.6	2.7	2.1	0.4	0.2		
20	1.3	0.1	0.8	1.6	2.7	2.0	0.4	0.2		
$20+$	0.0	0.0	0.0	0.0	1.6	0.7	0.1	0.06		

**Table G-2. Percent Reduction in Average July-September, Flow by Reach.** 

## **Effects of Flow Reduction SR spring/summer Chinook Salmon and SRB Steelhead Productivity**

We compared population productivity to population density and flow during the rearing life stage to estimate the effects of the flow related effects of proposed action on SR spring/summer Chinook salmon and SR Basin steelhead.

## **Population Trend Data**

We used redds counted in Johnson Creek from 1998 through 2022 as an index of EFSFSR Chinook salmon population size, redds counted in the SFSR from 2009 through 2022 as an index of the SFSR Chinook salmon population size (Idaho Department of Fish and Game reports), redds counted in the Secesh River system from 2009 through 2022 as in index for the Secesh River Chinook salmon population size (Kennedy et al. 2013; Stiefel et al. 2014; Stiefel et al. 2015; Stiefel et al. 2016; Belnap et al. 2017; Felts et al. 2018; Rabe et al. 2018; Felts et al. 2019; Felts et al. 2020; Nau et al. 2021; Poole et al., 2021; Ruthven et al., 2022) and estimated numbers of steelhead returning to the SFSR population as in index of the SFSR steelhead population size (Copeland et al. 2013; Copeland et al. 2014; Copeland et al. 2015; Stark et al. 2016; Stark et al. 2017; Stark et al. 2018; Stark et al. 2019a; Stark et al. 2019b; Stark et al. 2021; Steelhead Run Reconstruction Workgroup 2021; Baum et al. 2022).

# **2.2 Population Productivity**

For the three Chinook salmon populations, we calculated whole life cycle productivity (productivity) by dividing the recruit redds by the stock redds, assuming that 75% of adult returns were four years old and 25% were five years old. For the SFSR steelhead population, we calculated productivity by dividing estimated number of returns by the estimated number of returns four year previous, under the assumption of a consistent four year generation time with no repeat spawning (i.e., all spawners are four years old).

## **2.3 Flow Data used in Productivity/Flow Comparisons**

We used data from the Johnson Creek near Yellow Pine gage (USGS 13313000) (Johnson gage) for the EFSFSR Chinook salmon population and the SFSR steelhead population. Most of the EFSFSR Chinook salmon population, and a large portion of the SFSR steelhead population, spawns and rears in the Johnson Creek drainage. We used flow data from the SFSR measured at the Krassel Ranger Station gage (USGS 13310700) (SFSR gage) for the SFSR and Secesh River Chinook salmon populations. Most of the SFSR Chinook salmon population spawns and rears in the SFSR upstream form the EFSFSR and the SFSR gage is the only gage in that reach. There are no stream flow gages in the Secesh River drainage, but the SFSR gage is the closest gage outside of the Secesh River drainage.

## **2.4 Regression Equations and Reach Level Productivity Effects**

For the three SR spring/summer Chinook salmon populations, we compared the natural log of population productivity to average July-September (see section 2.3) recorded during the year following the brood year. For the one SR Basin steelhead population, we compared the natural log of population productivity to average July-September flow recorded during the brood year. For all populations, we also included the natural log of brood year population size in the regression models. The results are in Figures G-1 through G-4.

We calculated the flow-related effects of the proposed action on fish population productivity, for each stream reach, by inputting the estimated percent reduction in average July – September flow (Table G-2), and the average population size for the period of record, into the regression equations. We used the regression equation for SFSR Chinook salmon (Figure G-2) to calculate flow effects on Chinook salmon in the SFSR downstream from the EFSFSR, and the equation for EFSFSR Chinook salmon (Figure G-1) to calculate flow effects on Chinook salmon in all other

stream reaches. We used the regression equation for SFSR steelhead (Figure G-3) to calculate the flow effects on steelhead in all stream reaches. The percent reduction in Chinook salmon and steelhead productivity, by reach and MY, are in Tables G-3 and G-4.

We did not use the regression equation for Secesh River Chinook salmon (Figure G-4) in any of the effects analyses, but included the results here for informational purposes. The Secesh River Chinook salmon population is in the SFSR drainage, immediately adjacent to the EFSFSR and SFSR Chinook salmon population areas, and the population therefore experiences similar climatic and flow conditions as the SFSR and EFSFSR populations. Although not in the action area, the close proximity and similar conditions warrants inclusion of the Secesh River Chinook salmon population productivity versus flow relationship in this Appendix.

	<b>Percent Reduction in Chinook Salmon Productivity</b>									
<b>Mine</b> Year	<b>Meadow</b> <b>Creek</b>	<b>EFSFSR</b> <b>Above</b> <b>Meadow</b> <b>Creek</b>	<b>EFSFSR</b> <b>Betwee</b> $\mathbf n$ <b>Meadow</b> <b>Creek</b> and the <b>YPP</b>	<b>EFSFSR</b> <b>Betwee</b> n the <b>YPP</b> and <b>Sugar</b> <b>Creek</b>	<b>Sugar</b> <b>Creek</b>	<b>EFSFSR</b> <b>Between</b> <b>Sugar</b> <b>Creek and</b> Johnson <b>Creek</b>	<b>EFSFSR</b> <b>Betwee</b> $\mathbf n$ <b>Johnson</b> <b>Creek</b> and the <b>SFSR</b>	<b>SFSR</b> <b>Betwee</b> $\mathbf n$ <b>EFSFSR</b> and the <b>Salmon</b> <b>River</b>		
$-2$	$-2.3$	0.0	$-1.1$	$-1.0$	0.0	$-0.6$	$-0.1$	0.0		
$-1$	2.3	0.0	1.1	4.2	0.0	2.4	0.5	0.1		
$\mathbf{1}$	4.9	0.0	2.5	7.7	2.8	5.7	1.1	0.3		
$\overline{2}$	9.2	0.0	4.3	18.5	2.8	11.7	2.2	0.7		
$\overline{3}$	10.0	0.0	4.9	12.6	2.8	8.5	1.6	0.5		
$\overline{4}$	17.4	0.0	7.8	13.8	2.8	9.1	1.7	0.5		
$\overline{5}$	$-2.4$	0.1	$-0.9$	4.9	2.8	4.1	0.8	0.2		
6	13.6	0.3	7.8	13.1	2.8	8.7	1.7	0.5		
$\overline{7}$	29.0	0.3	14.3	20.1	2.8	12.6	2.4	0.7		
8	22.1	0.1	10.0	19.2	2.8	12.1	2.3	0.7		
9	6.5	0.0	3.2	11.0	2.8	7.5	1.4	0.4		
10	6.5	0.0	3.1	11.5	2.8	7.8	1.5	0.5		
11	6.4	0.0	3.2	12.6	2.8	8.5	1.6	0.5		
12	7.3	0.0	3.6	10.2	2.8	7.1	1.4	0.4		
13	8.0	1.0	4.3	11.2	2.8	7.6	1.5	0.4		
14	5.6	1.7	3.5	7.4	2.8	5.5	1.1	0.3		
15	6.3	1.3	2.7	5.2	2.8	4.2	0.8	0.2		
16	2.3	0.7	0.7	2.3	2.8	2.5	0.5	0.1		
17	2.4	1.8	1.7	2.7	2.8	2.8	0.5	0.2		
18	3.1	1.4	1.9	2.9	2.8	2.8	0.6	0.2		
19	2.0	1.0	1.3	1.7	2.8	2.2	0.4	0.1		
20	1.3	0.8	0.8	1.6	2.8	2.1	0.4	0.1		
$20+$	0.0	0.0	0.0	0.0	1.7	0.7	0.1	0.04		

**Table G-3. Percent Reduction in Chinook Salmon Productivity by Reach.**

	Percent Reduction in Average July - September Flow									
<b>Mine</b> Year	<b>Meadow</b> <b>Creek</b>	<b>EFSFSR</b> <b>Above</b> <b>Meadow</b> <b>Creek</b>	<b>EFSFSR</b> <b>Betwee</b> $\mathbf n$ <b>Meadow</b> <b>Creek</b> and the <b>YPP</b>	<b>EFSFSR</b> <b>Betwee</b> n the <b>YPP</b> and <b>Sugar</b> <b>Creek</b>	<b>Sugar</b> <b>Creek</b>	<b>EFSFSR</b> <b>Between</b> <b>Sugar</b> <b>Creek and</b> Johnson <b>Creek</b>	<b>EFSFSR</b> <b>Betwee</b> $\mathbf n$ Johnson <b>Creek</b> and the <b>SFSR</b>	<b>SFSR</b> <b>Betwee</b> $\mathbf n$ <b>EFSFSR</b> and the <b>Salmon</b> <b>River</b>		
$-2$	$-4.3$	0.0	$-2.1$	$-2.0$	0.0	$-1.2$	$-0.2$	$-0.1$		
$-1$	4.4	0.0	2.2	8.1	0.0	4.6	0.9	0.4		
$\overline{1}$	9.6	0.0	4.8	15.2	5.4	11.0	2.1	1.0		
$\overline{2}$	18.2	0.0	8.4	38.0	5.4	23.4	4.3	2.0		
3	19.9	0.0	9.5	25.4	5.4	16.7	3.1	1.4		
$\overline{4}$	35.5	0.0	15.4	27.9	5.4	18.0	3.3	1.5		
$\overline{5}$	$-4.5$	0.2	$-1.6$	9.6	5.4	7.8	1.5	0.7		
6	27.5	0.5	15.3	26.4	5.4	17.2	3.2	1.5		
$\overline{7}$	62.3	0.5	28.8	41.6	5.4	25.3	4.6	2.1		
8	46.2	0.2	19.8	39.6	5.4	24.3	4.4	2.0		
9	12.7	0.0	6.1	21.9	5.4	14.8	2.8	1.3		
10	12.8	0.0	5.9	23.0	5.4	15.4	2.9	1.3		
11	12.5	0.0	6.1	25.4	5.4	16.7	3.1	1.4		
12	14.3	0.0	6.9	20.4	5.4	13.9	2.6	1.2		
13	15.7	1.7	8.3	22.3	5.4	15.0	2.8	1.3		
14	10.8	3.0	6.7	14.6	5.4	10.7	2.0	0.9		
15	12.3	2.2	5.3	10.0	5.4	8.1	1.6	0.7		
16	4.3	1.2	1.4	4.3	5.4	4.8	0.9	0.4		
17	4.6	3.1	3.2	5.2	5.4	5.3	$1.0\,$	0.5		
18	5.9	2.4	3.6	5.5	5.4	5.5	1.1	0.5		
19	3.9	1.8	2.4	3.2	5.4	4.2	0.8	0.4		
20	2.5	1.4	1.6	3.1	5.4	4.1	0.8	0.4		
$20+$	0.0	0.0	0.0	0.0	3.2	1.3	0.3	0.1		

**Table G-4. Percent Reduction in Steelhead Productivity by Reach.**

## **2.5 Population Level Productivity Effects**

We used the amount of habitat in each affected stream reach, expressed as weighted intrinsic potential (Cooney and Holzer 2006), to scale the stream reach specific reductions in productivity to the population level. The amount of Chinook salmon and steelhead weighted intrinsic potential, in each of the analyzed stream reaches, is in Table G-5. We calculated the population level effects by multiplying the stream reach specific reductions in productivity (Tables G-3 and G-4) by the percent of habitat in the affected stream reach (Table G-5). The results for Chinook salmon and steelhead are in Tables G-6 and G-7, respectively.

**Table G-5. The amount of Chinook salmon and steelhead habitat, measured as weighted intrinsic potential, and expressed as m<sup>2</sup> and as a percentage of the population, for each analyzed stream reach.**



**Note:** The percentages are for the SFSR steelhead population (all stream reaches), the EFSFSR Chinook salmon population (all stream reaches except the SFSR), and the SFSR Chinook salmon population (SFSR between the EFSFSR and the Salmon River).

	<b>Percent Reduction in Productivity (Percent)</b>								
<b>Mine</b> Year	<b>Meadow</b> <b>Creek</b>	<b>EFSFSR</b> <b>Above</b> <b>Meadow</b> <b>Creek</b>	<b>EFSFSR</b> <b>Betwee</b> $\mathbf n$ <b>Meadow</b> <b>Creek</b> and the <b>YPP</b>	<b>EFSFSR</b> <b>Between</b> the YPP and <b>Sugar</b> <b>Creek</b>	<b>Sugar</b> <b>Creek</b>	<b>EFSFSR</b> <b>Between</b> <b>Sugar Creek</b> and Johnson <b>Creek</b>	<b>EFSFSR</b> <b>Between</b> <b>Johnson Creek</b> and the SFSR	<b>Total for the</b> <b>EFSFSR</b> <b>Chinook</b> <b>Salmon</b> Population	<b>SFSR Between</b> <b>EFSFSR</b> and the <b>Salmon River,</b> <b>Total for the SFSR</b> <b>Chinook Salmon</b> Population
$-2$	$\overline{0}$	$\mathbf{0}$	$\Omega$	$-0.001$	0.000	$-0.090$	$-0.016$	$-0.112$	$-0.001$
$-1$	$\overline{0}$	$\mathbf{0}$	$\theta$	0.003	0.000	0.355	0.061	0.439	0.003
$\mathbf{1}$	$\overline{0}$	$\mathbf{0}$	$\boldsymbol{0}$	0.005	0.007	0.831	0.140	1.021	0.007
$\overline{2}$	$\overline{0}$	$\theta$	$\overline{0}$	0.011	0.007	1.717	0.283	2.109	0.014
3	$\boldsymbol{0}$	$\boldsymbol{0}$	$\mathbf{0}$	0.008	0.007	1.240	0.207	1.524	0.010
$\overline{4}$	$\overline{0}$	$\boldsymbol{0}$	$\boldsymbol{0}$	0.008	0.007	1.337	0.222	1.643	0.011
5	$\boldsymbol{0}$	$\mathbf{0}$	$\mathbf{0}$	0.003	0.007	0.595	0.101	0.730	0.005
6	$\boldsymbol{0}$	$\boldsymbol{0}$	$\boldsymbol{0}$	0.008	0.007	1.279	0.213	1.572	0.010
$\overline{7}$	$\boldsymbol{0}$	$\mathbf{0}$	$\mathbf{0}$	0.012	0.007	1.849	0.303	2.271	0.015
8	$\overline{0}$	$\mathbf{0}$	$\overline{0}$	0.012	0.007	1.776	0.292	2.182	0.014
9	$\overline{0}$	$\mathbf{0}$	$\overline{0}$	0.007	0.007	1.103	0.185	1.356	0.009
10	$\overline{0}$	$\mathbf{0}$	$\overline{0}$	0.007	0.007	1.147	0.192	1.410	0.009
11	$\overline{0}$	$\mathbf{0}$	$\mathbf{0}$	0.008	0.007	1.240	0.207	1.524	0.010
12	$\mathbf{0}$	$\boldsymbol{0}$	$\mathbf{0}$	0.006	0.007	1.042	0.174	1.280	0.008
13	$\boldsymbol{0}$	$\boldsymbol{0}$	$\boldsymbol{0}$	0.007	0.007	1.120	0.187	1.377	0.009
14	$\overline{0}$	$\mathbf{0}$	$\mathbf{0}$	0.004	0.007	0.807	0.136	0.991	0.007
15	$\boldsymbol{0}$	$\boldsymbol{0}$	$\mathbf{0}$	0.003	0.007	0.615	0.104	0.755	0.005
16	$\mathbf{0}$	$\boldsymbol{0}$	$\boldsymbol{0}$	0.001	0.007	0.367	0.063	0.449	0.003
17	$\boldsymbol{0}$	$\boldsymbol{0}$	$\mathbf{0}$	0.002	0.007	0.405	0.069	0.496	0.003
18	$\overline{0}$	$\mathbf{0}$	$\overline{0}$	0.002	0.007	0.418	0.071	0.512	0.003
19	$\overline{0}$	$\mathbf{0}$	$\mathbf{0}$	0.001	0.007	0.318	0.054	0.389	0.003
20	$\overline{0}$	$\mathbf{0}$	$\Omega$	0.001	0.007	0.312	0.053	0.382	0.003
$20+$	0.0	0.0	0.0	0.0	0.004	0.103	0.018	0.126	0.0009

**Table G-6. Chinook Salmon Population Level Reduction in Productivity, By Reach, And Total for the EFSFSR And SFSR Chinook Salmon Populations.** 

	<b>Percent Reduction in Productivity (Percent)</b>								
<b>Mine</b> Year	<b>Meadow</b> <b>Creek</b>	<b>EFSFSR</b> Above <b>Meadow</b> <b>Creek</b>	<b>EFSFSR</b> <b>Betwee</b> $\mathbf n$ <b>Meadow</b> <b>Creek</b> and the <b>YPP</b>	<b>EFSFSR</b> <b>Between</b> the YPP and <b>Sugar</b> <b>Creek</b>	<b>Sugar</b> <b>Creek</b>	<b>EFSFSR</b> <b>Between</b> <b>Sugar Creek</b> and Johnson <b>Creek</b>	<b>EFSFSR</b> <b>Between</b> <b>Johnson Creek</b> and the SFSR	<b>SFSR</b> <b>Between</b> <b>EFSFSR And</b> the Salmon <b>River</b>	<b>Total for the SFSR</b> <b>Steelhead</b> <b>Population</b>
$-2$	$-0.005$	0.000	$-0.008$	$-0.002$	0.000	0.000	$-0.012$	$-0.017$	$-0.045$
$-1$	0.005	0.000	0.008	0.008	0.000	0.000	0.047	0.066	0.135
$\mathbf{1}$	0.012	0.000	0.019	0.015	0.010	0.000	0.109	0.153	0.319
2	0.022	0.000	0.033	0.038	0.010	0.000	0.222	0.310	0.635
3	0.024	0.000	0.037	0.025	0.010	0.000	0.162	0.227	0.485
$\overline{4}$	0.044	0.000	0.060	0.028	0.010	0.000	0.174	0.244	0.559
5	$-0.006$	0.001	$-0.006$	0.010	0.010	0.000	0.079	0.111	0.198
6	0.034	0.002	0.060	0.026	0.010	0.000	0.167	0.233	0.532
$\overline{7}$	0.076	0.002	0.112	0.042	0.010	0.000	0.238	0.332	0.813
8	0.057	0.001	0.077	0.040	0.010	0.000	0.229	0.320	0.734
9	0.016	0.000	0.024	0.022	0.010	0.000	0.144	0.202	0.418
10	0.016	0.000	0.023	0.023	0.010	0.000	0.150	0.210	0.432
11	0.015	0.000	0.024	0.025	0.010	0.000	0.162	0.227	0.463
12	0.018	0.000	0.027	0.020	0.010	0.000	0.136	0.191	0.403
13	0.019	0.007	0.032	0.022	0.010	0.000	0.147	0.205	0.443
14	0.013	0.012	0.026	0.015	0.010	0.000	0.106	0.149	0.332
15	0.015	0.009	0.020	0.010	0.010	0.000	0.081	0.114	0.260
16	0.005	0.005	0.005	0.004	0.010	0.000	0.049	0.069	0.147
17	0.006	0.013	0.012	0.005	0.010	0.000	0.054	0.076	0.176
18	0.007	0.010	0.014	0.006	0.010	0.000	0.055	0.078	0.180
19	0.005	0.007	0.009	0.003	0.010	0.000	0.042	0.060	0.137
20	0.003	0.006	0.006	0.003	0.010	0.000	0.042	0.059	0.128
$20+$	0.0	0.0	0.0	0.0	0.006	0.043	0.014	0.019	0.082

**Table G-7. SFSR Steelhead Population Level Reduction in Productivity, By Reach and Total.** 

# **2.6 Population Level effects on Fish**

We converted the effect on population productivity into the effect, expressed as the number of returning adults, but multiplying the percent reduction in population productivity by the average population size. The before action population productivity was assumed to be the productivity at average population density and average flow for the period of record (Table G-8).

<b>Mine Year</b>	<b>EFSFSR Chinook</b> <b>Salmon</b>	<b>SFSR Chinook</b> <b>Salmon</b>	<b>SFSR Steelhead</b>
$-2$	$-0.4$	0.00	$-0.4$
$-1$	$1.5\,$	0.02	1.3
$\mathbf{1}$	3.5	0.04	3.0
$\overline{2}$	7.2	0.09	6.0
3	5.2	0.07	4.6
$\overline{4}$	5.6	0.07	5.3
5	2.5	0.03	1.9
6	5.4	0.07	5.0
$\overline{7}$	7.8	0.10	7.7
8	7.4	0.09	6.9
9	4.6	0.06	3.9
10	4.8	0.06	4.1
11	5.2	0.07	4.4
12	4.4	0.05	3.8
13	4.7	0.06	4.2
14	3.4	0.04	3.1
15	2.6	0.03	2.5
16	1.5	0.02	1.4
17	1.7	0.02	1.7
18	1.7	0.02	1.7
19	1.3	0.02	1.3
20	1.3	0.02	1.2
$20+$	$0.4\,$	0.01	$0.8\,$

**Table G-8. Estimated effects expressed as number of returning adults.** 

**Note:** Negative numbers are positive effects.



**Figure G-1. Results of Multivariate Regression of EFSFSR Chinook Salmon Population Productivity Against Flow and Population Density.** 

**Note:** The leverage plots for flow and density are in panels A and B, respectively, and the whole model plot is in panel C.



**Figure G-2. Results of Multivariate Regression of SFSR Chinook Salmon Population Productivity Against Flow and Population Density.**

**Note:** The leverage plots for flow and density are in panels A and B, respectively, and the whole model plot is in panel C.



**Figure G-3. Results of Multivariate Regression of SFSR Steelhead Population Productivity Against Flow and Population Density.**

**Note:** The leverage plots for flow and density are in panels A and B, respectively, and the whole model plot is in panel C.



**Figure G-4. Secesh River Chinook Salmon.**

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